



State of Oregon  
Department of  
Environmental  
Quality

# Columbia Slough Sediment Study

Lower Slough between River Mile 5.9 and 8.7.

January 2012 update



**January 31, 2012 Lower slough update**

As a result of the evaluations conducted for the Whitaker Slough, conclusions associated with toxicity and bioaccumulation screening levels made as part of the Lower Slough evaluation were reconsidered. Because of the increased sensitivity observed in the chronic bioassay tests conducted on sediment in the Whitaker Slough and the lack of chronic testing results for sediment in the Lower Slough, we are no longer recommending that the toxicity screening levels be modified from default values. In addition, further scrutiny of the methods used to derive Lower Slough specific bioaccumulation values revealed that screening levels are approximately equal to the default values established in DEQ's bioaccumulation guidance. Consequently, pending further evaluation of Columbia Slough data, segment-specific risk-based levels are no longer applicable.

The table in Appendix H has been revised to reflect this change. Sections 7.2 and 7.3 of the text have been replaced with a note that data is being re-evaluated as part of the broader Columbia Slough study. A report on that work will be available in 2013.

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## **Executive Summary**

The objective of this study was to evaluate previous estimates of sediment ambient baseline concentrations and identify where hazardous substances are present at concentrations significantly exceeding baseline concentrations in a 4.3-km reach of Portland, Oregon's Columbia Slough. Two sampling methods were used to characterize shallow sediments: incremental and targeted composite. The incremental sampling (IS) method was used to determine representative baseline concentrations for the reach. For IS sampling, 50 sample increments were collected using a grid overlain on the entire study reach. Material from the 50 locations was randomly combined into three replicates such that each of three samples contained sediment from 30 sample locations. Targeted, composite samples were collected in the vicinity of public and private outfalls and areas of known bank contamination where elevated concentrations of hazardous substances were more likely to be present above ambient baseline levels. Target composite samples included up to nine sample locations within the outfall/bank area. Analytes included metals, pesticides, polychlorinated biphenyls (PCBs) (Aroclors and congeners), semi-volatile organic compounds, tributyltin, and polybrominated diphenyl ethers. Results indicate that the highest contaminant levels are generally associated with outfalls and that the IS samples provide a statistically robust indication of reach-wide sediment concentrations that can be used for establishing baseline levels.

IS samples were also used in conjunction with fish tissue data already collected, to determine Biota Sediment Accumulation Factors (BSAF) for bioaccumulative contaminants. BSAFs are used to determine site-specific risk-based clean-up levels for bioaccumulative contaminants in sediment.

In addition to sediment chemistry, bioassays were conducted on sediment collected from 10 locations in the study area. The results of these tests were used to evaluate the potential for benthic toxicity associated with sediment contamination.

Consistent with previous studies, PCBs, were found to be the primary contaminant of concern in this reach of the Columbia Slough due to the potential risk to humans and wildlife that may consume impacted fish.

Several target areas were identified as having elevated concentrations of PCBs or other contaminants in sediment as compared to ambient baseline levels. Those areas in which at least one contaminant was detected at a concentration that was an order of magnitude higher than the risk-based concentration or exceeded the newly established baseline level (whichever was higher) were highlighted. An area-weighted averaging approach was used to identify the five areas where cleanup could lower average PCB concentrations in the segment.

# Comprehensive Columbia Slough: Lower Slough Report

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## APPENDICES

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E - 2009 Columbia Slough Sediment Triplicate/Duplicate QC Review

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## **1.0 INTRODUCTION**

In 2008/2009 DEQ initiated a process through which parties determined to have likely contributed to contamination in Columbia Slough sediments could settle their liability for this contamination by paying into an account set up by DEQ. DEQ would use the money collected in this account to complete environmental investigation and cleanup in the portion of the Slough impacted by the contributing parties. The first parties to take advantage of this process are located in the Lower Columbia Slough. Six parties discharging to City and private outfalls in the section of the Slough between City Outfall 59 (near River Mile 5.9) and the upstream limit of the Lower Slough at Pen 2 Levee (River Mile 8.7) (see Figure 1) settled with DEQ. DEQ used settlement money to design and implement a sediment investigation in this area. This report describes the study, documents results, provides analysis and conclusions, and identifies next steps.

### **1.1 Report Organization**

This report is organized as follows:

- Section 1 provides a brief overview of the project and objectives.
- Section 2 describes the Columbia Slough Watershed with more detailed information on the focus area of this study.
- Section 3 summarizes sample collection methods, analytical procedures, and the data quality review.
- Section 4 presents the analytical results for samples and provides some comparisons to ambient and risk-based levels.
- Section 5 lays the foundation for developing a relationship between sediment concentrations and fish tissue levels.
- Section 6 summarizes data from other investigations in the Lower Slough to provide a comprehensive picture of the available data and where additional data may be needed.
- Section 7 presents the study conclusions.
- Section 8 outlines the additional work that may be needed to identify source control needs and develop a sediment remedy for the Lower Slough.

### **1.2 Study Objectives**

The Record of Decision (ROD) for the Columbia Slough specified three primary tasks: 1) source control, 2) sediment cleanup, and 3) long-term monitoring to evaluate the effectiveness of natural recovery at reducing sediment contamination to risk-based levels. This study will

characterize the nature and extent of hazardous substances in sediments in this reach of the Columbia Slough to support achieving each of these objectives.

### 1.2.1 Complete Source Control Evaluation

DEQ and the City of Portland developed the Columbia Slough Sediment Watershed Action Plan in 2005, which identified priority areas within the slough. The lower slough from Interstate 5 to Martin Luther King Jr. Blvd. was identified as one of five areas with elevated sediment concentrations and high total suspended solids (TSS) in stormwater runoff. The DEQ sampling efforts intend to support the Columbia Slough Watershed Action Plan by pinpointing higher concentrations of contaminants of concern at outfall and bank source areas within the study area and helping to identify where additional source control evaluation may be warranted.

### 1.2.2 Define Sediment Cleanup Areas

The cleanup goal for sediment in the Columbia Slough is the higher of established baseline concentrations (concentrations reflecting area-wide levels below which active cleanup may not be feasible due to potential recontamination) or risk-based levels. The incremental sample (IS) data collected in this study will provide a more defensible, site-specific, and up-to-date representation of baseline concentrations in the lower slough (see Section 3 below).

The study will identify “hot spot” zones, defined as those areas where sediment concentrations are more than an order of magnitude above the corresponding risk-based screening level or where they exceed baseline concentration for the target reach whichever is higher. These are areas where additional point and non-point source control measures for storm water utilities or contaminated bank soils may be needed. If hot spot zones are found, it is likely that further investigation of the depth and full extent of contamination will be required.

Risk based cleanup levels will also be re-evaluated based on toxicity test results and a comparison of fish tissue and sediment data. The toxicity tests will help to define what contaminant concentrations are likely to present a risk to benthic organisms. The fish tissue comparison will be used to estimate a more site-specific bioaccumulation-based sediment cleanup goal. It is recognized that both toxicity and bioaccumulation evaluations will be based on a somewhat limited data set and re-evaluation will be warranted as more data is obtained. This will be further discussed in Section 8.

### 1.2.3 Evaluate Natural Recovery

Once effective source control measures are implemented, new sediment entering the slough as a result of stormwater runoff should be cleaner and will eventually cover sediment containing residual contamination. This is one natural recovery mechanism we expect to occur in the Columbia Slough and, in fact, is likely already occurring as a result of source control measures implemented to date. Another aspect of this study was collection of samples designed to assess this mechanism by looking at contaminant concentration variation with depth at one location.

## **2.0 AREA BACKGROUND**

The Columbia Slough Watershed drains approximately 32,700 acres of land (see Figure 1). Portland's city limits end at approximately NE 185th Avenue on the east, but the watershed includes Fairview Lake and Fairview Creek, and portions of Troutdale, Fairview, Gresham, Maywood Park, Wood Village, and unincorporated Multnomah County. The Watershed historically contained a vast system of side channels, streams, ponds, lakes, and wetlands that covered the floodplain of the Columbia River between the mouths of the Willamette and Sandy Rivers. High water seasonally inundated the floodplain, cutting new channels and depositing sediment. Native Americans used these waterways and the uplands for fishing, hunting, and gathering food. Fishing continues in this area.

Over the years, the watershed and waterway have been drastically altered. Beginning in 1918, levees were built and wetlands were drained and filled to provide flood protection and allow for development. The waterway was channelized, and dozens of streams were diverted from natural channels to underground pipes. Today, the Columbia Slough comprises a 19-mile main channel that parallels the Columbia River, as well as approximately 12 additional miles of secondary waterways. Other remaining major surface water features include Fairview Creek, Fairview Lake, and Smith and Bybee Lakes. Floodplain development has resulted in an extensively managed surface water system that includes levees, pumps, and other water control structures. The levee system has greatly changed the historic floodplain and reduced the area available to floodwaters.



**Figure 1 Columbia Slough Watershed**

The Columbia Slough Watershed includes virtually every type of land use: residential neighborhoods, commercial and industrial development, agriculture, Portland International Airport (PDX), interstate highways, railroad corridors, 54 schools, and large open spaces. Much of Portland’s industrial and commercial land is located within the Watershed. In addition to industrial development in the area north of Columbia Boulevard and the Rivergate area, land is preserved for industrial uses in the Columbia South Shore area between NE 82nd and NE 185th Avenues north of Sandy Boulevard.

The Slough is divided into three sections, based on hydraulic characteristics:

- The **Upper Slough** starts at the mouth of Fairview Lake on the east and flows west to the mid-dike levee at NE 142nd Avenue. It receives water from Fairview Lake, Fairview Creek, Wilkes Creek, stormwater outfalls, natural springs, groundwater, and overland flow.

- The **Middle Slough** extends from the mid-dike levee near NE 142nd Avenue to the Pen 2 levee near NE 18th Avenue. It includes a substantial southern arm complex of sloughs and lakes, including Prison Pond, Mays Lake, Johnson Lake, Whitaker Slough, Whitaker Ponds, and Buffalo Slough. The Middle Slough receives water from the Upper Slough, stormwater outfalls, natural springs, overland flow, and groundwater. Pumps are used to move water from the Upper and Middle Slough to the Lower Slough.
- The **Lower Slough** starts at the Pen 2 levee, near NE 18th Avenue, and extends approximately 8.5 miles to the Willamette River. The lowlands of the Lower Slough Watershed are subject to flooding because they are not protected by levees. Water flow and levels in the Lower Slough are affected primarily by the Columbia River and Willamette River stage and the ocean tides, as well as by pumping. During high tide, the Columbia and Willamette Rivers create a backwater effect that complicates flow patterns.

During the Columbia Slough Screening Level Risk Assessment performed in 1994, virtually every sediment sample analyzed contained one or more contaminants at concentrations exceeding conservative screening levels based on impacts to aquatic life and fish consumers. Sediment contaminants of concern include metals, pesticides, polyaromatic hydrocarbons (PAHs), phthalates, and polychlorinated biphenyls (PCBs).

Other concerns in the Columbia Slough include an approved Total Maximum Daily Load (TMDL) for the water body and a fish consumption health advisory. The TMDL, part of the Clean Water Act, includes DDT/DDE, dioxin, lead, and PCBs reduction strategies for the water column. Oregon Department of Human Services has issued a health advisory discouraging the consumption of fish in the Slough because of high PCB levels.

## **2.1 Project Area**

The project area is located within the upper portion of the Lower Slough segment. The sediments in this area, between City outfall 59 and the Pen 2 levee, have consistently contained high levels of contamination, particularly PCBs and metals. This segment also represents one of the oldest industrial areas along the Slough. Unlike the Middle and Upper Slough, the Lower Slough is tidally influenced as it is directly connected to the Willamette River. Salmon have been observed throughout the Lower Slough.

## **2.2 Contaminant Sources of Concern**

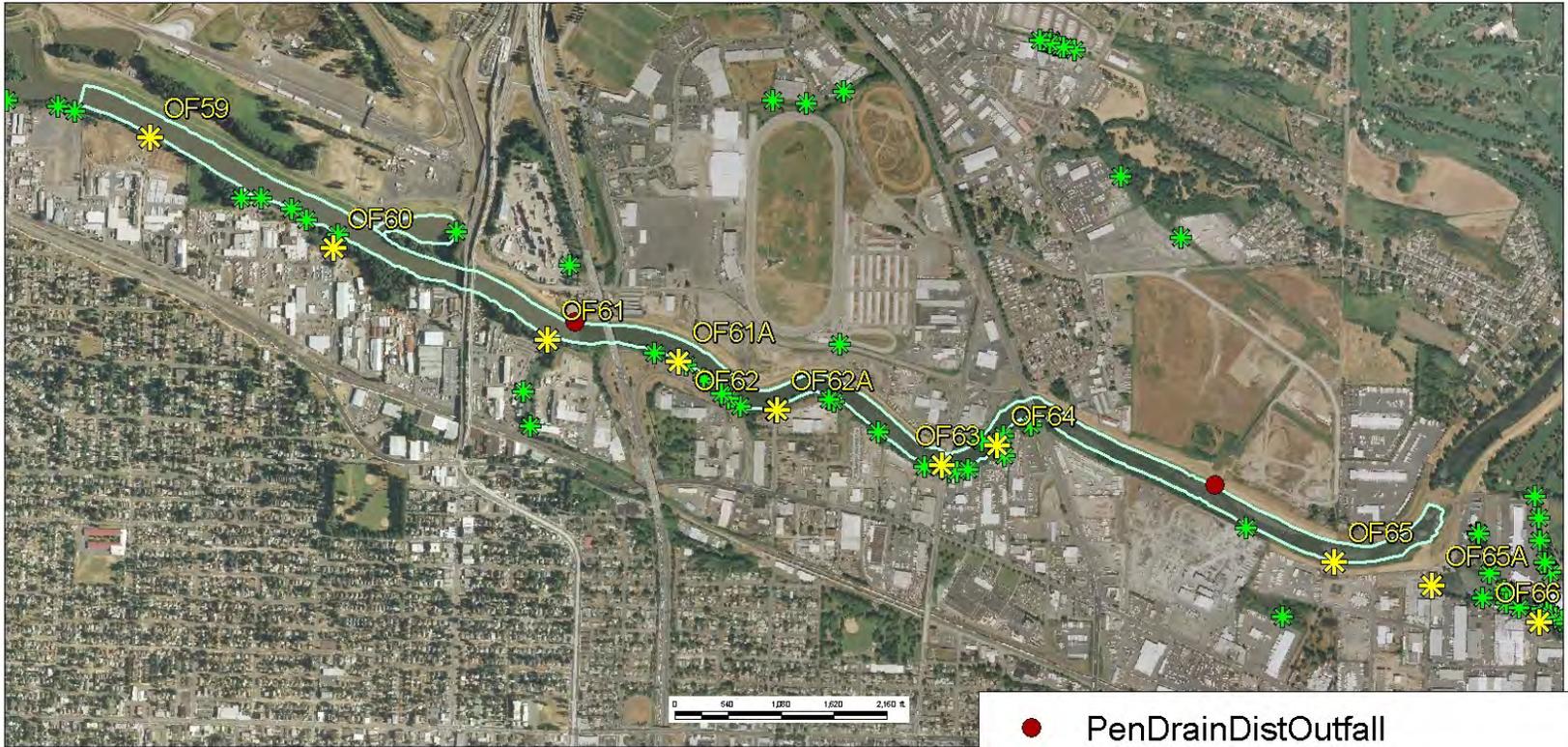
Contaminated bank areas and discharges from outfalls, both public and private are thought to be the major contributors to sediment contamination in the Slough.

### **2.2.1 Outfalls**

City of Portland (COP) and Oregon Department of Transportation (ODOT) outfalls tend to drain larger areas than private outfalls and tracking down particular contaminant sources can be more

complicated. There are seven COP outfalls and two ODOT outfall along the south shore of the study area. Four of these outfalls were once combined sewer overflow (CSO) outfalls but sanitary wastes throughout the Slough watershed is now directed to the Columbia Blvd treatment plant. Legacy pollutants associated with CSO discharge may however be present in Slough sediments.

There are approximately 20 private outfalls (many of which are associated with DEQ cleanup sites discussed below) along the south bank of the Slough and three on the north bank. The reason for few outfalls on the north bank is that the Peninsula Drainage District maintains a system of waterways and pipes on the north side of the Slough that collects most storm water internally and drains it through two large pump stations that discharge on the north shore. See Figure 2 for project area and outfall locations.



**OUTFALL 59-65  
DEQ SEDIMENT STUDY**



- PenDrainDistOutfall
- ★ City of Portland outfalls
- ✱ Non-City Outfalls (BES, Dec08)
- OF59-65channel

Base imagery ca. 2005 from Oregon Imagery Explorer web mapping service.  
Outfalls from City of Portland Bureau of Environmental Services.

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**Figure 2 Map of Project Area and Outfalls**

### 2.2.2 DEQ Cleanup Sites

Over thirty cleanup sites listed in DEQ's data base are located in the Lower Columbia Slough. Contaminants of concern at these sites include PCBs, PAHs, metals, and petroleum hydrocarbons. Some of these sites have entered into settlements with DEQ and have implemented source control measures. Some have conducted sediment investigations and others are still in the screening stage. Figures 3 and 4 shows cleanup site locations shaded in red.

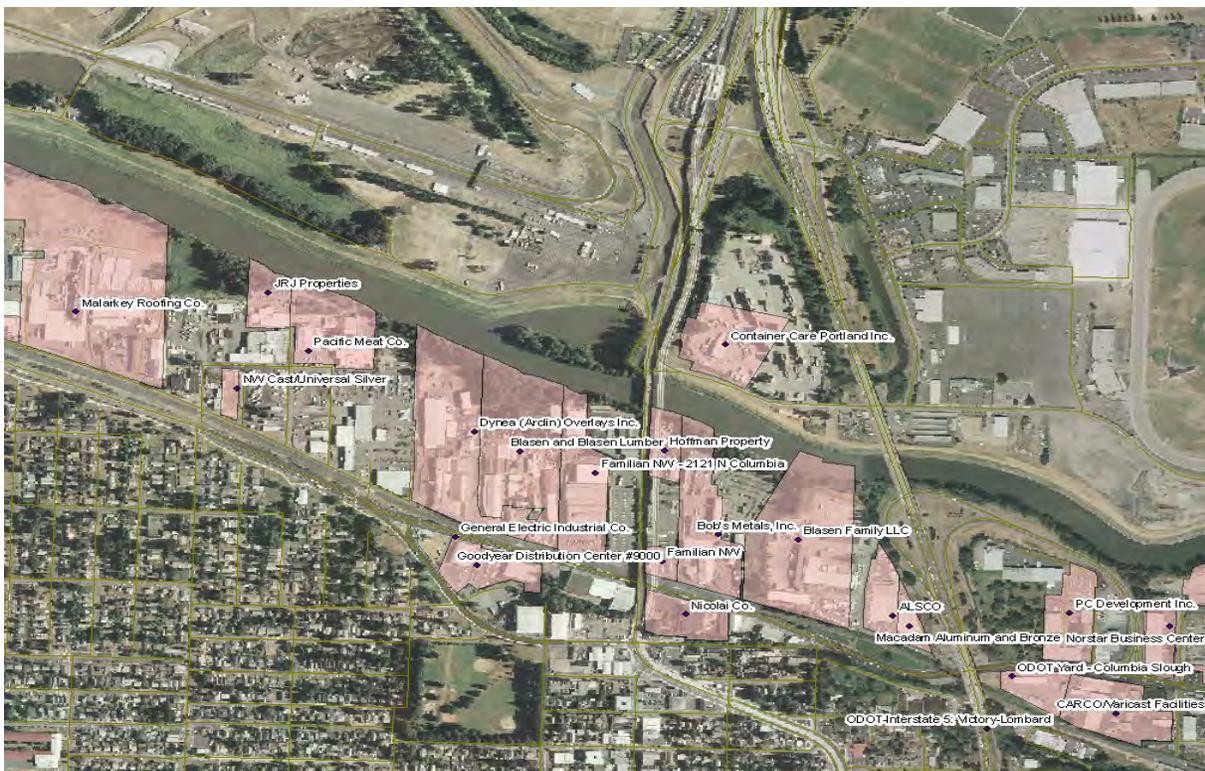


Figure 3 Cleanup sites in West Half of Study Area



**Figure 4 Cleanup Sites in East Half of Study Area**

DEQ is actively working to ensure that source control measures are implemented where needed at these sites and anticipates that major sources will be managed by 2013.

### **2.3 Previous Sampling Investigations**

Several other sampling efforts have been completed in the Lower Slough as part of specific site investigations, as part of the City's long-term monitoring, and to support dredging evaluations for bridge construction or channel maintenance. This data is summarized in Section 6 and has been incorporated into the Slough-wide data base to provide a larger dataset for evaluating data gaps and identifying "hot-spot" areas.



**Figure 5 Columbia Slough, near Pacific Meats Cleanup Site**

### **3.0 SAMPLE DESIGN AND STRATEGY**

The general sampling strategy included two sample designs: an incremental sampling (IS) strategy to determine average sediment concentrations throughout the outfall 59-Pen 2 levee study reach and targeted samples in locations of suspected contamination within the study area. The IS provides a reliable, defensible, and cost-effective method to determine average concentrations within the reach. The size of a decision unit for an IS sampling grid depends on the purpose of the data and how it will be used. The purpose of the Lower Columbia Slough decision unit is to generate a baseline sediment concentration for the typical three mile home range of carp. The baseline sediment concentrations can then be used to help determine bioaccumulation rates in fish. The baseline sediment concentrations will help to assess the feasibility of active cleanup measures within the reach. The sample design was chosen by DEQ staff after researching available methods and discussing with DEQ chemists, toxicologists, Columbia Slough project managers, and program managers.

#### **3.1 Collection and Field Activities**

Field activities were carried out in accordance with the Quality Assurance Project Plan, *Columbia Slough Outfall 59-65 Sediment Study* (DEQ, 2009); document number DEQ09-LQ-

0046-QAPP. Deviations from the QAPP are described in the report where appropriate. The QAPP was developed using EPA document # *EPA-823-B-01-002* and *Recommended Guidelines for Sampling Marine Sediment, Water Column and Tissue in Puget Sound, Puget Sound Water Quality Action Team 1997* as guidance. Sediment was analyzed for a variety of chemicals at all locations and bioassays were performed at selected locations, typically major outfalls. Additionally, sample collection procedures and analyses were matched to previous City of Portland studies when applicable. Preservation and labeling followed laboratory guidelines.

A Nationwide Permit for Survey Activities (NWP-2009-439) from the Army Corps of Engineers was obtained to collect sediment within the waters of the United States. An Oregon Department of State Lands dredge/fill permit was not required for the project. Multnomah County Drainage District provided logistical support including Slough access and boat storage. DEQ Laboratory and Environmental Assessment Division provided sampling equipment for field collection.

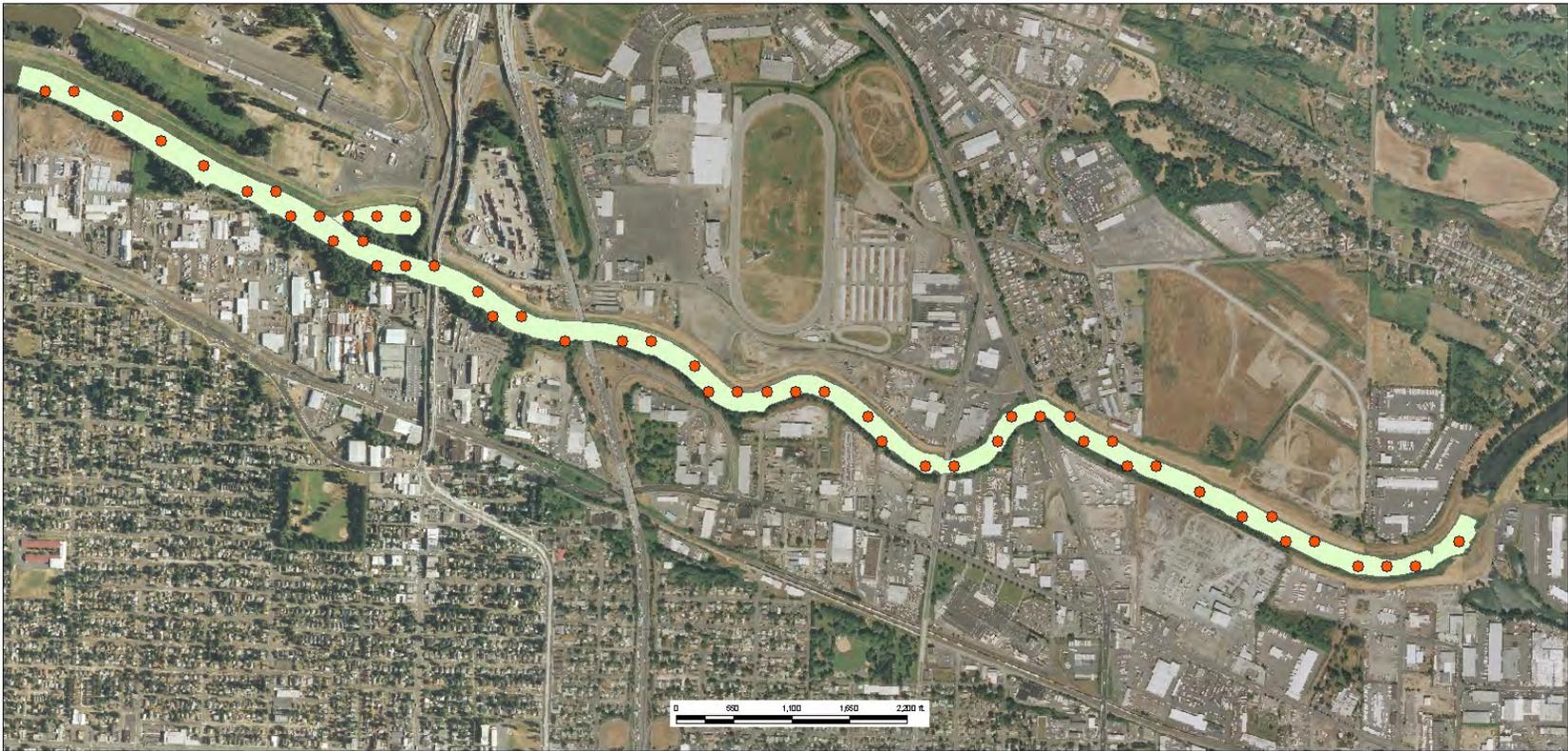
### **3.2 Sediment Sample Location Selection**

Sample locations were selected based on the two sampling strategies: IS and targeted composites at suspected source locations. See Appendix A for detailed sample location maps. Sites were located with a Garmin handheld GPS, with an average accuracy of +/- 15 feet. While all efforts were made to collect the sediment at mapped locations, some site locations deviated due to safety or accessibility issues. The majority of the sites were within 15 feet or less of the intended sample point. The two exceptions were: Simpson Cove locations for targeted and IS samples were moved closer to the bank to allow for access from the shore and Pacific Meats Outfall #3 locations were moved towards the Slough channel, to allow for access from a boat. Appendix B describes actual locations of all sample increments.

Incremental Sampling Locations: The IS strategy was designed to integrate sediment information throughout the 2.7 mile length of this Slough reach. A total of 50 sample increments of the upper 10 cm of sediment were collected from a random-start grid overlain on the entire reach. Sediment from the 50 locations was randomly combined such that each of three samples (designated A, B, and C), contained sediment from 30 of the 50 different locations shown in Figure 6.

Targeted Samples: The targeted sampling strategy involved a small array of composite sampling points immediately upstream and downstream of 17 public and private outfalls and 4 known bank contamination areas (Figure 7). The number and location of individual sample points in each composite were based on outfall size and sonic bottom data provided by the City that shows sediment disturbance around outfalls. Sediment was collected from up to eight locations at the larger City outfalls and two ODOT locations; while as few as three locations were sampled at the smaller outfalls. Samples were evenly spaced along the length of bank areas with known contamination. Sample collection depth was 10 cm for targeted sites with the exception of two 5 cm samples that were co-located with 10 cm samples at Pacific Meats outfalls 1&2. The vertical

profile obtained in this area will help determine if higher concentration material is being “buried” by recently deposited sediment.



**OUTFALL 59-65  
DEQ SEDIMENT STUDY**

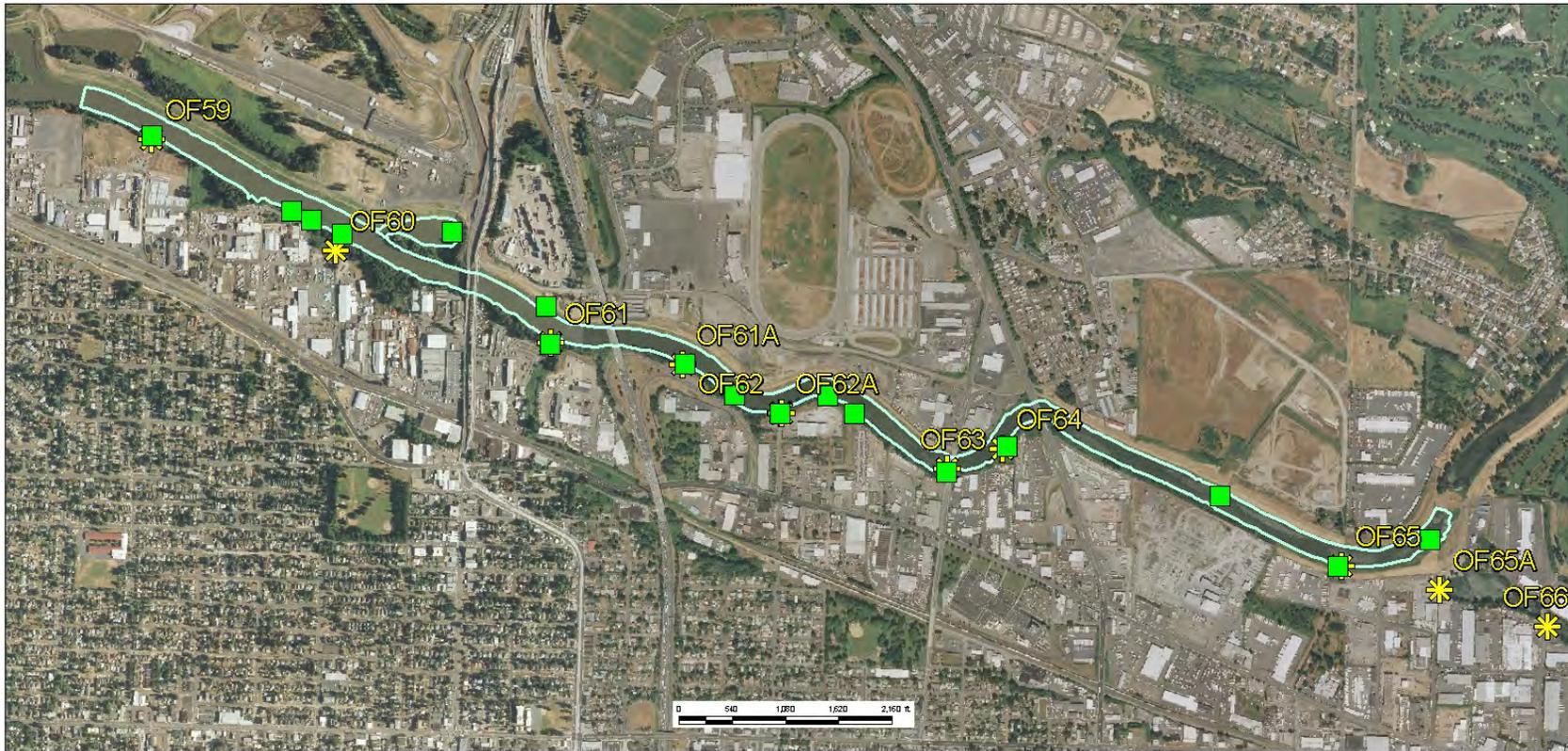


- DEQ2009\_1Ssample
- OF59-65channel
- Lower Columbia Slough study area

Base imagery ca. 2005 from Oregon Imagery Explorer web mapping service.  
Outfalls from City of Portland Bureau of Environmental Services.

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**Figure 6 Incremental Sample Locations**



**OUTFALL 59-65  
DEQ SEDIMENT STUDY**

- 2009 Targeted Samples
- ✱ City of Portland outfalls
- OF59-65channel

Base imagery ca. 2005 from Oregon Imagery Explorer web mapping service.  
Outfalls from City of Portland Bureau of Environmental Services.

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**Figure 7 Targeted Sample Locations**

### 3.3 Sampling Procedure

All samples were collected during a three week period in September 2009 during low water levels. The DEQ crew used an electric powered boat or wading to gain access to the sites. Samples were collected using an Ekman Dredge (Figure 8) at deep water sites or by stainless steel spoon or glass scoop (Figure 9) at shallow sites. Once samples were collected and labeled, they were placed on ice for preservation and transport to the laboratory.



Figure 8 Ekman Dredge Sampler



**Figure 9 Glass Scoop Sediment Sampler**

**Incremental Sampling:** At each increment point 50 g of sediment was weighed with a digital scale using the sampling methods described above. The sediment was then placed in the appropriate triplicate glass jar. The sediment was later homogenized at the laboratory.

**Targeted Samples:** Each sample consisted of three to eight composite elements and the total sediment mass required varied between a total of 400 g for chemical analyses only and 3400 g for bioassay and chemical analyses. Sample volume from each composite point depended on the number of composite points and the analyses to be performed. The field crew placed the sediment in labeled glass jars for analysis and the samples were later homogenized by the DEQ lab. Glass jars were separately filled for chemical analyses and bioassays from the same sample points. Field information was collected at each site including sample depth, sediment color, presence of woody debris, observation of oil sheen, and other information (see Appendix B).

### **3.4 Analytical Methods**

Sediment samples were sent to DEQ's Laboratory and Environmental Assessment Division (LEAD) and Pace Analytical, Inc for chemical analysis. Bioassay samples were sent to Northwestern Aquatic Sciences (NAS), based in Newport, Oregon.

#### **3.4.1 Chemical Analyses**

The chemical analyses chosen reflect legacy and current-use compounds found in Columbia Slough sediment. Previous studies were examined for historical comparisons. Tributyltin (TBT) and polybrominated diphenyl ethers (PBDEs) were new compounds not previously tested in

Columbia Slough sediment studies. Tributyltin (TBT) is the active ingredient of many products that act as biocides against a broad range of organisms. It is primarily used as an antifoulant paint additive on ship and boat hulls, docks, fishnets, and buoys to discourage the growth of marine organisms such as barnacles, bacteria, tubeworms, mussels and algae, though its use on small boats was prohibited in 1988. PBDEs have been used in a wide array of products as a flame retardant.

Metals, PCB Aroclors, pesticides, semi-volatile organic compounds (SVOCs) and grain size were analyzed in all samples. PCB congeners, TBT, PBDEs, total organic carbon (TOC) and bioassays were analyzed at selected sites, typically at larger diameter outfalls and in the incremental samples. See Table 1 for the list of analyses performed on each sample.

Once all samples were collected, LEAD prepared the sediment for chemical analysis which for the increment samples included drying, homogenization, sieving (<2mm), and sub-sampling. Some volatilization may occur for some semi-volatile organics and mercury during the preparation, although the volatilization is not expected to be substantial. Samples were then frozen to prolong holding time. The analytical methods used can be found in Table 2.

#### 3.4.2 Bioassay Analysis

Two freshwater sediment toxicity tests were performed on 12 sediment samples:

- 1) a 10-day *Chironomus dilutes* bioassay and
- 2) a 10-day *Hyalella azteca* bioassay.

These tests were chosen because they are commonly used freshwater bioassays which have been used on cleanup sites throughout Oregon. Bioassays samples were collected at selected sites (see Figure 10), usually larger city outfalls, see Appendix A for the list of sites.

Bioassay methods were based on the ASTM 2001 method and EPA method # 100.2. Samples were refrigerated in dark conditions at the DEQ laboratory until the end of the sampling. Samples were sent via UPS to Newport packed on ice. Once received at NAS, sample sediment was homogenized and split between the two analyses.

**Table 1 Sample Locations and Analyses Performed**

| Composite Name                                       | Label   | Latitude | Longitude  | Sample Strategy     | Depth | QA         | Analyses |              |              |           |       |              |       |     |            |          |
|--|---------|----------|------------|---------------------|-------|------------|----------|--------------|--------------|-----------|-------|--------------|-------|-----|------------|----------|
|  |         |          |            |                     |       |            | Metals   | PCB Aroclors | PCB Congener | Pesticide | SVOCs | Tributyl Tin | PBDEs | TOC | Grain Size | BioAssay |
|  |         |          |            |                     | cm    |            |          |              |              |           |       |              |       |     |            |          |
| Columbia Slough-Between Outfall 59-65                |         | 45.5866  | -122.6746  | Increment Sampling  | 10    | Triplicate | X        | X            | X            | X         | X     | X            | X     | X   | X          |          |
| Columbia Slough at PEN Levee 2 Pen2                  | Pen2    | 45.58338 | -122.64805 | Composite           | 10    |            | X        | X            |              | X         | X     |              |       |     | X          |          |
| Columbia Slough at City of Portland Outfall 65 OF65  | OF65    | 45.58259 | -122.65192 | Composite           | 10    |            | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |
| Columbia Slough at 13 <sup>th</sup> Ave. pumpstation | 13thPS  | 45.58435 | -122.65688 | Composite           | 10    |            | X        | X            |              | X         | X     |              |       |     | X          |          |
| Columbia Slough at City of Portland Outfall 64       | OF64    | 45.58548 | -122.66557 | Composite           | 10    |            | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |
| Columbia Slough at City of Portland Outfall 63       | OF63    | 45.58485 | -122.66749 | Composite           | 10    |            | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |
| Columbia Slough at Prec+BR                           | Prec+BR | 45.58627 | -122.67116 | Shoreline Composite | 10    | Duplicate  | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |
| Columbia Slough at Wastech                           | Wastech | 45.58684 | -122.67288 | Shoreline Composite | 10    |            | X        | X            |              | X         | X     |              |       | X   | X          | X        |
| Columbia Slough at City of Portland Outfall 62       | OF62    | 45.58640 | -122.67433 | Composite           | 10    |            | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |
| Columbia Slough at ODOT outfall                      | ODOT    | 45.58677 | -122.67637 | Composite           | 10    |            | X        | X            |              | X         | X     |              |       | X   | X          |          |
| Columbia Slough at ODOT Outfall 61A –                | OF61A   | 45.58774 | -122.67815 | Composite           | 10    |            | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |

| Composite Name                                 | Label        | Latitude | Longitude  | Sample Strategy     | Depth | QA        |  |  | Analyses |              |              |           |       |              |       |     |            |          |
|--|--------------|----------|------------|---------------------|-------|-----------|--|--|----------|--------------|--------------|-----------|-------|--------------|-------|-----|------------|----------|
|  |              |          |            |                     |       |           |  |  | Metals   | PCB Aroclors | PCB Congener | Pesticide | SVOCS | Tributyl Tin | PBDEs | TOC | Grain Size | BioAssay |
| Columbia Slough at Blasen                      | Blasen       | 45.58805 | -122.68234 | Shoreline Composite | 10    |           |  |  | X        | X            |              | X         | X     | X            |       | X   | X          |          |
| Columbia Slough at Schmeer Pumpstation         | Schmeer      | 45.58849 | -122.68250 | Composite           | 10    |           |  |  | X        | X            |              | X         | X     |              | X     | X   |            |          |
| Columbia Slough at City of Portland Outfall 61 | OF61         | 45.58838 | -122.68360 | Composite           | 10    | Duplicate |  |  | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |
| Columbia Slough at Simpson Cove                | Simpson Cove | 45.59103 | -122.68738 | Composite           | 10    |           |  |  | X        | X            |              | X         | X     |              | X     | X   |            |          |
| Columbia Slough at City of Portland Outfall 60 | OF60         | 45.59096 | -122.69197 | Composite           | 10    |           |  |  | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |
| Columbia Slough at PacificMeats outfall 1      | PM1-10       | 45.59131 | -122.69317 | Composite           | 10    |           |  |  | X        | X            |              | X         | X     |              | X     | X   | X          |          |
| Columbia Slough at PacificMeats outfall        | PM1-5        | 45.59131 | -122.69317 | Composite           | 5     |           |  |  | X        | X            |              | X         | X     |              | X     | X   |            |          |
| Columbia Slough at PacificMeats outfall 2      | PM2-10       | 45.59154 | -122.69378 | Composite           | 10    |           |  |  | X        | X            |              | X         | X     |              | X     | X   | X          |          |
| Columbia Slough at PacificMeats outfall 2      | PM2-5        | 45.59154 | -122.69378 | Composite           | 5     |           |  |  | X        | X            |              | X         | X     |              | X     | X   |            |          |
| Columbia Slough at PacificMeats outfall 3      | PM3          | 45.59194 | -122.69536 | Composite           | 10    |           |  |  | X        | X            |              | X         | X     |              | X     | X   |            |          |
| Columbia Slough at City of Portland Outfall 59 | OF59         | 45.59362 | -122.69936 | Composite           | 10    |           |  |  | X        | X            | X            | X         | X     | X            | X     | X   | X          | X        |

**Figure 10 Targeted Sample Locations-Bioassays**



**OUTFALL 59-65  
DEQ SEDIMENT STUDY**

Base imagery ca. 2005 from Oregon Imagery Explorer web mapping service.  
Outfalls from City of Portland Bureau of Environmental Services.



- + Bioassay Locations
- \* City of Portland outfalls
- \* Non-City Outfalls (BES, Dec08)
- Lower Columbia Slough study area

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**Table 2 Sampling and Analytical Methods**

## Sampling and Analytical Methods

| <b>Analysis</b>   | <b>Method</b>       | <b>Laboratory</b>         |
|---|---------------------|---------------------------|
| Metals by ICPMS, Total Recoverable                        | EPA 6020            | DEQ LEAD                  |
| Metals by ICPMS(Total Mercury)                            | EPA 7473            | DEQ LEAD                  |
| Organic: Aggregate Constituents & Properties              | EPA 9060            | DEQ LEAD                  |
| Grain size  | DEQ SOP             | DEQ LEAD                  |
| Polychlorinated Biphenyls(PCBs)                           | EPA 1668B           | Pace Analytical, Inc      |
| PCB Aroclors  | EPA 8082            | Pace Analytical, Inc      |
| Pesticides  | EPA 1699 & EPA 8081 | Pace Analytical, Inc      |
| Semi-volatile Organics                                    | EPA 8270D           | DEQ LEAD                  |
| Tributyl Tin(TBT)   | EPA 8232            | DEQ LEAD                  |
| PBDE  | EPA 1614            | Pace Analytical, Inc      |
| Bioassay: 10 day <i>Chironomous dilutes</i> Toxicity Test | NAS-XXX-CT4b        | Northwest Aquatic Science |
| Bioassay: 10 day <i>Hyaella azteca</i> Toxicity Test      | NAS-XXX-HA4b        | Northwest Aquatic Science |

From document DEQ09-LQ-0046-QAPP

### 3.5 Data Usability

Sample collection and analyses followed all methods, quality control procedures and reporting as described in the Quality Assurance Project Plan (DEQ09-LQ-0046-QAPP) (DEQ, 2009) and associated documents, except as noted. DEQ conducted a data validation check to ensure data acceptability. Below is a brief summary of data quality controls outlined in the QAPP and areas where the project deviated from the QAPP.

### 3.6 Summary of Quality Control

Three categories of quality control (QC) were used to assess data quality: sample design and methods, field sampling QC and analytical QC. Sampling designs were chosen to provide quality control for IS and targeted sampling.

#### **Sample Design and Methods:**

Integrated Samples: A triplicate sampling QC method, based on ‘*Draft Guidance on Multi-Increment Soil Sampling*’ Alaska Department of Environmental Conservation, 2009) was used for the incremental sampling. Early designs for the project included a more traditional approach to Multi-Increment® sampling with many decision units placed around individual outfalls. The projected field labor cost of collecting 30 or more increments and replicates at each outfall proved prohibitively expensive so a modified design was devised. The entire reach was our “decision unit” and enabled determination of a reach-wide average or background concentration. This length was also appropriate for evaluating bioaccumulation as it is consistent with the average home range of carp in the Lower Slough. Ordinary, targeted, composite samples collected at outfalls produced a reasonable assessment of the outfall’s potential to be a point source of contamination. Our IS procedure controls compositional heterogeneity and distributional heterogeneity, two major sources of error in sediment sampling. Grouping and segregation error is controlled by collecting multiple randomly located sample increments to address distributional heterogeneity. The collection of triplicate samples allows for the calculation of relative standard deviation ( $RSD=100 \times \text{Standard Deviation}/\text{mean}$ ) which provides an indication of how well the IS sample represents the decision unit. All results under 30% RSD were considered representative of the decision unit. If the RSD is greater than 35% the data distribution becomes ‘non-normal’ and the confidence of the results is weakened.



**Figure 11 Sample drying & waiting to be processed with IS analytical methods**

Targeted Samples: Standard duplicate samples were collected at a rate of 10 percent of the shoreline/outfall sites, in this case at 2 sites-one shoreline site and one outfall site. Relative percent difference was calculated to determine precision/accuracy of results. The analytical quality control target included precision of 20% for all analytes. Accuracy limits varied, depending on laboratory methods. The reporting limit was requested to be the lower of the Limit of Quantitation or Columbia Slough Risk Based Screening Levels. In some cases, estimates were reported and used. Analytical instruments were maintained and inspected within manufacturer's guidelines.

**Field Sampling QC:**

Daily Activities Log: Daily Activities were recorded on Chain of Custody forms and Columbia Slough Sediment Field Forms, which include date/time, location changes, how sample was collected, sample depth, odor and color. Field notes are summarized in Appendix B. Sample depth was expected to be 10 cm, however 28 of 71 samples (including increment samples) did not reach the 10 cm sample depth. The average depth for all 10 cm samples was 8.6 cm. It was noted that rockier/sandier sediments seemed to produce smaller depths. Wood debris and garbage also limited sample depth at less than 5 samples.

Location: Simpson Cove increment and targeted samples were moved to bank locations-as the cove was too shallow to boat and sediments too soft for wading. Pacific Meats Outfall #3 sites were also moved to the closest possible point that was boatable, due to shallow soft sediments. Updated locations can be found in Appendix B.

Minor tracking changes were noted on the field sheet, including crossed out bottle numbers and other minor typos.

Quality control in the field included decontamination of field sampling gear between sites, use of electric boat motor, and isolation of the sediment, whenever possible, to minimize cross contamination.

**Analytical QC:**

All analytical QA/QC was deemed acceptable, except for deviations noted below. Quality Control included duplicates, matrix spikes, laboratory control standards/calibrations and equipment calibrations. Dilutions were needed for several analyses to minimize interferences. Duplicate sample precision and relative percent difference for the incremental samples were evaluated and are discussed in Section 4.3.2.

Appropriate methods were used for all analyses, ensuring good comparability with other data. Analytical accuracy and precision were determined to be generally acceptable, with noted exceptions. Qualifiers were assigned to data points that exceeded project data criteria. All data

was reported and considered representative of the samples collected, except as noted by qualifiers.

Bioassay quality control included water quality testing, replicates and a control sediment test. Test results were considered acceptable, as controls were above the minimum acceptable survival rate of 80% - *Hyalella* (97.5% average survival rate); *Chironomus* (87.5% average survival rate).

### 3.7 Deviations from Quality Assurance Project Plan

Many analytical quality control concerns were attributed to matrix interference, i.e. sediment samples were very ‘dirty’ with many compounds. Sample dilution was required for pesticides, PCBs and PBDEs analyses. Method Reporting limits for pesticides were also much higher than normal. Laboratory matrix spikes failed high for several batch analyses, which may bias those results high or low. These results are qualified in Appendix F & G and flagged as JH for estimate biased high and JL for estimate biased low.

A significant delay in the chemical analyses at LEAD was encountered due to laboratory machinery break downs. As a result, it was determined that PCB, PBDE and pesticide analytical work that had not been completed should be sent to a contract laboratory, Pace Analytical. Since samples had been frozen the sediment was analyzed within the holding time period. The pesticide method was changed to EPA 8081 from EPA 1699 for the targeted sample sites because of budget concerns and indications from previous studies that pesticides are not a risk-driver for this section of the slough. Both pesticide methods EPA 1669 and EPA 8081 were performed on IS samples for method comparison.

Some estimates were used after careful consideration by the project managers. For example PCB Aroclor results were used although they were biased high because of interference from other co-eluting Aroclors.

**Targeted Samples:** Outfall 61 duplicate review determined 81% of analyses met acceptable criteria. Many duplicates for metals and SVOC compounds did not meet the 20% precision limit. The Precision+B&R duplicate passed 88% of the precision limits. The precision check for metals and PCBs passed, but some SVOCs and pesticides failed. Since the incremental samples passed QA/QC checks, DEQ concluded that laboratory methods produced acceptable results and that variability in other duplicate checks was due to a heterogeneous environment. This variability is expected in a diverse environment, based on previous studies within the slough (City of Portland 2009). Appendix E contains a complete review of targeted sample quality assurance.

**Incremental Samples:** For the IS samples, all results under 30% relative standard deviation (RSD) were considered representative of the decision unit. The QA criteria for a representative sample were met for 96% of the analytes. These data will be used as an average value for conditions across the decision unit. Total silver, tributyl tin (TBT), hexachlorobenzene, PCB 81, 2,4'-DDT and Aroclor 1254 did not pass the RSD evaluation. Aroclor 1254 RSD was at 38%- just above the acceptable confidence limit. This value will be used as a mean concentration across the decision unit, however there is some uncertainty associated with the value. The other substances are not risk-drivers in this section of the slough and, although there is some uncertainty in the result, these values can be considered estimates for the average conditions across the decision unit with little impact on subsequent environmental decisions. PBDE RSD's could not be fully calculated as there was not enough sample material to analyze PBDE in Triplicate C. Appendix E contains a complete list of incremental sample RSD values.

**Table 3 Incremental Sample Relative Standard Deviation (RSD) Failures**

| Analyte           | Relative Standard Deviation (%) |
|-------------------|---------------------------------|
| Total Silver      | 51                              |
| 2,4'-DDT          | 44                              |
| Hexachlorobenzene | 79                              |
| Aroclor 1254      | 38                              |
| PCB 81            | 32                              |
| Tributyltin       | 109                             |

*Values below 30% pass standard QC*

Results met the project completeness goal of 100% for targeted samples, after some collection locations were moved by the field crew because of safety concerns. The incremental sampling met the project completeness goal of 100% except for PBDEs, which were not a primary target analyte. Project managers believe the dataset is useable and meets the goals outlined in the Columbia Slough Outfall 59-65 Sediment Study. Future data users must evaluate whether this data is acceptable for their project's data quality objectives.

## **4.0 SAMPLE RESULTS**

Sample results are summarized in the following sections. IS results are presented first as they are used in subsequent data comparisons.

### **4.1 Increment Sample Method Evaluation and Comparison**

The IS data shows low variability between the three replicates. The QA criteria for a representative sample (results under 30 percent RSD) were met for 96 percent of the analytes. For comparison purposes, the mean values were calculated from the three IS replicates and then compared to mean values estimated from the discrete samples collected by the City of Portland in 2006 within the Lower Columbia Slough (City of Portland, 2009). The purpose of the comparison is to consider how well standard methods used for estimating average concentrations over an area compare to more rigorous values generated by the IS approach.

In 2006, the City collected 23 sediment samples in the Lower Slough. Among the analytes included in the investigation were Aroclors by Method 8082, pesticides by a modified version of method 8081 and metals by EPA 6020. The results presented in Table 4 show that the IS samples were relatively consistent with the City sampling averages.

The comparison is not intended to be comprehensive and the evaluation was limited to selected chemicals in common between the two sampling events. DEQ evaluated metals, PCBs as represented by Aroclor 1254, and the organochlorine pesticides 4,4'-DDD, 4,4'-DDE, and dieldrin. These representative substances were selected because they are some of the most commonly detected environmental contaminants present in the Lower Slough, and they represent the majority of human health or ecological risk.

In the case of Aroclor 1254, the detection frequency was low at 38% (10 of 26 samples). This low detection frequency, even in the lower slough where PCBs are a significant contaminant is not surprising, given the analytical method used was EPA 8082. This method relies on pattern matching, and if the basic Aroclor patterns are not matched, PCBs are not quantified (EPA 2002). It is probable that any PCBs detected in Slough sediment are highly weathered and some of the PCBs will not match the analytical standards, potentially accounting for the relatively low frequency of detection, even though the detection limit in the 2006 investigation was very good, at 0.16 ug/kg.

For the organic analytes included in the 2006 investigation (Table 4), the analytical results are positively skewed, approximating a log-normal distribution. Therefore, the non-detected values were replaced by estimated “imputed” values using US EPAs ProUCL software (EPA 2010). Each of the metals evaluated were detected in every sample. A few descriptive statistics were calculated using these data and shown in Table 4.

**Table 4 Comparison of City 2006 Sediment Data relative to 2009 IS samples.**

| Analyte                 | 2006 Discrete Sample Summary |                     |      | 2009 IS summary |      |      |         |                    | Previous Slough baseline values |
|-------------------------|------------------------------|---------------------|------|-----------------|------|------|---------|--------------------|---------------------------------|
|                         | CV <sup>1</sup>              | Median <sup>2</sup> | Mean | IS A            | IS B | IS C | IS mean | Ratio <sup>3</sup> |                                 |
| <b>Organics (ug/kg)</b> |                              |                     |      |                 |      |      |         |                    |                                 |
| Aroclor 1254            | 217                          | 5.8                 | 33.5 | 27.3            | 16.5 | 37   | 27      | 0.8                | 24                              |
| Dieldrin                | 136                          | 0.82                | 1.3  | 0.64            | 0.76 | 0.69 | 0.70    | 0.54               | NA                              |
| DDE                     | 78                           | 7.3                 | 8.3  | 5.9             | 6.9  | 6.1  | 6.3     | 0.74               | 6.1                             |
| DDD                     | 154                          | 2.8                 | 4.1  | 3.1             | 4.4  | 3.5  | 3.7     | 0.9                | 7                               |
| <b>Metals (mg/kg)</b>   |                              |                     |      |                 |      |      |         |                    |                                 |
| Arsenic                 | 37                           | 6.6                 | 6.8  | 5.2             | 5.2  | 5.6  | 5.3     | 0.78               | 8                               |
| Copper                  | 59                           | 48                  | 47   | 37              | 34   | 41   | 37.5    | 0.8                | 54                              |
| Lead                    | 83                           | 48                  | 59   | 40.8            | 37.6 | 45.5 | 41      | 0.69               | 90                              |
| Zinc                    | 55                           | 273                 | 269  | 244             | 227  | 261  | 244     | 0.91               | 314                             |

1. CV- Coefficient of variation expressed as a percentage of the mean. The CV is presented as a measure of variation in the data. It illustrates which analytes have relatively more or less variability, and more or less certainty.
2. The median is the value at which 50 percent of the observed concentrations are both above and below the value. It is resistant to outliers, and, when shown with the mean, provides an indication of whether the data are symmetric or positively skewed. It is often used as a better indicator of where most of the data are in skewed data sets (e.g., dieldrin).
3. Ratio of IS mean to discrete sample mean.

The results presented in Table 4 show that the IS samples were relatively consistent between replicates indicating the method was successful in reducing sediment sample heterogeneities. The mean from the three IS replicate samples is compared to the mean from the 23 (non-duplicate) discrete samples collected in 2006 generating the ratio of the two means. The comparison shows that for this group of analytes the IS estimates of mean were consistently lower than the average of discrete samples. The difference was less for the metals and greater for the organics - up to a 54 percent lower estimate of the mean for dieldrin. It should be noted that some of the dieldrin values were non-detect and had to be estimated for this comparison using “imputation” methods (US EPA 2010), so the true dieldrin concentration could be slightly to somewhat lower than that estimated by discrete samples.

Comparison of mean and median values for the two sample set shows that for metals these two statistics are similar, indicating that they both indicate similar average concentration. For the bioaccumulative organics, the medians are consistently lower than the IS mean and the coefficients of variation are relatively high. This is typical of positively skewed environmental data. In these cases, a median may be a better estimate of average concentrations and this appears to be

supported by comparison to the IS collected means, with the exception of Arcolor 1254. The reason why Aroclor behaves differently than the other organics is unknown, but may be due to this analyte representing a complex mixture and the analytical methodology.

The confidence intervals or error bars associated with the 2006 discrete and 2009 IS mean values are compared in Table 5.

**Table 5: 95 Percent Confidence Intervals for Select Analytes in both Datasets**

| 95 Percent Confidence Intervals <sup>a</sup> |                        |                        |                         |                  |
|--|------------------------|------------------------|-------------------------|------------------|
| Analyte                                      | 2006 Discrete Interval | Discrete Interval Size | 2009 IS Sample Interval | IS Interval Size |
| <b>Organics (ug/kg)</b>                      |                        |                        |                         |                  |
| Aroclor 1254                                 | 2.1-65                 | 63                     | 1.8 - 52                | 50.2             |
| Dieldrin                                     | 0.52 - 2.0             | 1.5                    | 0.55 -0.85              | 0.3              |
| DDE  | 5.5 - 11               | 5.5                    | 5.1 - 7.5               | 2.4              |
| DDD  | 1.3 - 6.9              | 5.6                    | 2.0 - 5.3               | 3.3              |
| <b>Metals (mg/kg)</b>                        |                        |                        |                         |                  |
| Arsenic                                      | 5.7 - 7.9              | 2.2                    | 4.8 - 5.9               | 1.1              |
| Copper                                       | 35 - 59                | 24                     | 8.9 - 10.7              | 1.8              |
| Lead   | 38 - 80                | 42                     | 31 - 54                 | 23               |
| Zinc   | 205 - 334              | 129                    | 202- 286                | 84               |

a - Interpreted as meaning that the true mean sediment concentration has a 95% chance of being within the interval shown. Based on 23 2006 samples (excluding three duplicates).

There was a significant improvement in the confidence interval for the IS dataset. The improvement is especially notable for pesticides, with the IS interval between 20-60 percent of the discrete sample interval. For the metals, the IS intervals were about ½ the size of the discrete samples interval and substantially better for copper. Overall, the results indicate a significant improvement in estimation of mean concentration using the IS methodology. Although the level of replication was low with only three samples, the quality of the data is high and the method is successful in reducing sampling heterogeneities. A greater level of confidence can be placed in the estimated mean concentrations of the IS dataset.

For comparison, Slough-wide baseline values (DEQ 2002) calculated using the sediment data collected throughout the entire Slough in 1994/1995 (Parametrix, 1995) are also shown. Surprisingly, the historical data generated average concentrations for the organics shown that are very consistent with the IS averages determined in this study. Metals concentrations are consistently lower, but on the same order of magnitude as the historically determined baseline values. The IS data set for the Lower Slough reflects more recent data with a higher level of confidence and is more representative of ambient concentrations in the Lower Slough. In the Lower Slough the average concentration values for the IS data set will replace the baseline

values previously developed for the entire Columbia Slough. As expected, the average concentrations (IS data) for some bioaccumulative metals, SVOCs, PCBs and pesticides were above DEQ bioaccumulative risk-based sediment screening levels (DEQ 2007). As outlined in the 2005 Columbia Slough ROD, active remediation of sediment in the Slough to concentrations below baseline is likely to be infeasible due to potential for recontamination. However, a combination of active remediation of sediment exceeding baseline concentrations, source control, and natural recovery is expected to be effective in eventually reaching risk-based cleanup levels.

Another comparison of interest was completed by calculating the area-weighted average concentration for two compounds, total PCBs and dieldren, using the discrete samples present in the Slough segment, including data collected in this study, the City of Portland 2006 study, Pacific Meat data, and ODOT data. The method involved creating polygons around each sample point that are sized based on proximity to the next closest sample point in each direction. The concentration at the sample point was assumed to apply to the entire polygon surrounding it. This process generated a segment-wide weighted average concentration for PCBs of 92.65 ug/kg and for dieldren of 1.82 ug/kg. This compares to the IS averages of 45.5 and 0.7 ug/kg respectively. The higher values for the area-weighted concentrations likely reflects the fact that discrete samples were targeted to release areas which are more contaminated, as discussed in the next section, and by this method were assumed to represent a disproportionate amount of the sediment.

## **4.2 Targeted Samples**

The sediment chemistry results for targeted sites were elevated in comparison to the average IS concentrations in the Lower Slough study area. The following sections discuss contaminants of concern for the Lower Slough and summarize detection characteristics at the targeted source areas. Appendix C provides details on all analytes.

### **4.2.1. Detection Rates for Analyte Classes**

Naturally occurring metals were detected in nearly all samples, ranging from 100 percent for common metals to 46 percent for rare metals, such as antimony. The most common PAHs were found in nearly all samples. PCB Aroclor 1260 was found in 81 percent of sampled sites and PCB Aroclor 1254 was found in 54 percent of sites. Pesticides were analyzed using two methods. The standard lower resolution Method 8081 used for targeted samples detected few analytes. Method 1669, a higher resolution method used for the IS triplicates, detected low concentrations of numerous pesticides, including DDT, DDE and DDD.

A smaller sub-set of 16 samples was analyzed for tributyltin (TBT), PBDEs and PCB congeners. These tests were included on an exploratory basis to assess the value of conducting more comprehensive sampling for these analytes in the future. TBT was found in 63 percent of the

samples analyzed but was generally detected at concentrations below risk-based screening levels. PCB Congeners and PBDEs were found in all analyzed samples. There is currently no screening value for PBDEs. Total PCB congeners were generally lower than the Aroclor sum for the same sample.

#### 4.3.2. Common Contaminants at Targeted Sites

Contaminants of concern (COCs) were selected based on their toxicity and frequency of detection throughout the reach. Standard plots were generated to display the data for the targeted sites relative to the IS mean (reach average or baseline) and the Columbia Slough Sediment Screening values (risk-based values). Targeted site location acronyms (x axis) are defined in Table 1.

#### Metals

Chromium: Targeted chromium values for the reach were generally below risk-based values. The average concentration of chromium at targeted locations was 61 ppm, slightly above the IS mean of 44 ppm and risk-based value of 42 ppm. The sample at Outfall 59 was an outlier at 481 ppm.

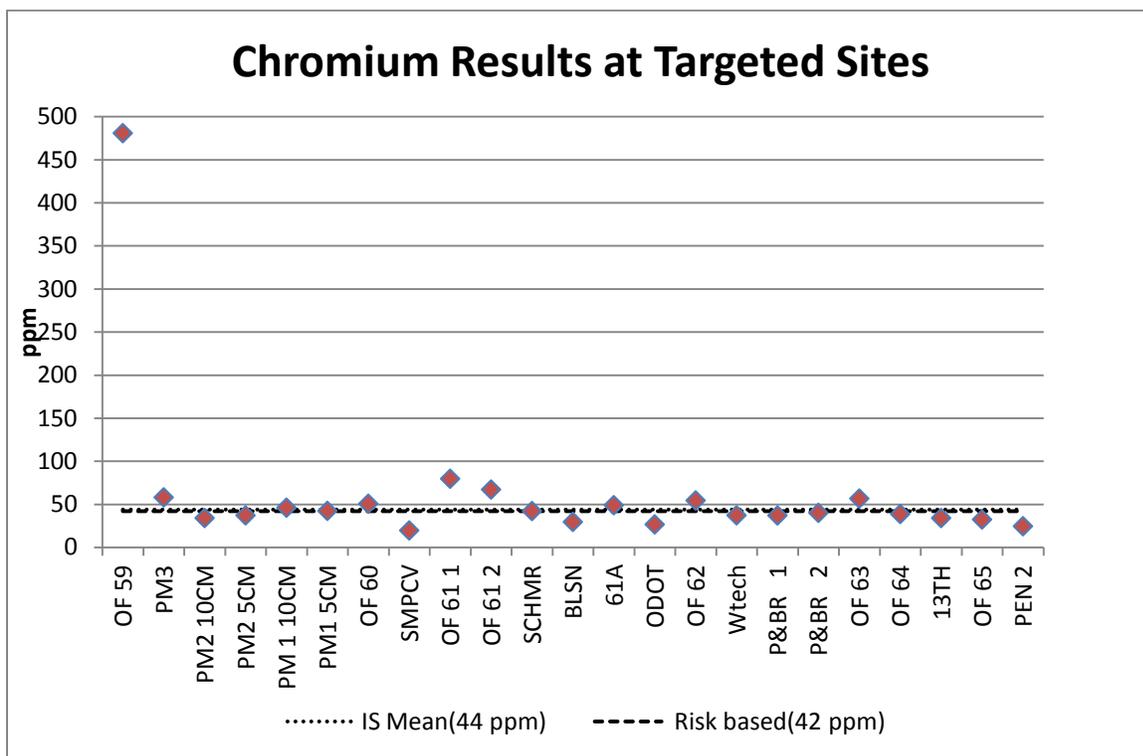
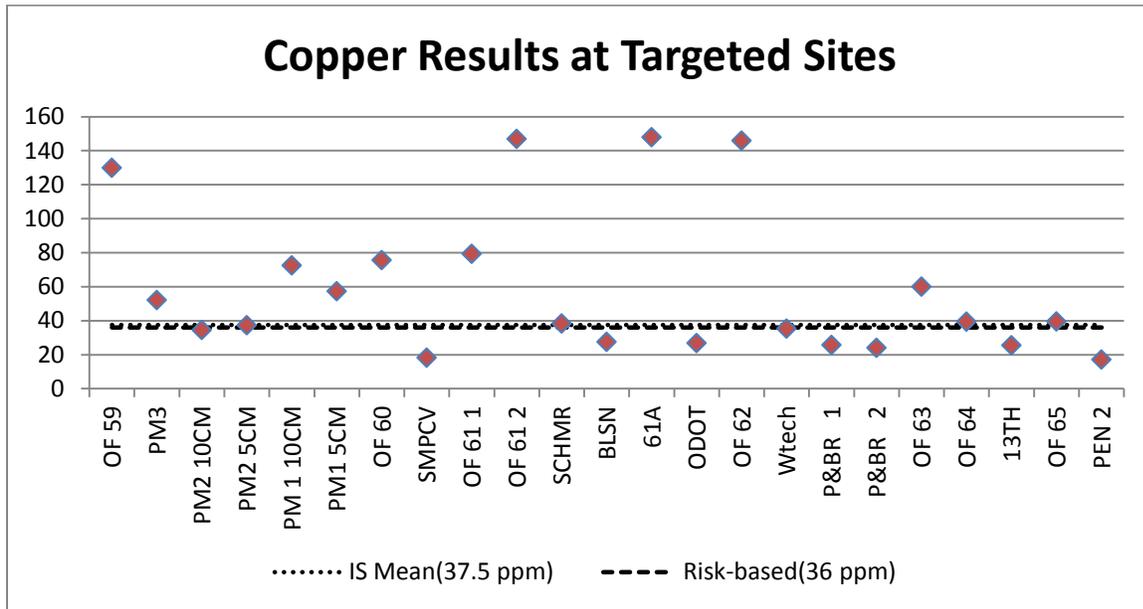


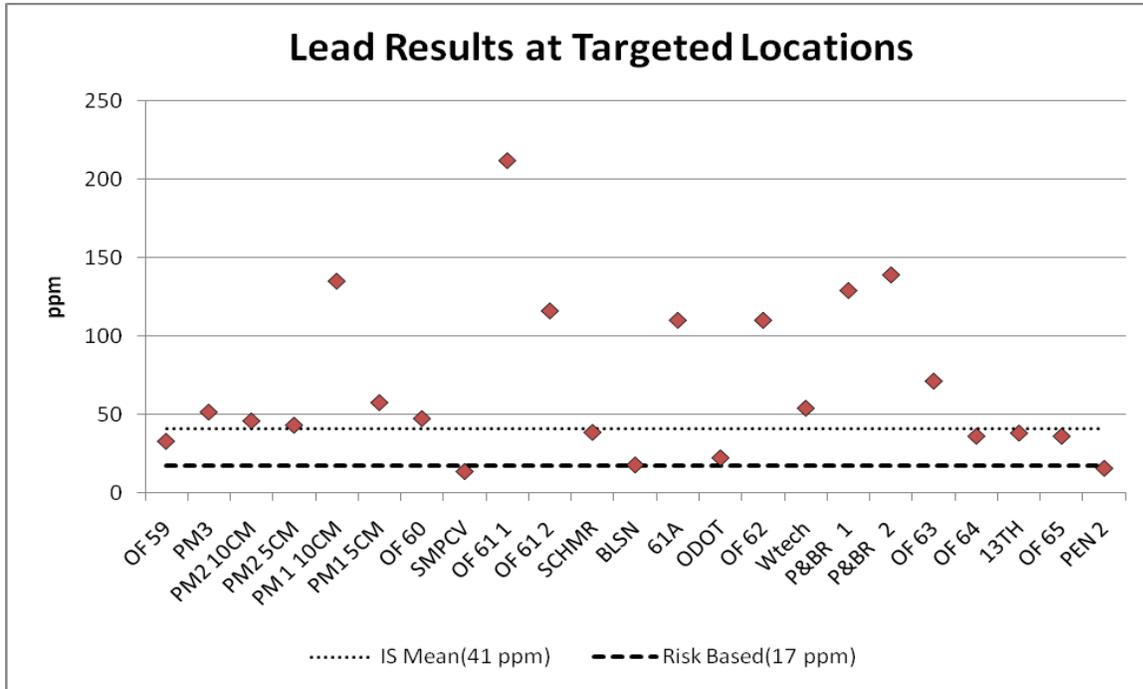
Figure 12 Chromium Results at Targeted Samples

Copper: About 62 percent of targeted copper samples had concentrations that exceeded the risk-based level. The average concentration of targeted samples was 59 ppm, again, above the IS (37.5 ppm) and risk-based (36 ppm) values. Samples at Outfalls #59, 61, 61A and 62 had copper concentrations between 130-148 ppm.



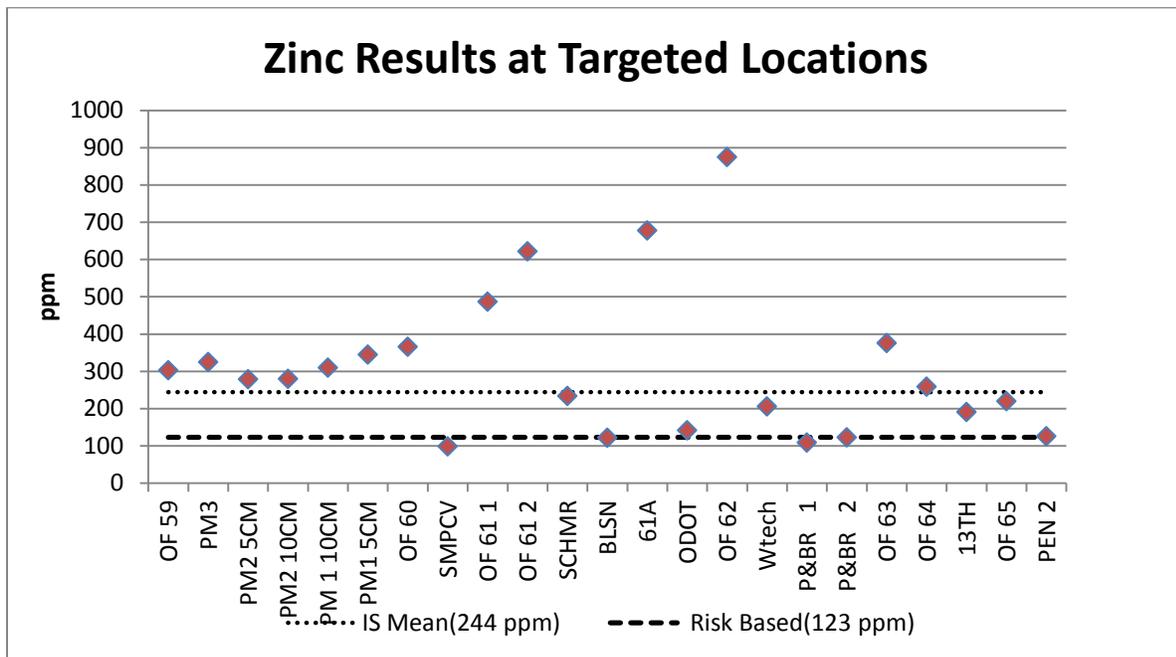
**Figure 13 Copper Values for Targeted Samples**

Lead: Lead was detected at all sites and the average lead concentration of targeted samples in the reach was 68 ppm, above the IS (41 ppm) and risk-based (17 ppm) values. All targeted sites except two exceeded risk-based concentrations.



**Figure 14 Lead Values for Targeted Samples**

Zinc: Zinc concentrations exceeded the risk based value at 54 percent of the targeted sites, with an average concentration of 304 ppm. Samples at outfalls #s 60, 61, 61A, 62, 63 and Pacific Meats #1 & #3 all exceeded the risk base value.



**Figure 15 Zinc Results for Targeted Sites**

## Total Poly-Aromatic Hydrocarbons (PAHs)

Total PAH is the sum of the concentrations of all individual PAHs at each target site. For calculation simplicity, if an analyte was non-detect, the analytical detection limit was used in the calculation. A total PAH risk-based value has not been calculated.

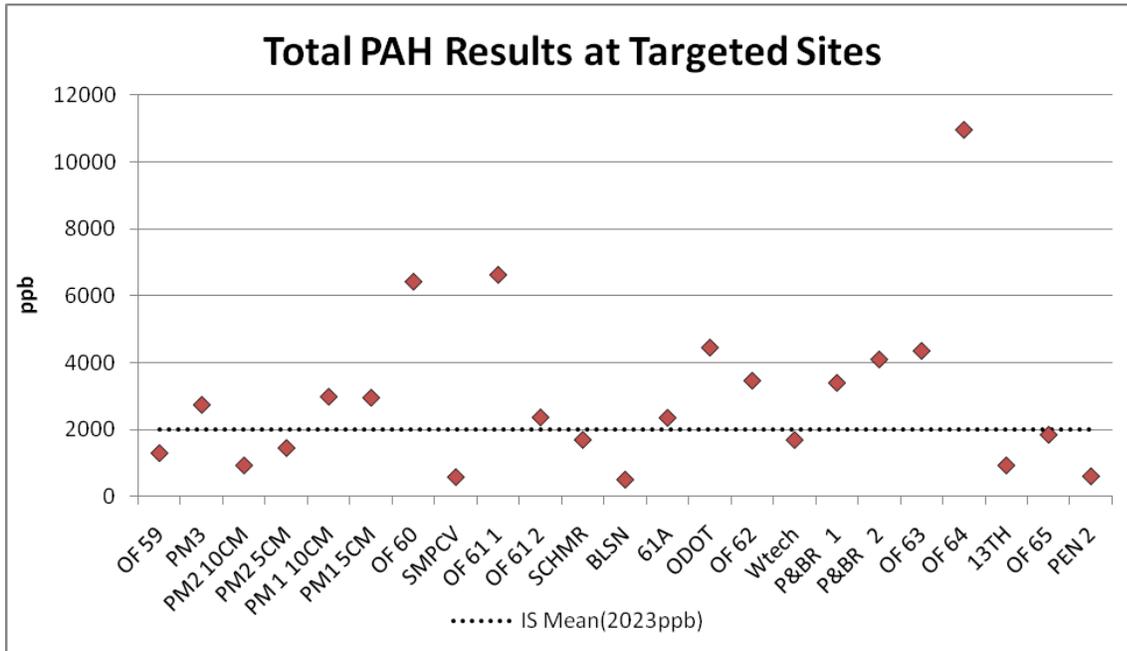


Figure 16 Total PAH's for Targeted Sites

**PCBs:**

Aroclor 1254: The risk-based concentration and the IS average values were similar for Aroclor 1254. Consequently, targeted concentrations were either above or below both values. It should be noted, however, that the risk-based value from the 2007 Columbia Slough Screening Level Table is not truly risk-based in that it reflects the standard analytical detection limit for Aroclor 1254. The true risk-based value is below this. Consequently, any detection of Aroclor 1254 is above the true risk-based value. See discussion of site-specific risk-based value for this compound presented in Section 5. The highest targeted values were near the Pacific Meat cleanup site (ESCI#145) and Outfall 60, which is the closest city outfall to Pacific Meat.

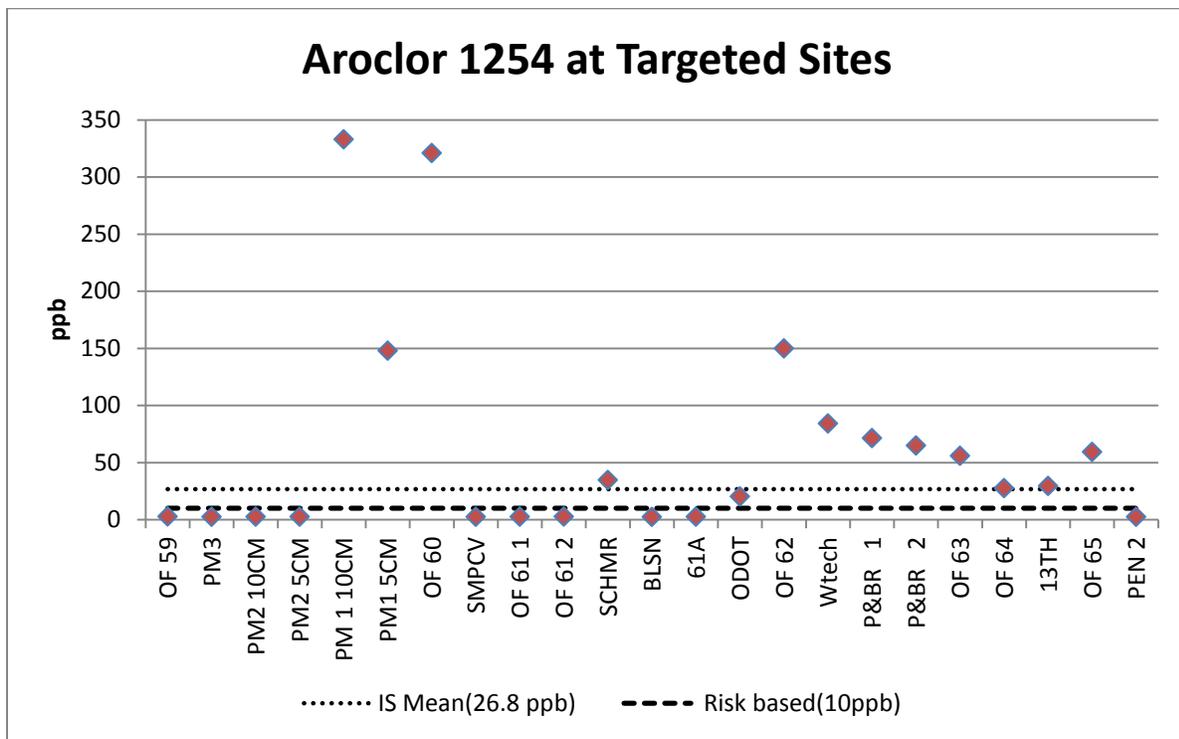
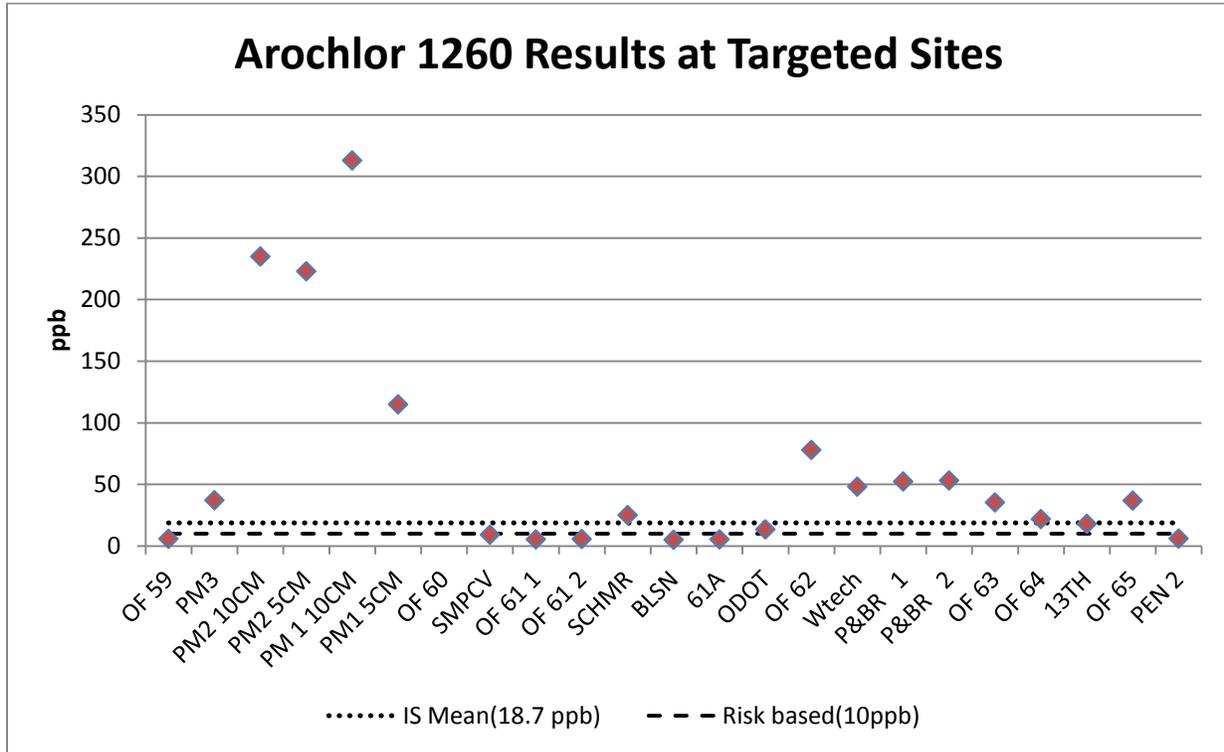


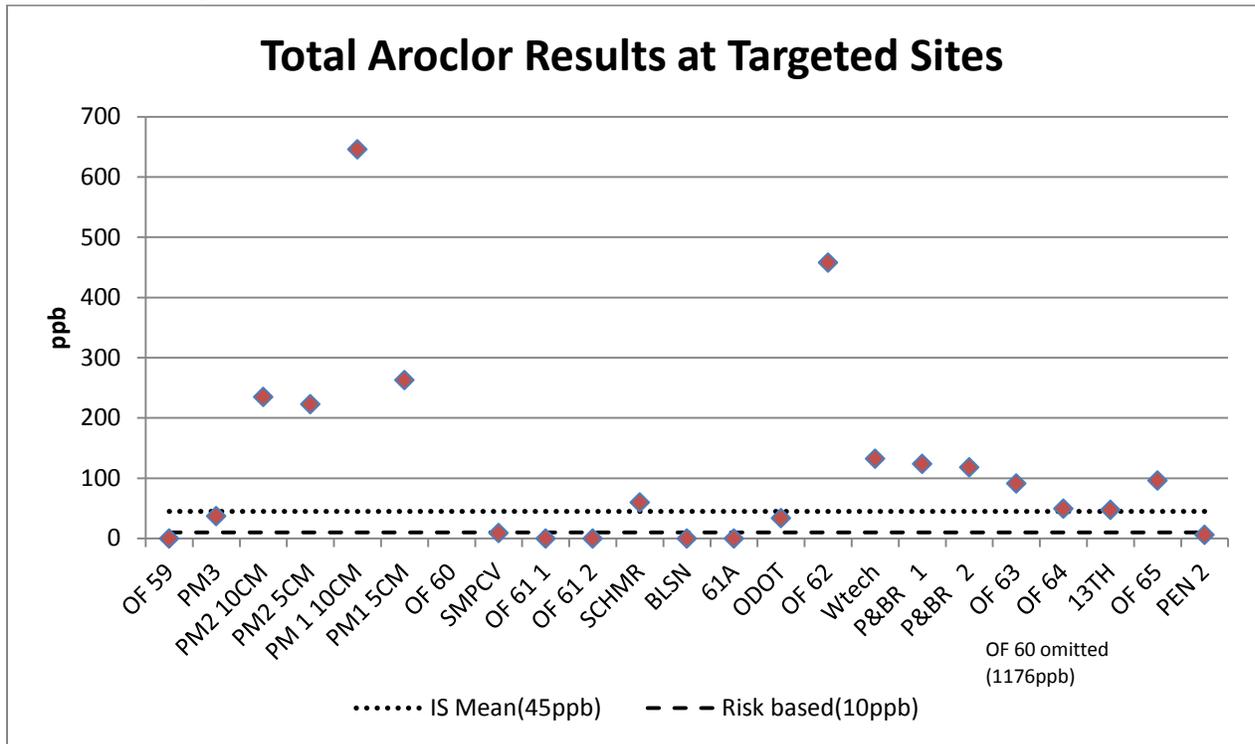
Figure 17 Aroclor 1254 at Targeted Sites

**Aroclor 1260:** As with Aroclor 1254, the IS average reach value was close to the standard analytical detection limit of 10 ppb for Aroclor 1260 and targeted concentrations were either above or below both values. The highest targeted value was located near Outfall 60 (855ppb) and omitted from the figure below, due to difficulty in showing this value on the graph.



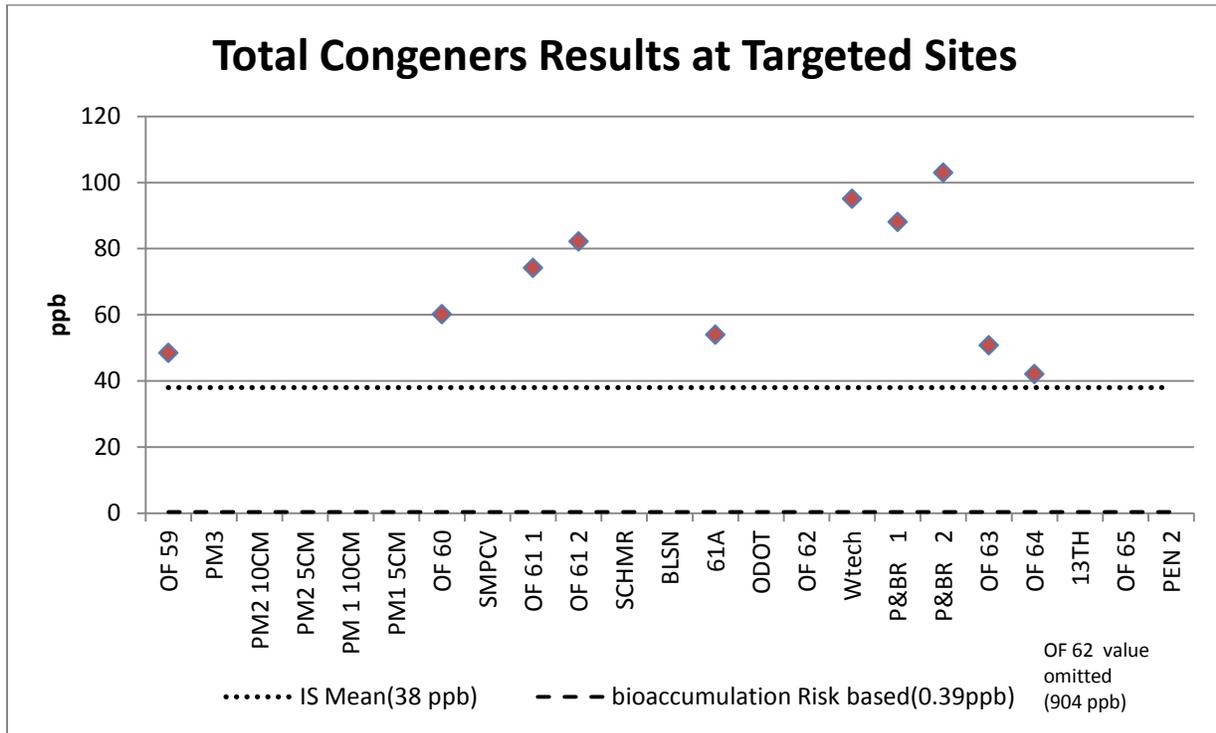
**Figure 18 Arochlor 1260 Results at Targeted Sites**

**Total PCB Aroclors:** Total Aroclor values were similar to those in individual Aroclor graphs above. Outfall 62 was unique as results showed additional Aroclor 1248 contamination, where analyses of samples for all other outfalls in the study only found Aroclor 1254 and 1260. The Outfall 60 value (1176 ppb) was omitted from figure so values near the risk based number could be viewed easily.



**Figure 19 Total Aroclor Results at Targeted Sites**

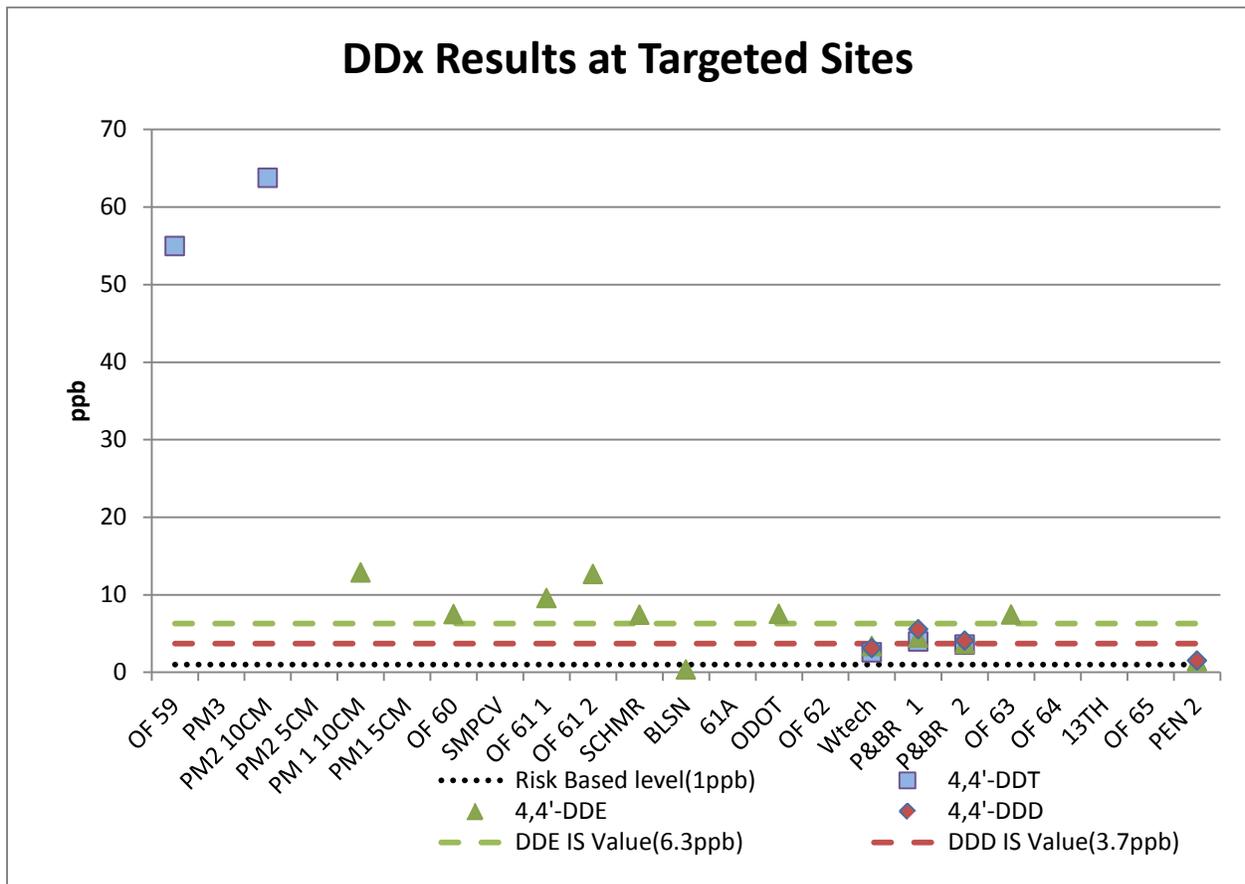
Total PCB Congeners: Total PCB congeners were also measured for a subset of samples. All results were below 103 ppb, with the exception of that for Outfall 62 which had a concentration of 904 ppb. This data point was omitted from the graph below to limit graph compression. The congener method is able to produce lower detection limits than the Aroclor method, therefore the bioaccumulation risk-based value was used for a comparison.



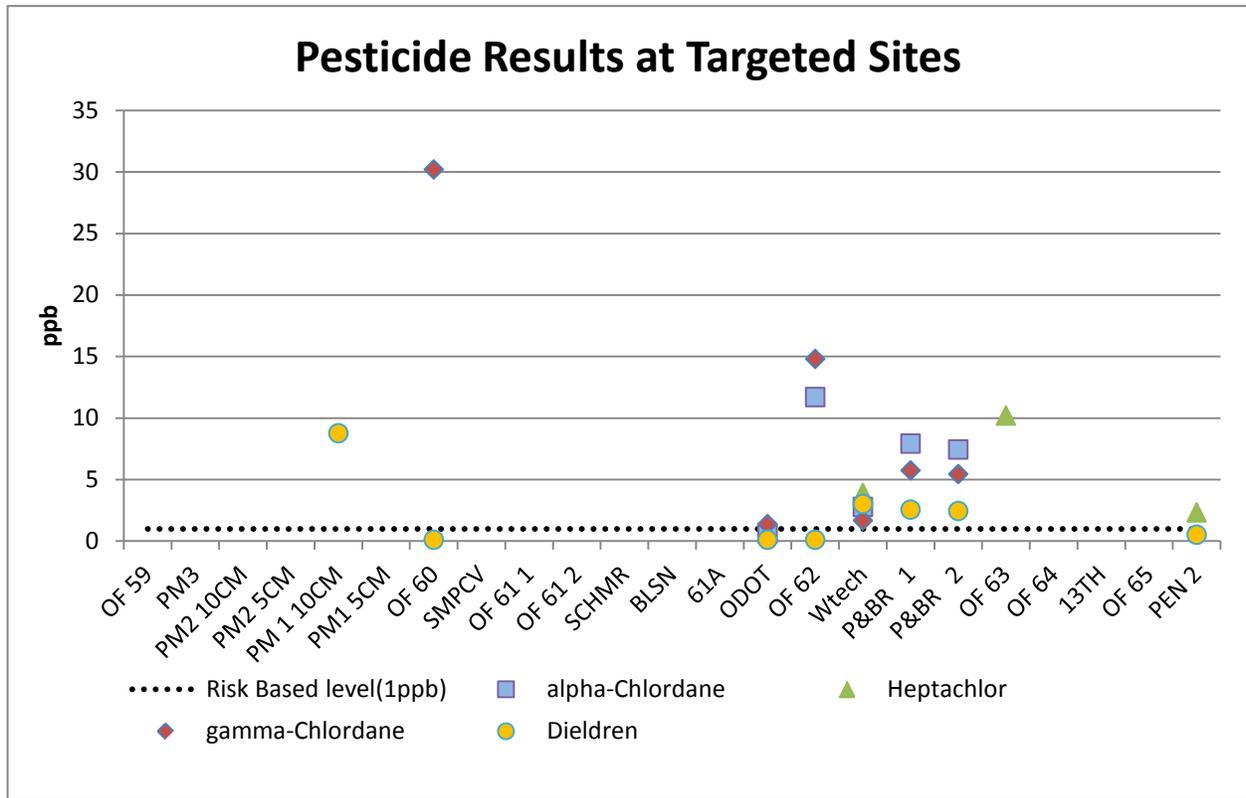
**Figure 20 Total Congeners Results at Targeted Sites**

**Pesticides:**

Pesticides were generally found at targeted locations; however, method detection limits for pesticides were above the detection limit referenced in the 2007 Slough Screening Level Table of 1 ppb for many targeted sites and many of the pesticide results were estimated. The IS average for the reach was below the “risk-” based values for all pesticides except DDD and DDE. Note again that the risk-based value was set at the typically achievable concentration of 1 ppb. The true risk-based value is below 1 ppb. Although pesticides were analyzed at all sites, only detected values are graphed in Figure 21 and 22.



**Figure 21 DDx Pesticide Results at Targeted Sites**

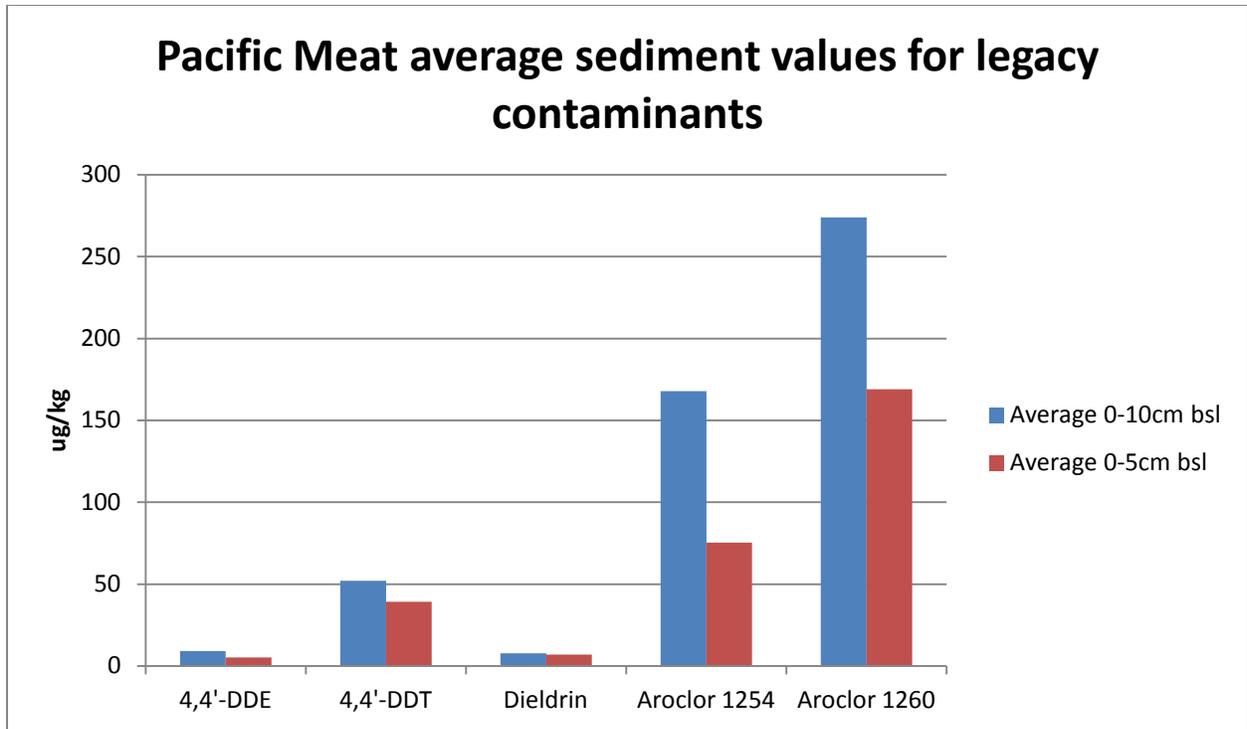


**Figure 22 Pesticide Results at Targeted Sites**

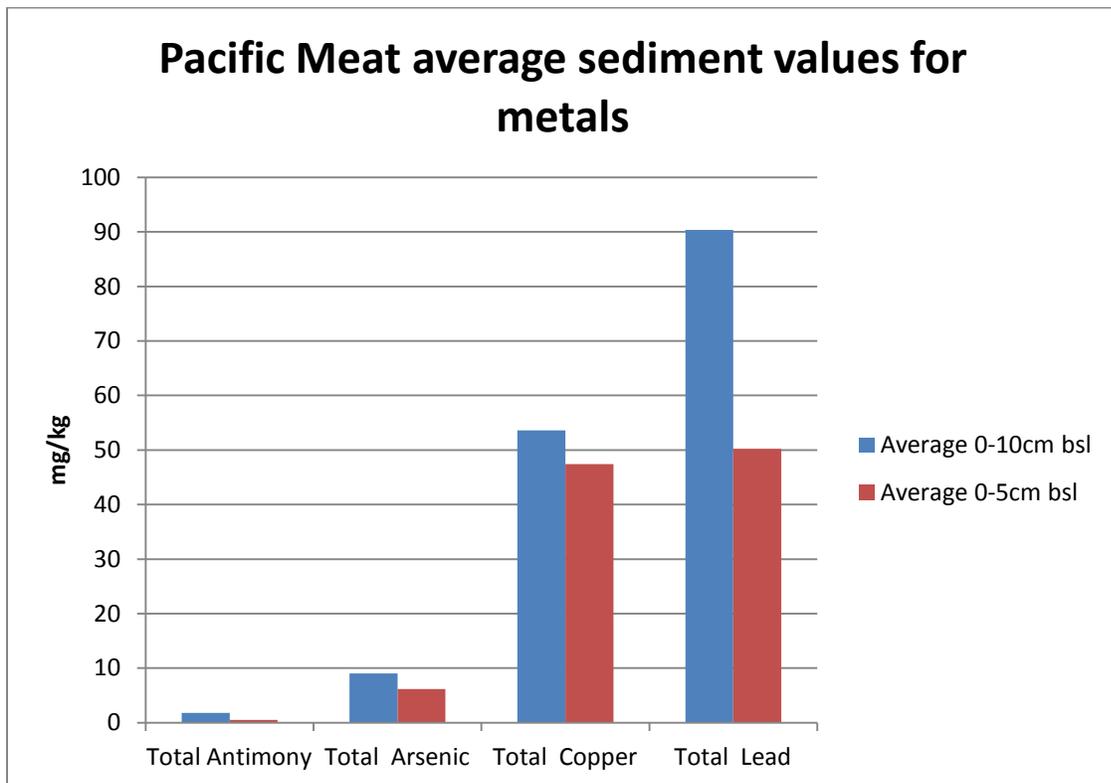
#### 4.2.3 Vertical Profile Data

Sediment samples were collected at 0-5 cm and 0-10 cm depths at Pacific Meats, an area of known PCB and pesticide contamination, to determine if natural deposition is covering older contaminated sediment with cleaner material. Previous sampling has shown that contaminant concentrations from 1-2 feet below the mud line are up to an order of magnitude higher than concentrations detected in the interval from 0-1 feet. In this study the 0-10 cm below surface level (bsl) and 0-5 cm bsl values from both outfalls were combined to determine average concentrations at these depths.

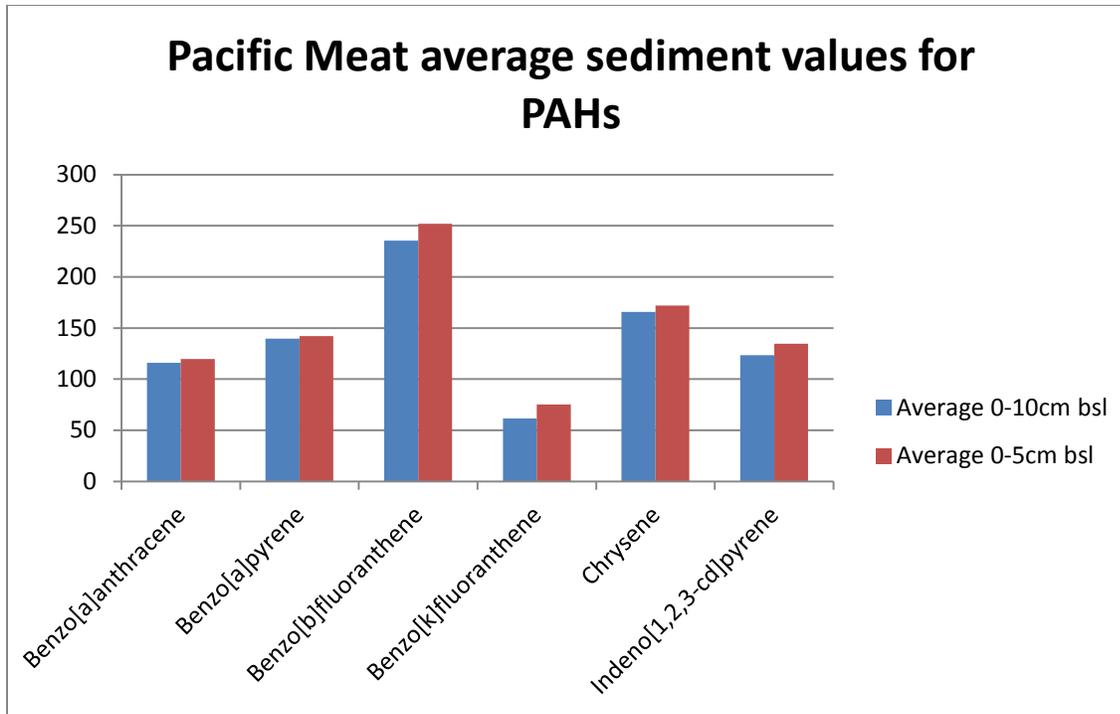
The graphs below show that PCB and metal contaminants are considerably lower in concentration in the shallow sediment interval. To a lesser extent, this is also true for pesticides. Selected PAH's show a slight concentration increase in shallower sediments, as current on-site source control has yet to be implemented. The results suggest that many contaminants and particularly legacy contaminants such as PCBs and pesticides, are being covered up by cleaner sediment.



**Figure 23 Average Sediment Concentrations at Different Depths: Legacy Contaminants**



**Figure 24 Average Sediment Concentrations at Different Depths: Selected Metals**



**Figure 25 Average Sediment Concentrations at Different Depths: Selected PAHs**

### 4.3 Bioassay Summary

Two acute bioassay tests were performed on sediment samples collected near major outfalls. Locations are shown in Figure 10 and described in Table 6. The two toxicity tests were the 10-day mortality sediment toxicity test with the amphipod *Hyalella azteca* and the 10-day growth and mortality sediment toxicity test with the midge *Chironomus dilutus*. Details of these tests are documented in the laboratory reports from Northwest Aquatic Sciences (NAS 2009 a,b). Bioassay results were evaluated by DEQ for statistical significance of responses associated with test sediment relative to those observed in the laboratory control, and for magnitude of the bioassay endpoint response. Prior to interpreting the bioassay responses, the mean responses of each of the eight replicates were normalized to the control sediment response by dividing each mean test sample response by the control sample response. The normalized data were then evaluated for statistical significance. Because bioassay response data cannot meet the conditions required for use of parametric hypothesis tests, the tests for statistical significance were performed by using both the nonparametric hypothesis test referred to as the Mann-Whitney test on the median, and the robust permutation test on the mean response (Higgins 2004). The conclusions are summarized below. Results were identical when using the median and mean responses (Table 6).

Effect size was evaluated using the reference envelope approach which has also been used to interpret Portland Harbor Sediment Toxicity Tests (EPA 2009). This approach pools all toxicity

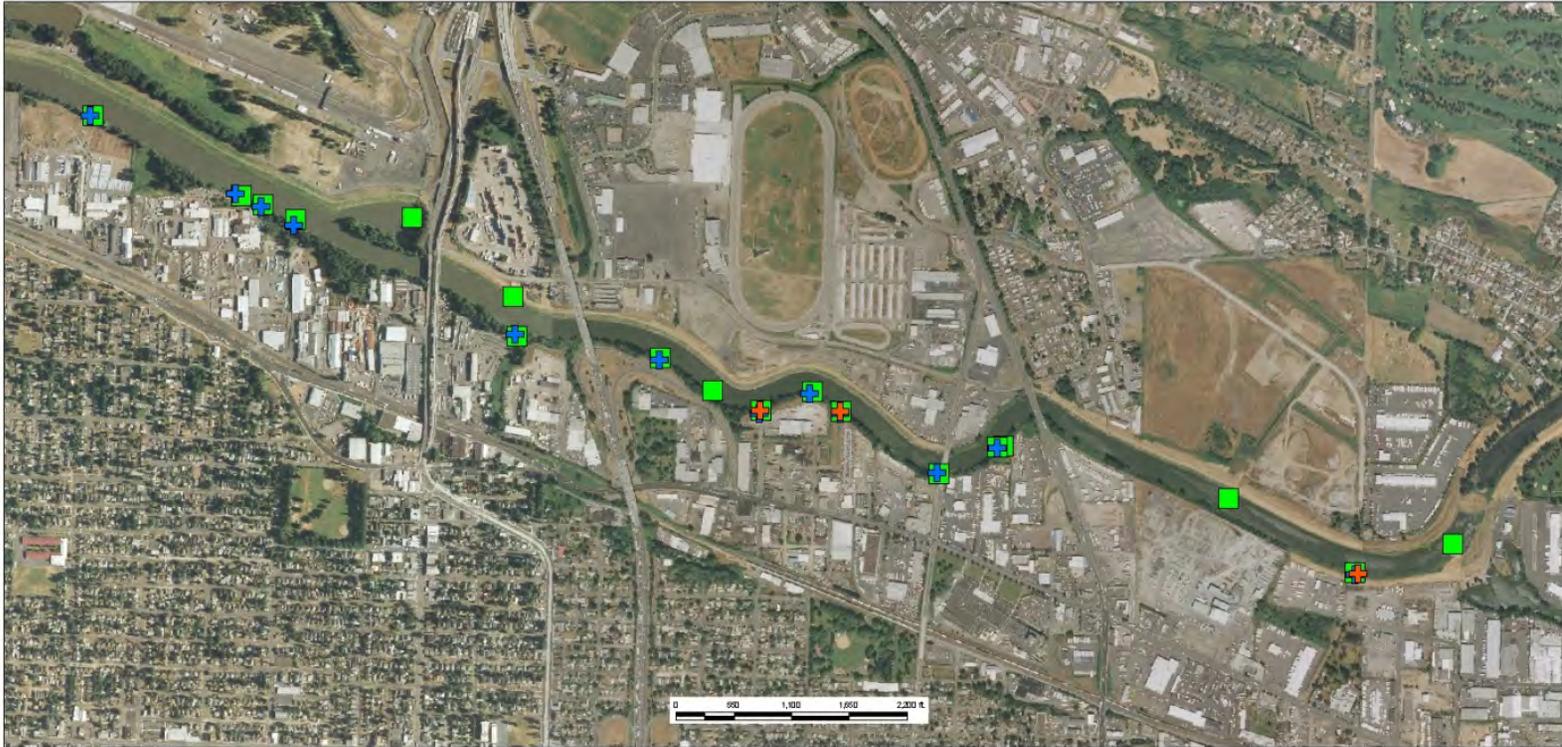
test results to better identify where impacts are observed. Developed as described in EPA 2009, it is based on a large number of reference area bioassay in the lower Willamette and provides a good regional reference point. The reference envelope was used to differentiate the effect size into three categories - no effect, minor effect, and severe effect. The results are presented in Table 6 and the details of the interpretation are presented in Appendix H. As shown in Table 6, there were statistically significant effects in the *Chironomus* mortality test for four samples. Of these four samples, three of them also showed an effect in the growth endpoint. No samples showed any effect for the *Hyalella* mortality endpoint. Overall, the three samples with significant responses in both *Chironomus* tests are deemed to demonstrate a toxic effect. Because minor effects were observed in both *Chironomus* tests in the OF 62 sample, this sample was interpreted to be possibly toxic.

**Table 6 Results of Sediment Bioassay Tests.**

| Site         | <i>Hyalella</i>                      |                                 | <i>Chironomus</i>                    |                                 |                                      | Overall Interpretation |
|--------------|--------------------------------------|---------------------------------|--------------------------------------|---------------------------------|--------------------------------------|------------------------|
|              | Percent Survival                     | Reference Envelope Effect Level | Percent Survival                     | Reference Envelope Effect Level | Growth (Biomass)                     |                        |
|              | Statistically Different <sup>a</sup> |                                 | Statistically Different <sup>a</sup> |                                 | Statistically Different <sup>a</sup> |                        |
| Control      | N/A                                  | N/A                             | N/A                                  | N/A                             | N/A                                  | N/A                    |
| OF 65        | No                                   | No Effect                       | Yes                                  | Severe Effect                   | No                                   | Toxic                  |
| OF 63        | No                                   | No Effect                       | No                                   | No Effect                       | No                                   | Not Toxic              |
| OF 64        | No                                   | No Effect                       | No                                   | Minor Effect                    | No                                   | Not Toxic              |
| Prec +BR     | No                                   | No Effect                       | Yes                                  | Severe Effect                   | Yes                                  | Toxic                  |
| Prec + BR QA | No                                   | No Effect                       | Yes                                  | Severe Effect                   | Yes                                  | Toxic                  |
| Wastech      | No                                   | No Effect                       | No                                   | No Effect                       | No                                   | Not Toxic              |
| OF 62        | No                                   | No Effect                       | Yes                                  | Minor Effect                    | No                                   | Possibly Toxic         |
| OF 61A       | No                                   | No Effect                       | No                                   | No Effect                       | No                                   | Not Toxic              |
| OF 61        | No                                   | No Effect                       | No                                   | Minor Effect                    | No                                   | Not Toxic              |
| OF 61 QA     | No                                   | No Effect                       | No                                   | No Effect                       | No                                   | Not Toxic              |
| OF 60        | No                                   | No Effect                       | No                                   | No Effect                       | No                                   | Not Toxic              |
| Pac Meat     |                                      |                                 |                                      |                                 |                                      |                        |
| OF 1         | No                                   | No Effect                       | No                                   | No Effect                       | No                                   | Not Toxic              |
| OF 59        | No                                   | No Effect                       | No                                   | No Effect                       | No                                   | Not Toxic              |

N/A – Not Applicable – Control Sample

a- Statistical significance tested by comparing medians (Mann-Whitney U Test) and means using permutation tests.



**OUTFALL 59-65  
DEQ SEDIMENT STUDY**

Base imagery ca. 2005 from Oregon Imagery Explorer web mapping service.  
Outfalls from City of Portland Bureau of Environmental Services.



-  Toxic Bioassay OF59-65
-  Bioassay Locations
-  Targeted Samples
-  Lower Columbia Slough study area

G:\015\_Projects\Northwest\_Region\Columbia\_slough\MXD\050F59-65\_DEQ\_sed\_study\_rev 7/24/2009

**Figure 26 Bioassay Locations and Overall Interpretation**

## 5.0 BIOTA-SEDIMENT ACCUMULATION FACTORS

To obtain site-specific biota-sediment accumulation factors (BSAFs) DEQ used data from both two sediment sampling events:

- 1) The discrete sediment data with 23<sup>1</sup> sediment samples collected by the City in the Lower Slough in 2006 (BES 2009), and
- 2) The IS dataset collected by DEQ in 2009.

This sediment data was correlated to the City's 2005 fish tissue and fish lipid (GeoSyntec 2007) data set, for the calculations. BSAFs were calculated for several organic bioaccumulative contaminants that were analyzed in all three sampling events. The sediment and tissue data sets are reasonably spatially and temporally related. For the 2006 dataset BSAFs<sup>1</sup> were calculated using the following equation from DEQ's sediment bioaccumulation guidance (DEQ 2007).

$$BSAF = \frac{C_{fish} / F_L}{C_{sed} / F_{oc}}$$

Where:

- BSAF = Biota-Sediment Accumulation Factor (unitless)
- $C_{fish}$  = Chemical concentration in fish (ug/kg-wet weight)
- $C_{sed}$  = Chemical concentration in Sediment (ug/kg-dry weight)
- $F_L$  = Fish Lipid, as a fraction (e.g., kg/kg) wet weight.
- $F_{oc}$  = Sediment organic Carbon, as a fraction (e.g., kg/kg) dry weight.

The average fish lipid fraction from the City of Portland study is 0.087 and the average organic carbon fraction from the City of Portland data/IS samples is 0.023.

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<sup>1</sup> Three sediment duplicates removed for estimation of BSAFs. For fish lipid, 4 duplicate samples were averaged.

## 5.1 BSAFs from City 2006 Data

The fish in the 2005 fish tissue dataset were carp (family *Cyprinidae*) which are reported to have a home range of up to three miles stream length (Geosyntec, 2007) so individual fish may be exposed to a wide range of sediment contaminants. As shown in Table 4, with the exception of Aroclor 1254, median sediment values from the discrete samples tended to be most similar to the results of IS samples. For this reason, BSAFs were calculated assuming median values as a measure of central tendency for both sediment and fish tissue. Since Aroclor 1254 differed from other analytes, and the mean value appeared to be more representative, the mean was used in this case. The BSAF estimates along with comparison to DEQ and EPA published values are presented in Table 7.

**Table 7 BSAF estimates using 2006 discrete sediment samples <sup>a</sup>**

| Analyte      | BSAF Estimate <sup>b</sup> | DEQ 2007 <sup>c</sup> | U.S. EPA <sup>d</sup> |
|--------------|----------------------------|-----------------------|-----------------------|
| Aroclor 1254 | 1.8                        | 4                     | 1.4                   |
| Dieldrin     | 1.1                        | 24                    | 12.7                  |
| DDE          | 2.4                        | 28                    | 1.3                   |
| DDD          | 1.7                        | 24                    | 0.4                   |

a - All 23 lower slough sediment compared to 15 Lower Slough fish samples

b - Aroclor BSAF estimates based on mean concentration, others on median, based on results shown in Table 4.

c - Default estimates from DEQs 2007 bioaccumulation guidance.

d - U.S. EPA 2010: [http://www.epa.gov/med/Prods\\_Pubs/bsaf.htm](http://www.epa.gov/med/Prods_Pubs/bsaf.htm) (median values for Carp).

The results in Table 7 suggest somewhat less bioaccumulation (on average) than would be predicted using DEQ defaults. However, the DEQ defaults were selected based on 75<sup>th</sup> percentile values from WDOH 1995, whereas the calculated median values represent the 50<sup>th</sup> percentile, by definition. Moreover, the WDOH data likely include species other than carp, which may be at a higher trophic position in the food web. The DEQ defaults represent a higher bioaccumulation potential, and may be a reasonable representation of the “high end”.

When compared to USEPAs BSAF database, the value calculated for Aroclor 1254 is close to the value reported for total PCBs, while the dieldrin value is lower. The median EPA values for DDE and DDD are relatively close but lower than our calculated values. Differences may partially result from the fact that EPA fish samples for the pesticides were primarily fillets, while

samples analyzed for PCBs were primarily whole body, consistent with the DEQ study. The site-specific values generated in the study are assumed to best represent local conditions.

## 5.2: BSAFs from Linear Regression of 2009 IS Sediment and Fish Tissue Data

The IS sample results are believed to represent average sediment conditions. These results were used in a simple linear regression method<sup>2</sup> to derive a relationship between sediment and carp tissue for the common bioaccumulative compounds. Median concentrations of each analyte in sediment and carp tissue were used to create a single sediment/tissue data pair for the common analytes. The logs of the value were plotted to create a regression with a slope equal to the BSAF. Sediment organic carbon and analytical chemistry were taken from the IS sediment samples, and fish lipid and analytical chemistry were taken from carp samples. Sediment and tissue concentrations were normalized to fraction of organic carbon and lipid, respectively (as in the equation above). The results are shown in Figures 26 to 29. Figures 26 and 27 show all analytes combined, and PCBs and pesticides separately.

The combined sediment data of Figure 26 explain 70% of the variation in fish tissue concentrations, a very good result. Figure 27 suggests that the two classes of compounds (pesticides and PCBs) have different slopes and separate regressions were calculated as shown in Figures 28 and 29. For pesticides and PCBs, the regressions explain approximately 80 and 90% of the overall variance in the fish tissue concentrations respectively, an excellent result. The slope of these regression lines are 0.83 and 1.4 for PCBs and pesticides respectively. The regression slopes are equivalent to BSAFs, estimated using the linear regression method. The pesticide estimate of 1.4 compares to BSAF estimates in Table 7 of 1.1 to 2.4 indicating that the regression method using IS samples and the method using medians from discrete samples (Table 7) give comparable results. For Aroclor 1254, the regression method estimate of 0.83 is lower than the values estimated using discrete samples (1.8) from Table 7 but is generally consistent with theoretically expected values for benthic associated species (US EPA 2009 and 2010). The IS methodology appears to effectively reduce variation in sediment results allowing computation of reliable site specific bioaccumulation uptake factors.

There remain various uncertainties associated with this preliminary analysis. The spatial relation of the sediment and fish tissue data is uncertain, and PCB congener data were available for only a single fish. The effects of PCB weathering and differential accumulation by fish cannot be determined using simple Aroclor results by Method 8082. PCB mixtures undergo differential weathering in the environment, based on degree of chlorination and steric hindrances. Method 8082 uses pattern recognition of an un-weathered Aroclor standard that likely does not resemble the weathered environmental mixture. Moreover, many PCB congeners co-elute on

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<sup>2</sup> Logarithms used to linearize data and reduce the influence of extreme values.

chromatographic columns. Resolution of the complex compositional differences in environmental PCB mixtures would be required to assess detailed differences in accumulation, and these cannot be assessed with method 8082. It is anticipated that future work will further refine and validate these estimates.

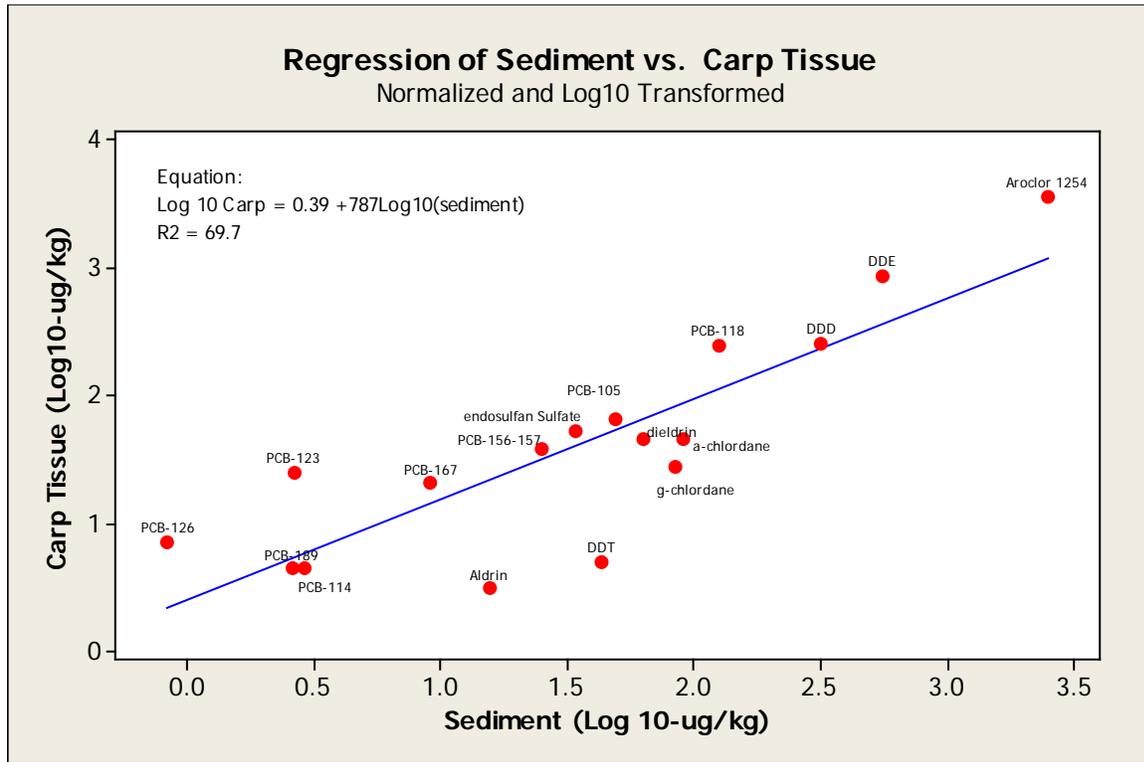


Figure 27 Normalized Regression of Sediment vs Carp Tissue

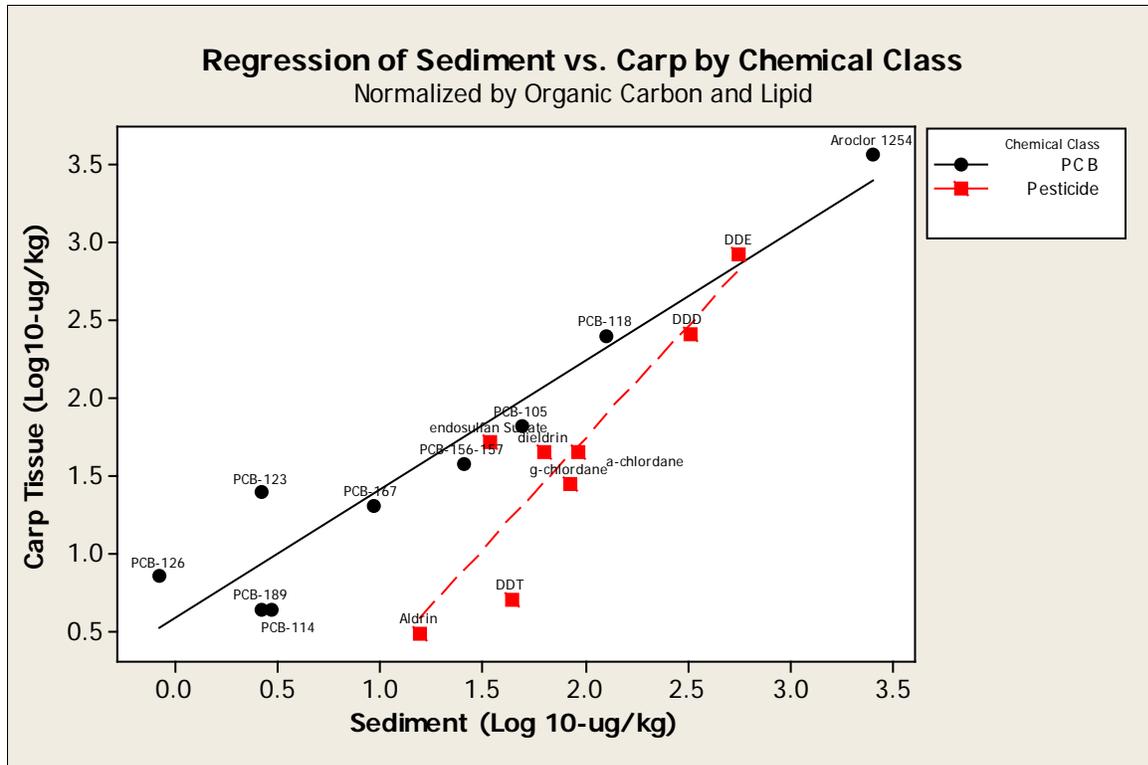


Figure 28 Chemical Class Regression of Sediment vs. Carp

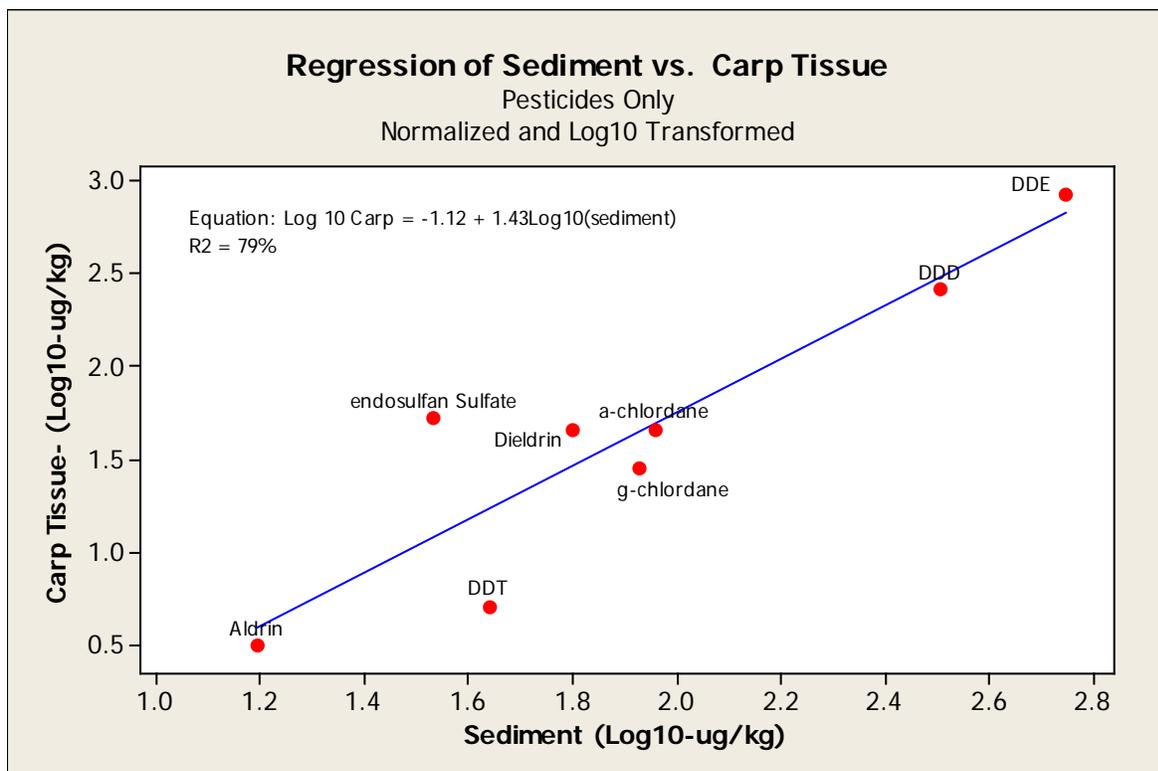
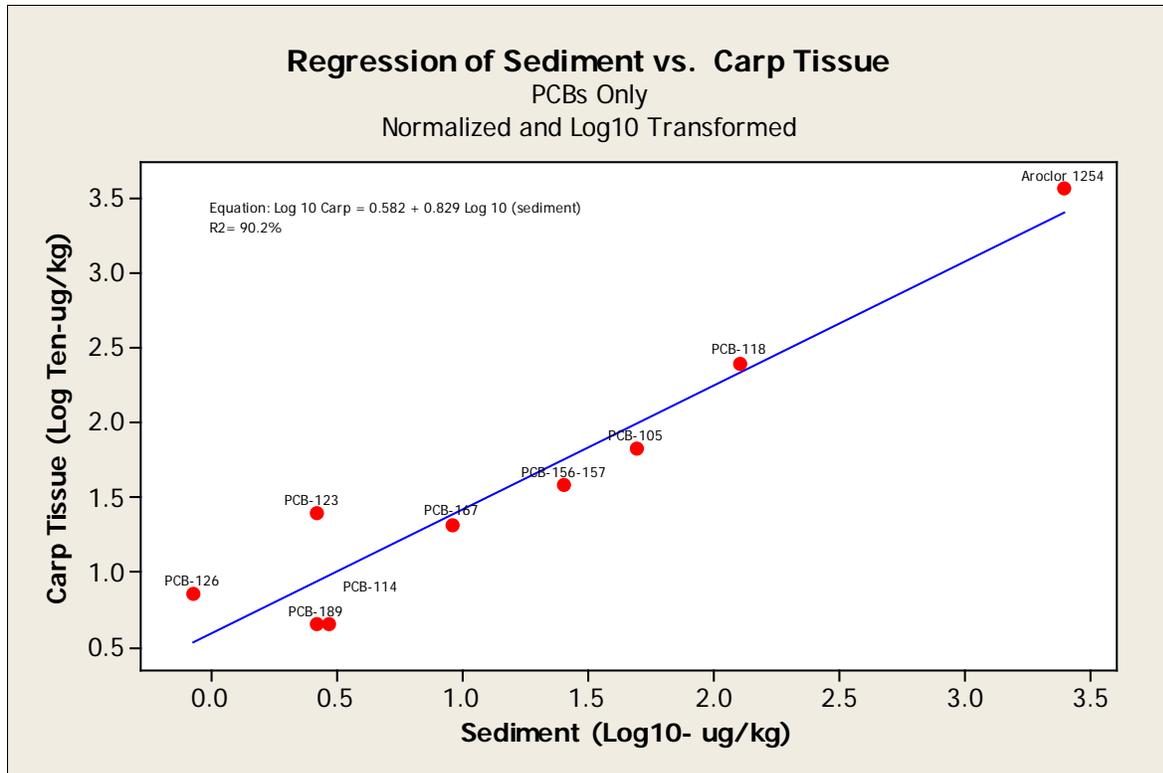


Figure 29 Regression of Sediment vs Carp Tissue: Pesticides



**Figure 30 Regression of Sediment vs. Carp Tissue: PCBs**

## 6.0 ADDITIONAL SEDIMENT DATA IN THE LOWER SLOUGH

Data collected by other parties in the Lower Slough since 2006 are summarized below to present available recent sediment information in the Slough. The US Army Corps of Engineers data was collected in 1998, but contains vertical profiling of the Lower Slough sediment and is thought to provide useful albeit somewhat dated information. Each project is summarized below followed by a table that describes contaminants above the IS levels. Sample locations are shown below in Figure 30.



**Figure 31 Additional Sediment Data in the Lower Slough**

## 6.1 Past Sediment Investigations within the 2009 Study Area

### 6.1.1 Pacific Meat

In October 2006 sediment samples were collected from a 600-foot reach of the Lower Slough offshore of the Pacific Meat (ECSI #145) cleanup site. Sediment samples were collected by the responsible party from 16 points in the vicinity of the property and two private outfalls. Sediment cores as deep as three feet below the mud line were collected and sediment down to 2 feet below the mud line were analyzed for PCBs, PAHs, cyanide, and metals. The highest total PCB results from the Pacific Meat site were 2,040  $\mu\text{g}/\text{kg}$  at the 0-1 foot depth and 2,450  $\mu\text{g}/\text{kg}$  at the 1-2 feet. In general, the deeper sediments from the 1-2-foot depth interval contained higher contaminant concentrations than shallower samples. Nearby outfalls include three private outfalls sampled in the DEQ 2009 study; including PM1, PM2 and PM 3.

### 6.1.2 Corps of Engineers

The Portland District US Army Corps of Engineers collected 15 sediment surface and core samples from depths as great as nine feet below the sediment surface from a portion of the Lower Slough in March 1998. The Corps selected sites based on their proximity to City outfalls but locations are generally too distant and limited in number to directly characterize outfall sediments. Multiple samples were collected from each core and composited to create an upper and lower profile sample. PCB (total) concentrations ranged up to 990  $\mu\text{g}/\text{kg}$  and were commonly between 300-700  $\mu\text{g}/\text{kg}$  in the upper part of the core samples. Concentrations in the

lower part of the core were considerably lower. DDx results for shallow core samples were commonly around 200 µg/kg, approximately an order of magnitude greater than the corresponding surface sample results. Shallow core samples contained lead at concentrations ranging from 110-360 mg/kg. Deeper core samples and surface samples for lead and other metals were comparable to City surface samples. Core samples contained elevated levels of mercury with concentrations ranging up to 4.5 mg/kg.

### 6.1.3 Oregon Department of Transportation(ODOT)

ODOT collected 15 sediment samples near Interstate 5 and Vancouver Bridge for bridge replacement projects. When detected, PCBs and pesticides, mainly DDx, were found in surface samples (0-5 ft) and not detected at lower depths (5-25 ft). PCBs were found in surface samples at up to 15 ug/kg and DDxs were found at concentrations between 15-150 ug/kg.

## **6.2 Past Sediment Investigations Outside the 2009 Study Area**

### 6.2.1 St Johns Landfill

Several sediment investigations have been performed around the perimeter of the St Johns Landfill in the main stem of the Lower Columbia Slough as well as side channels, the North Slough and the Blind Slough. For the “Stage 1” investigation City samples collected in the summer of 2006 were integrated into a fall 2006 investigation by METRO (owner of St Johns Landfill). Sample locations were spaced approximately 500-1000 feet along the Sloughs. METRO samples were collected from 0-0.5, 2-4, and 4-6 ft below mud line. City samples were collected only from the surface. Samples were analyzed for metals, herbicides, pesticides, PCBs, PAHs, phthalates, semi-volatile organic compounds, total organic carbon, and grainsize. The Stage 2 investigation concentrated on two areas where elevated sediment concentrations were found in Stage 1. Both the West Mud Flats and head of Blind Slough were suspected areas of elevated contamination due to their long history of leachate seep areas prior to placement of the landfill cover in the early 1990s. Results indicate moderate levels of sediment contaminants surround the landfill. Concentrations found at the West Mud Flats, head of Blind Slough, and near City Outfall 54 were generally elevated. Aroclor 1254 concentrations up to 64 ug/kg and total DDx concentrations up to 95 ug/kg were found at West Mud Flats. Aroclor 1254 concentrations up to 96 ug/kg were found at Blind Slough. The highest concentrations were often found in the 2-4 foot below mud line interval. Metals were somewhat elevated in surface sediment near Outfall 54.

### 6.2.2 Union Carbide (ECSI#176)

Union Carbide conducted a sediment investigation with DEQ oversight that included collection of sediment samples from upstream and downstream of the site and in a series of transects adjacent to the site. Samples were collected from the surface (0 – 6 inches) and subsurface (up to 3 feet below mud line). Samples were analyzed for PCBs, metals, SVOCs, and pesticides.

The highest PCB concentration found (86 ug/kg for Aroclor 1260) was near City Outfall 54. Aroclor 1242 was found at 77 ug/kg near the shore opposite the facility. The highest concentrations of metals were found upstream of the site, near Outfall 54 and there was no concentration gradient with depth. The highest concentrations of PAHs were found mid-channel with no apparent lateral concentration trends. Pesticides (primarily DDx) were elevated near Outfall 54 and mid-channel near the facility.

### **6.2.3 Pacific Carbide(ECSI# 268)**

Bioassay tests using sediments near the Pacific Carbide site found mortality to benthic organisms in 2006. Slough banks adjacent to site are composed of lime material which contains elevated PAHs, metals, and some PCBs c from 0-2 feet deep. Total PAHs were found in sediment at up to 770,000 ug/kg and total PCB Aroclors were found in sediment at up to 1,100 ug/kg.

## **6.3 City of Portland 2006 Investigation**

As discussed in Section 5, the City of Portland (COP) collected 23 surface sediment samples from the Lower Slough in 2006 as part of their long-term monitoring of Slough-wide conditions. Samples were analyzed for PCBs, semi-volatile organic compounds including PAHs, metals, pesticides, petroleum hydrocarbons, grain size, and organic carbon. Total PCBs, detected at concentrations ranging from 60-330 µg/kg are generally higher in this reach of the Slough than other Slough reaches and are relatively consistent in concentration.

COP 2006 metals results include: copper from 56-100 mg/kg, lead from 56-139 mg/kg, nickel from 27-36 mg/kg, zinc from 304-576 mg/kg. In general these metals are found at higher concentrations in the Lower Slough than other parts of the Slough. COP samples contained total DDT, DDE and DDD (DDx) at concentration of 12-62 µg/kg. Concentrations of total PAH compounds and oil are relatively higher in this reach compared with other parts of the Slough system.

## **7.0 CONCLUSIONS**

### **7.1 Baseline concentrations in Lower Slough**

This study supports the use of the 2009 incremental sample results as the new “baseline” or “ambient” concentration for the lower slough. While it only represents a portion of the Lower Slough, we believe it is generally representative of the entire Lower Slough based on a comparison to the 2006 data, knowledge of site sources, and the connectivity between the 2009 segment and the remainder of the slough. The IS concentrations do not vary significantly from the baseline values derived by DEQ in 2002 as can be seen for pertinent compounds in Table 4. The new baseline levels for metals in the Lower Slough are consistently lower than the previously established values which will continue to be used for Middle and Upper Columbia Slough evaluation. This is either due to reduced sources contributing to contamination over time or higher metals concentrations elsewhere in the Slough that raised the baseline value. Pesticide

screening level concentrations are consistently lower than previous values. This is due in part to lower detection limits achieved using analytical method U.S. EPA 1699 and in part to the fact that there are no known pesticide sources in the lower slough. Baseline screening levels for polycyclic aromatic hydrocarbon (PAHs) are consistently higher in this section of the Slough than the Slough-wide levels. This may reflect the more industrial nature of this portion of the Slough and the higher percentage of roadway surfaces, and runoff from the I-5 Freeway. A Lower Slough-specific screening level table, Table 8, incorporating the new baseline values and risk-based evaluation discussed below is provided at the end of this section.

## 7.2 Benthic Toxicity in the Lower Slough

Data is being re-evaluated as part of the broader Columbia Slough study.

## 7.3 Bioaccumulation

Data is being re-evaluated as part of the broader Columbia Slough study.

## 7.4 Elevated Concentration Zones

Several locations in the study were identified as having significantly elevated concentrations of contaminants in sediment. Those values that exceed the baseline value or one order of magnitude above a risk-based screening level, whichever is higher, are highlighted in this section along with some preliminary information on potential contaminant sources. Table 10 identifies sample locations with the contaminant concentrations meeting these criteria highlighted. Locations are shown in Figure 33.

**Table 8 Elevated Concentration Zones**

| Site                          | Significantly Elevated Concentrations |        |            |      |     |       | Bioassay |
|-------------------------------|---------------------------------------|--------|------------|------|-----|-------|----------|
|                               | PCBs                                  | Metals | Pesticides | PAHs | TBT | SVOCs | Toxicity |
| Outfall 59                    | -                                     |        | Yes        | -    | -   |       | -        |
| Pacific Meats & Outfall 60    | Yes                                   | -      | Yes        | -    | -   | -     | -        |
| Outfall 61A                   | -                                     | -      | -          | -    | Yes | -     | -        |
| Outfall 62 & Precision/B&R    | Yes                                   | -      | Yes        | -    | -   | -     | Yes      |
| Outfall 63 & Vancouver Bridge | Yes                                   | -      | Yes        | -    | -   | -     | -        |
| Outfall 65                    | -                                     | -      | -          | -    | -   | -     | Yes      |

Starting at the downstream end of the study area, the first location is adjacent to City of Portland Outfall 59. Chromium was detected at 481 mg/kg at this location which is more than ten times the baseline (44.1 ppm) and copper and nickel were elevated as well; however the bioassay

conducted on this sample did not indicate toxicity. During sampling the field crew observed what appeared to be roofing sand in the sediment near the outfall. Roofing sand often contains metals useful for moss control on roofs. DDT was detected at 55.6 ug/kg. No known source of pesticides is known in this section, but historic use of DDT in the area is possible. The inactive DEQ Cleanup site known as Malarkey Roofing (ECSI No. 690) is located nearby and discharges to this outfall.



**Figure 32 Roofing sands at Outfall 59**

The second location identified as elevated is centered on Pacific Meats outfalls No. 1 and No. 2 and City of Portland Outfall 60. PCBs were detected at up to 646 ug/kg total Aroclors at the Pacific Meat outfalls. Just upstream from these private outfalls, samples collected adjacent to City Outfall 60 contained higher concentrations, up to 1,176 ug/kg PCB Aroclors, the highest surface sediment concentration detected in the study area. It is not clear if there is any connection between Outfall 60 and the Pacific Meats site. Another possible source for PCBs at Outfall 60 is an inactive DEQ cleanup site known as Goodyear Distribution Center #9000 (ECSI No. 2105). In 1995, DEQ received a report of a spill of transformer oil at the Goodyear Building. It was reported that vandals drained and gutted six transformers releasing 100 gallons of transformer oil with PCBs concentrations exceeding 7 ppm. Stormwater from this site appears to drain to Outfall 60. NW Cast (ECSI #999) is also within the Outfall 60 drainage area and formerly housed a transformer oil and metal smelting shop. On-site soil sampling in 2001

found total PCB's in 8 of 9 soil samples ranging from 500 ppb to 27,500 ppb. The NW Cast site owners are conducting a stormwater Source Control Evaluation under DEQ oversight.

Pacific Meats is an active DEQ Cleanup site. A former operator at the property reclaimed PCB oil-containing transformers and caused considerable upland contamination. A soil cap is being placed on upland areas of contamination. There was evidence of overland flow of waste from site settling ponds as well as discharge of contaminated sediments through outfalls. Sediment sampling at discrete points was performed in 2006 by Pacific Meats at depths of 0.5 and 2.0 feet below the mudline. The deeper samples showed the highest concentrations with total Aroclors detected at up to 2450 ug/kg, consistent with a historic release. Deeper samples also contained elevated DDT and dieldrin values, while shallow samples did not, also consistent with historic releases.

The third area of elevated concentrations is ODOT Outfall 61A where tributyltin was found at 285 ug/kg. This is two orders of magnitude above the risk value for tributyltin of 2.3 ug/kg (EPA 2005). The source of tributyltin at this outfall is unknown. Tributyltin was also found at Outfall 61 and Outfall 63. No tributyltin sources have been identified for Outfall 63. Although bioassays in this area did not indicate toxicity, Outfall 61A has relatively high levels of metals (particularly copper, lead, and zinc) and PAHs in sediment, which may be a reflection of the metal foundries Macadam Aluminum and Bronze (ECSI No. 2765) and CARCO/Varicast (ECSI No. 3389) that discharge to this outfall. Stormwater from 84 acres of I-5 freeway and Columbia Blvd. also drains to Outfall 61A which could be a source of metals and PAHs.

Further upstream is the fourth area of elevated concentrations where Outfall 62 discharges and adjacent to the shoreline extending upstream passed the Wastech (ECSI No. 1271), Precision Equipment (ECSI No. 152), and B&R Auto wrecking sites (ECSI No. 4149). Outfall 62 sediments had the second highest total PCB Aroclor concentration in surface sediments and had the highest concentration of PCB congeners. Specific PCB congener concentrations detected at all these sites appear to match up reasonably well, suggesting a common source. This area was historically low lying and received a considerable amount of imported fill beginning in approximately 1925. The area contained a construction debris landfill, which was active from 1971 to 1973. A used oil recycling facility operated in the area from 1943 to 1974 and apparently discharged waste liquids to an open pond and ditch that emptied into the Slough. Dieldrin and delta-BHC pesticides were also elevated in this area. Bioassay samples at Precision/B&R indicated toxicity and the sample from Wastech indicated possible toxicity.



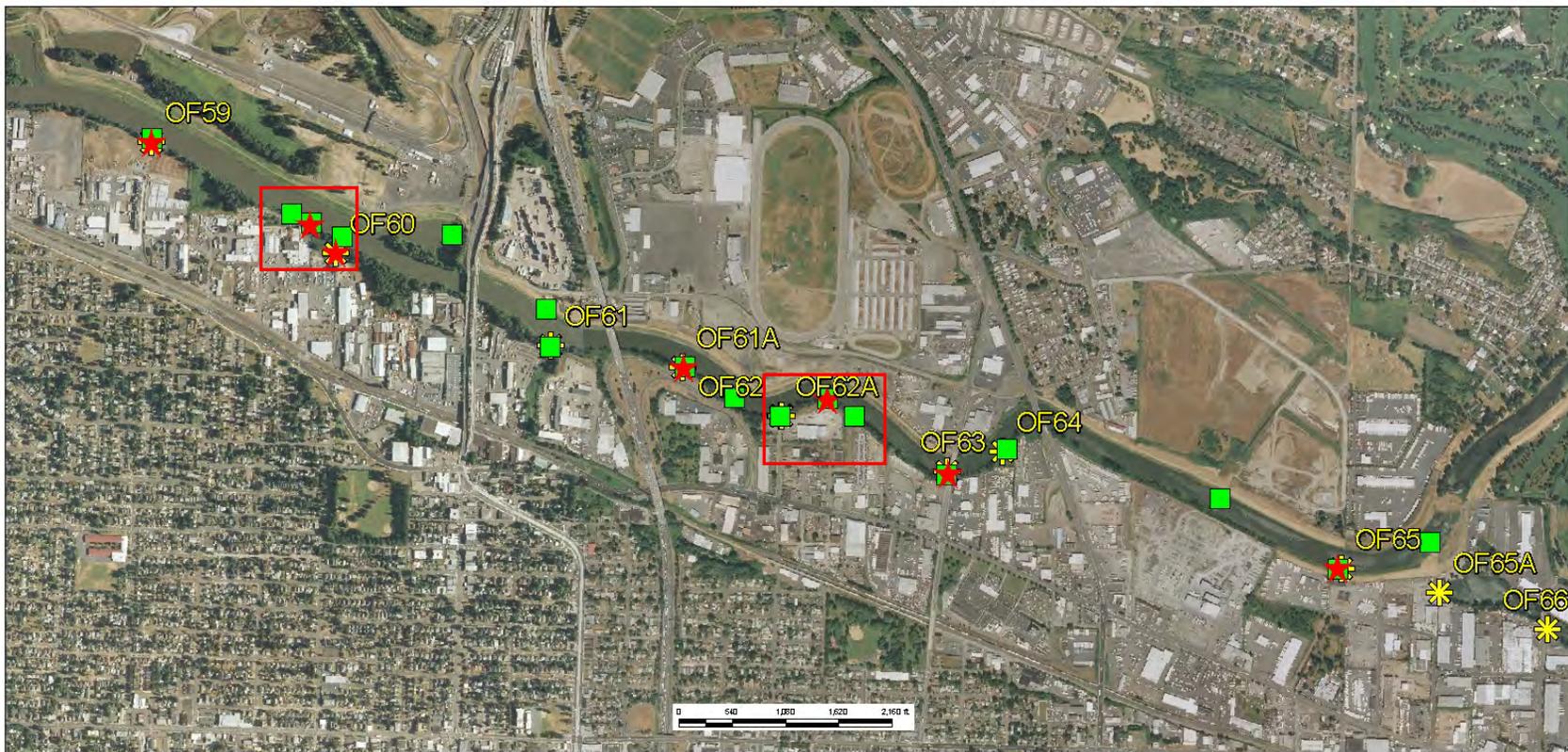
**Figure 33 Wastech bank sample area**

The fifth area of elevated concentrations is adjacent to Outfall 63 with heptachlor concentrations at 10.2ug/kg , just above the elevated definition criteria. It is likely that this value reflects historic releases. The total Aroclors concentration here was 91.4 ug/kg and samples collected at the nearby Vancouver Bridge construction site had PCBs up to 128 ug/kg, both exceeding the elevated criteria for PCBs. There are no known sources of PCB's in this area. Private outfalls in the vicinity of the sample area, including Ott's Friction (ECSI #5444) and Woodfeathers may also contribute to concentrations in the slough. Ott's is a distributor of truck parts and Woodfeathers is a distributor of roofing materials. Historically, Ott's housed a meat packing plant from 1920's to about 1980. Additionally, the south end of Vancouver Bridge, near the composite area, suffered a major fire to its wooden pilings in 2008.

Finally, the bioassay sample from Outfall 65 indicated toxicity. Chemistry results from the area indicated total PCB Aroclors at 96.3 ug/kg. Concentrations for other contaminants are not particularly high and other bioassays conducted on other samples with higher concentrations did not indicate toxicity, so the reason for toxicity at this outfall is not completely clear. The

Columbia Land Reclamation or Land Reclamation landfill was located on the nearby Hanson Pipe Cleanup site (ECSI #3893). Records indicate that the landfill operated from around 1974 until 1979. A City 2006 sample just downstream of this location contained total PCB Aroclors at 329.5 ug/kg. We assume that these elevated concentrations are connected, and that possible conditions associated with historic landfill discharges may contribute to toxicity.

**Figure 34 : All Elevated Concentration Zones between Outfall 59-65**



**OUTFALL 59-65  
DEQ SEDIMENT STUDY**

Elevated Concentration  
Zones with more than  
one targeted sample



- ★ Elevated Concentration Zones
- 2009 Targeted Samples
- ✱ City of Portland outfalls

Base imagery ca. 2005 from Oregon Imagery Explorer web mapping service.  
Outfalls from City of Portland Bureau of Environmental Services.

## **7.5 Natural recovery**

At the Pacific Meats Outfalls #1 and #2 sediment contaminants were compared in the intervals from 0-5 cm and 0-10 cms. PCBs, pesticides, and metals concentrations were all consistently higher in the deeper interval while PAHs were slightly higher in the shallow interval. Results suggest that legacy contaminants released historically are being covered up by cleaner sediment. Source control or the curtailed used of contaminants like PCBs may be effectively reducing new accumulation in recently deposited sediments. PAH contamination, on the other hand, appears to be slightly higher in recently deposited sediments, suggesting that sources for PAHs may not be controlled effectively. This may reflect increased traffic and associated vehicular releases that occur on roadways.

## **8.0 RECOMMENDATIONS/FUTURE STEPS**

### **8.1 Sediment Cleanup Prioritization**

Bioaccumulative compounds were determined to contribute the most toward risk for ecologic and human receptors and correspondingly warrant the highest priority for cleanup. Because exposure associated with bioaccumulative contaminants occurs throughout the area fish travel within the slough, it is appropriate to assess exposure based on average concentrations throughout this area. In order to identify priority areas for remedial action of bioaccumulative contaminants, weighted average concentrations of contaminants of concern were calculated. As discussed in Section 7.3, bioaccumulative contaminants of greatest concern for the Lower Slough are PCBs and dieldrin.

The IS average for these compounds exceeds the risk-based concentrations using the BSAF developed for the Lower Slough as documented in Section 5.2. The available discrete sediment data for the Slough segment was used to prioritize areas to target for sediment cleanup efforts. The process consisted of the following steps:

1. Theissen polygons were generated for all available discrete data points in the Slough segment. This included the following data sets: DEQ 2009 data, ODOT Vancouver bridge construction data, Pacific Meat sediment data, and City of Portland 2006 sediment data. Because the discrete data was primarily focused on contaminant source areas, it is expected to generate a higher area weighted average than the IS samples.
2. Five-hundred-foot buffers were drawn around each data point to limit the size of the polygons generated for points with limited surrounding data. This left some “blank” areas or polygons with no associated discrete data points.
3. Areas for each polygon were calculated and the area of non-represented sediment (zones between polygons) was calculated.

4. Area-weighted average concentrations for PCBs and dieldren were calculated by assuming the concentration at the data point reflected the concentration of the polygon and applying the IS average concentration to the zones between polygons.

This process generated a segment-wide weighted average concentration for PCBs of 92.65 ug/kg and for dieldren of 1.82 ug/kg. This compares to the IS averages of 45.5 and 0.7 ug/kg respectively. As a guideline for identifying priority areas, the highest concentration polygons were successively reduced to the method detection limit (MDL) (assumed to be 10 ug/kg for PCBs and 1 ug/kg for dieldren) for these contaminants until the overall weighted average fell below the IS average.

For PCBs, five priority areas were identified as documented in Table 11 below and illustrated in Figure 34.

**Table 9 PCB priority areas for cleanup**

| <b>Sample location</b>                         | <b>PCB concentrations</b> |
|--|---------------------------|
| Pacific Meat – 9 locations                     | 123 - 2040 ug/kg          |
| City of Portland OF 62–Wastech, Precision, B&R | 904 ug/kg                 |
| OF 63 and Vancouver St. bridge                 | 128 ug/kg                 |
| City of Portland OF 61                         | 138 ug/kg                 |
| Downstream of OF 65                            | 329.54 ug/kg              |

Clearly the Pacific Meat shoreline area has the highest priority for cleanup and is the best characterized area. Feasible cleanup measures can likely be developed with minimal additional data collection. Based on our limited study of the vertical variation of contaminant levels indicated in the Pacific Meat area, it appears likely that contamination is associated with historical as opposed to on-going releases.

Each of the other areas may require additional sediment sampling to define the lateral and vertical extent of significant sediment contamination and evaluating feasible cleanup approaches.

The second highest priority area is the OF 62/Wastech/Precision B&R reach. The likely sources for contamination in this area are an historic oil reclamation facility and contaminated fill along the bank of the Slough.



Figure 35 PCB Priority Cleanup Areas- shaded in light purple.

The dieldren evaluation was somewhat problematic due to the large number of samples in which dieldren was not detected and the elevated method detection limits. Several of the priority areas identified for PCBs were also identified as a priority for dieldren. DEQ recommends that cleanup evaluation focus on the priority areas identified for PCBs, and dieldren concentrations be re-evaluated once those actions have been completed. This focus is supported by the fact that PCB concentrations detected in sediment and fish tissue in this portion of the slough are more significantly elevated relative to risk-based levels than the dieldren concentrations.

In addition to the bioaccumulation evaluation, areas where benthic toxicity may be an issue were also evaluated. The OF62, Wastech, Precision and B&R area was the area with the clearest signal of benthic toxicity and this area has already been identified as the second highest priority for further investigation and cleanup. The only other location where toxicity was identified was OF 65. This is also a priority location but other than a former construction debris landfill there is little information on the likely sources.

### **8.3 Remedial Options**

The US Army Corps of Engineers and the Multnomah County Drainage District have long wanted to do extensive dredging in the Lower Slough for flood control purposes but have been reluctant to do this work due to the significant cost for disposal of contaminated sediments. A safe and local repository for contaminated sediments would be necessary. Dredging of contaminated sediment is costly and carries some risk of stirring up and spreading residual contamination. Thin layer capping may be an option however it would discourage any future dredging operations whether for flood control or mass removal. Monitoring natural attenuation may be an option as it appears that sources are primarily historic in nature and they are being buried by recent, cleaner sediment.

DEQ plans to initiate discussions with local, state and federal officials and other public and private advocacy groups on cleanup approaches and strategies for the lower Slough with a goal of reaching consensus on a preferred strategy. This information will be essential to determine costs and funding mechanism to complete the cleanup.

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Appendix A: Maps of Sample Locations

Figure A-3 Detailed Columbia Slough Sediment Study Map 2-11.

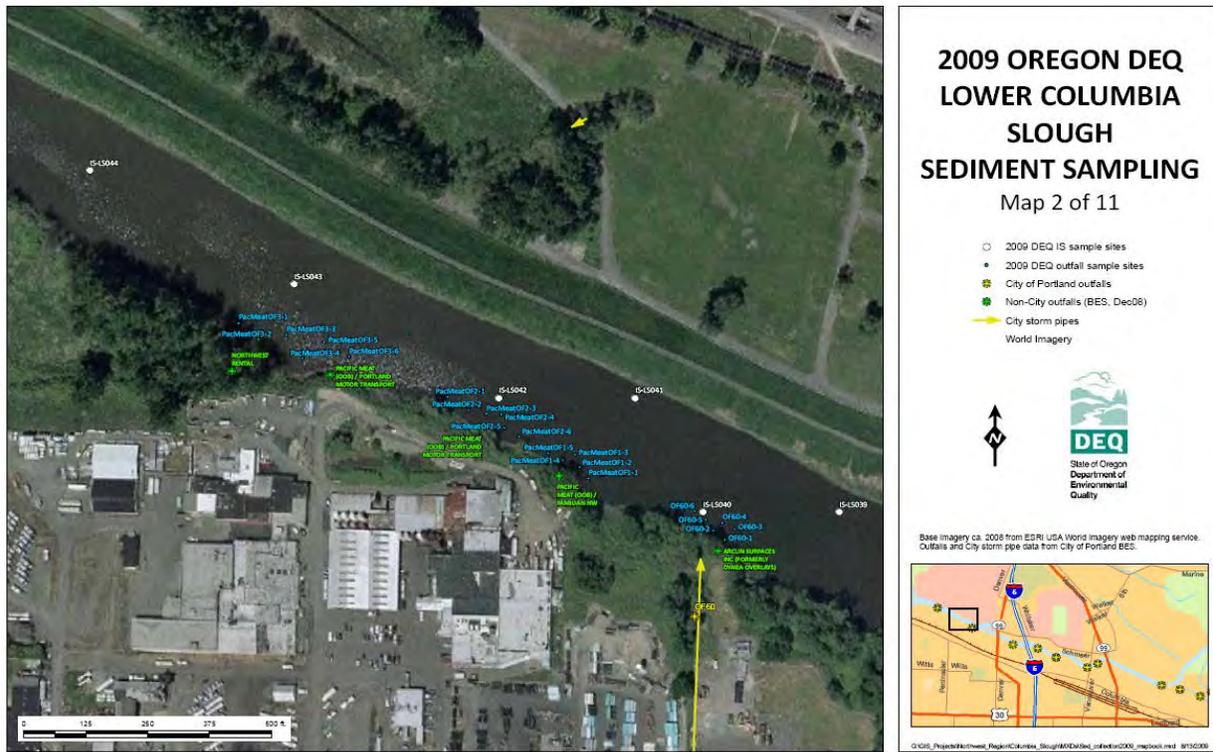
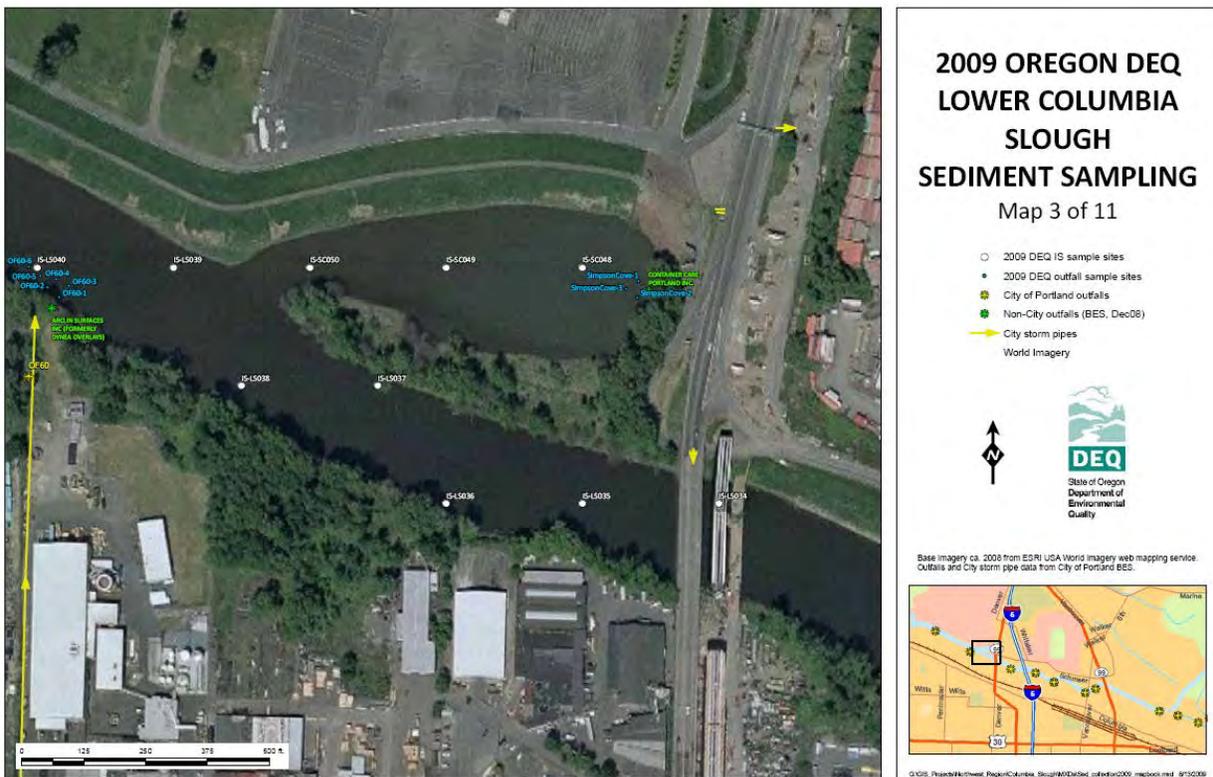


Figure A-4 Detailed Columbia Slough Sediment Study Map 3-11.



Appendix A: Maps of Sample Locations

Figure A-5 Detailed Columbia Slough Sediment Study Map 4-11.

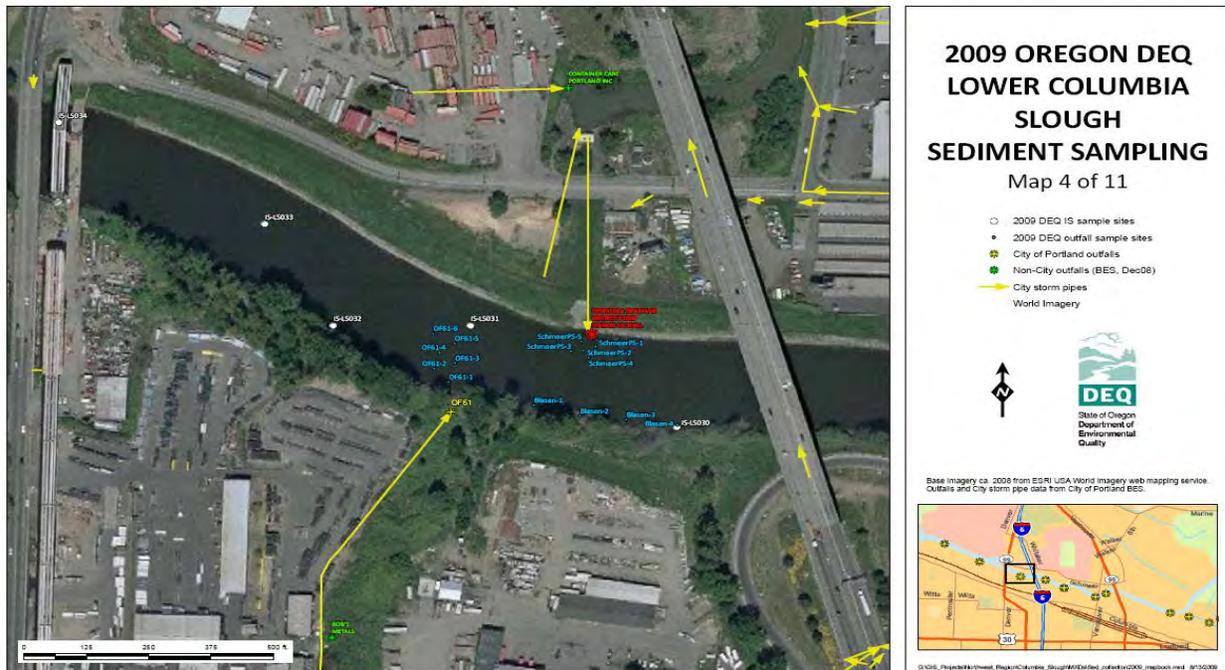
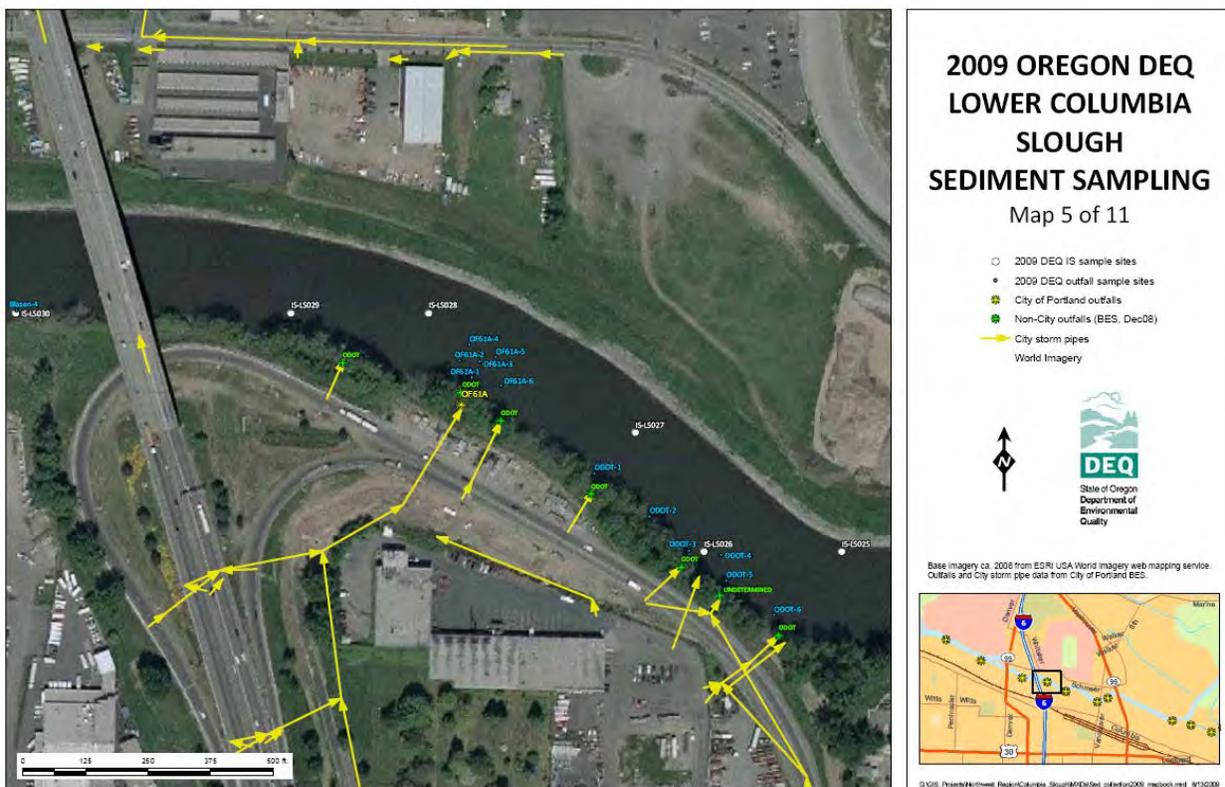


Figure A-6 Detailed Columbia Slough Sediment Study Map 5-11.



Appendix A: Maps of Sample Locations

Figure A-7 Detailed Columbia Slough Sediment Study Map 6-11.

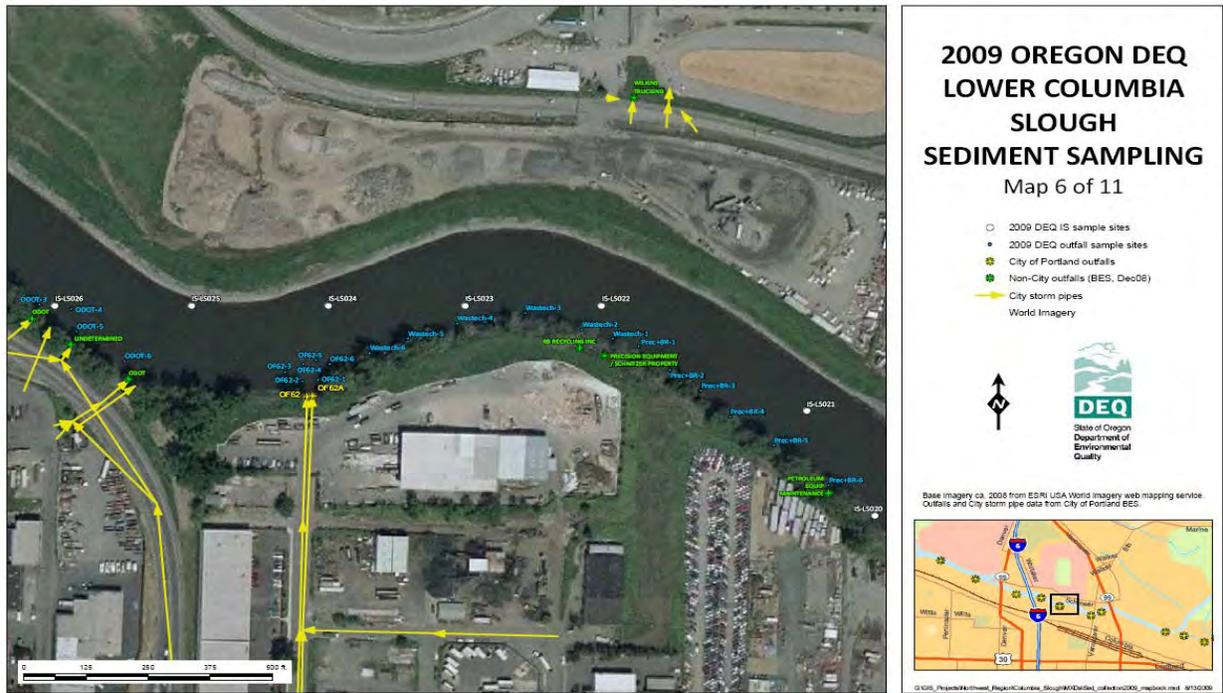


Figure A-8 Detailed Columbia Slough Sediment Study Map 7-11.



Appendix A: Maps of Sample Locations

Figure A-9 Detailed Columbia Slough Sediment Study Map 8-11.

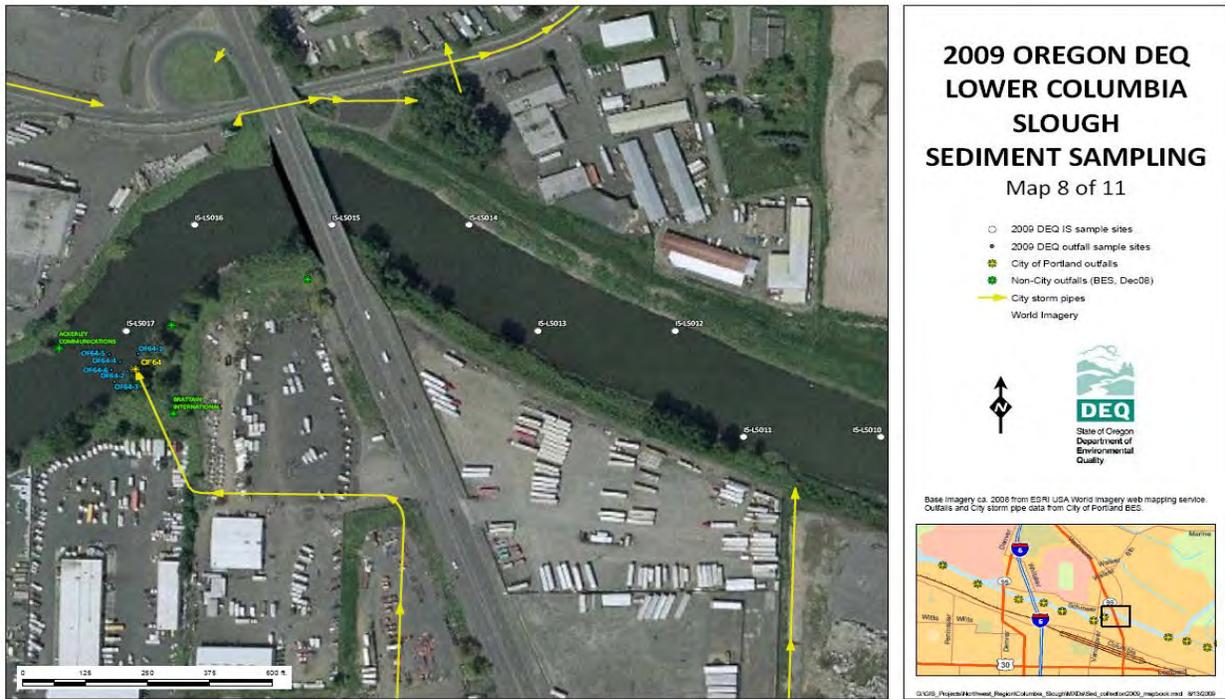
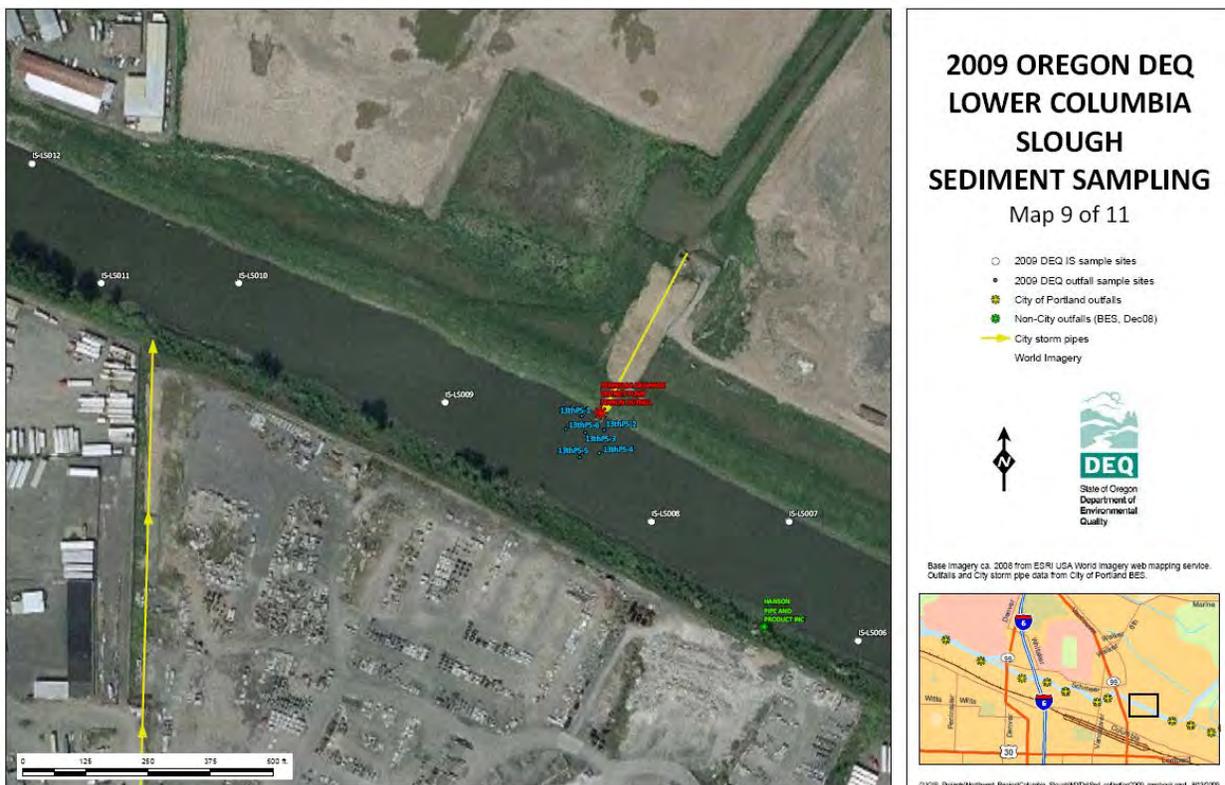


Figure A-10 Detailed Columbia Slough Sediment Study Map 9-11.



Appendix A: Maps of Sample Locations

Figure A-11 Detailed Columbia Slough Sediment Study Map 10-11.

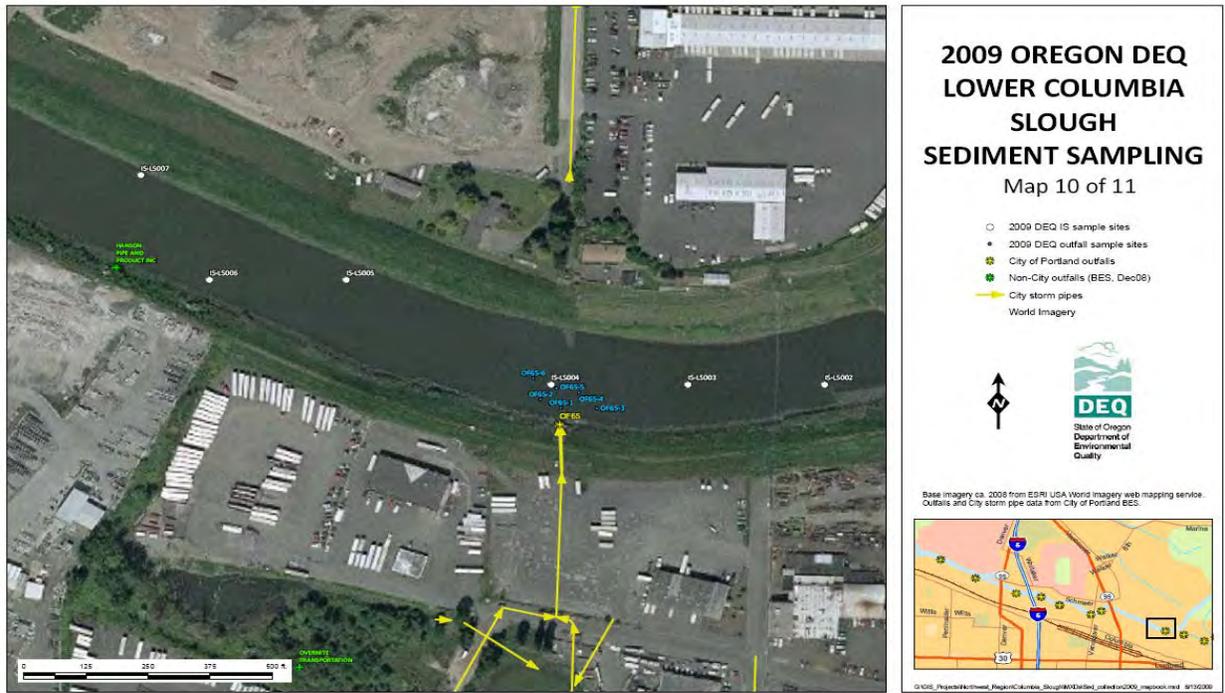
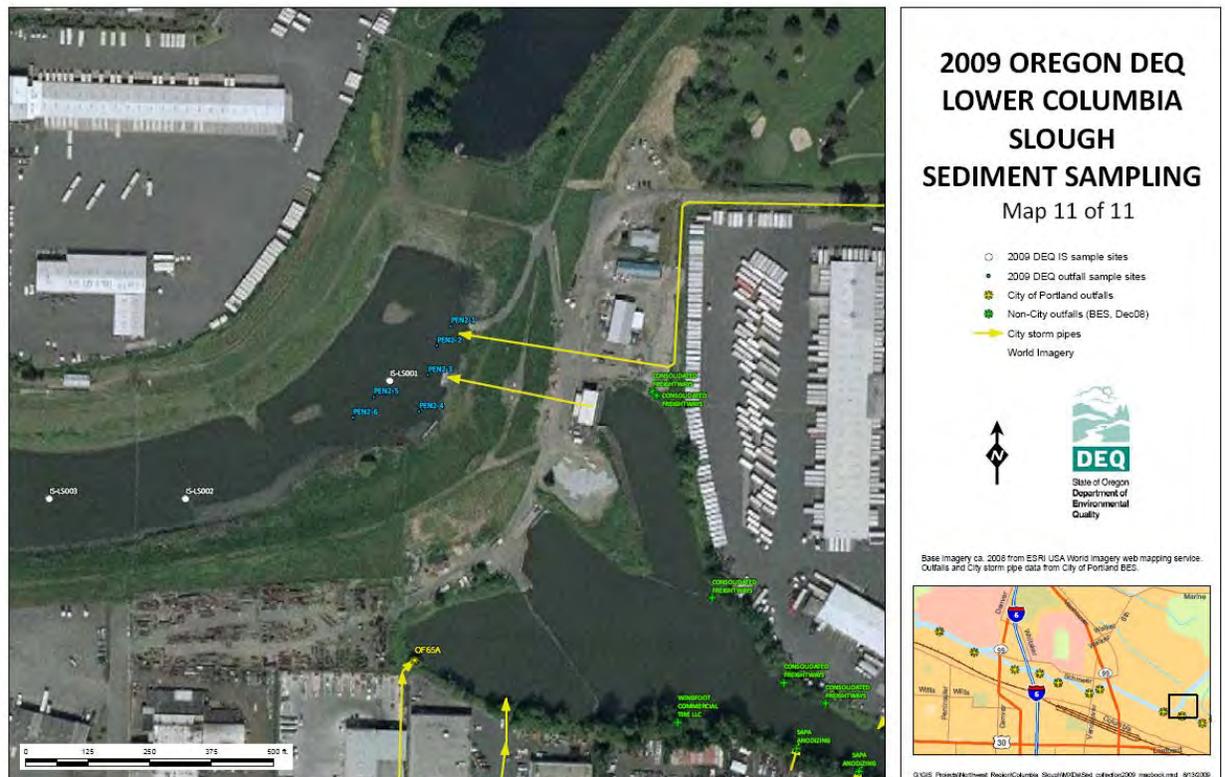


Figure A-12 Detailed Columbia Slough Sediment Study Map 11-11.



Appendix A: Maps of Sample Locations



**FIGURE 2**  
**OUTFALL 59-65**  
**DEQ SEDIMENT STUDY**

- Lower Columbia Slough study area
- City of Portland outfalls
- Non-City Outfalls (BES, De d08)
- Peninsula Drainage District Outfall



Base imagery ca. 2005 from Oregon Imagery Explorer web mapping service.  
Outfalls from City of Portland Bureau of Environmental Services.

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| <b>Appendix B: Table of Sample Locations</b> |  |               |               |                |                          |                           |           |                                   |                     |       |
|--|--|---------------|---------------|----------------|--------------------------|---------------------------|-----------|-----------------------------------|---------------------|-------|
| Lasar #                                      | Site Name  | Site Location | Latitude      | Longitude      | latitude(dd)<br>Centroid | Longitude(dd)<br>Centroid | depth(cm) | Number of Scoops<br>50mg= 1 scoop | Volume<br>requested |       |
| 35962  | <b>Columbia Slough at City of Portland Outfall 65</b>      |               |               |                | 45.58259                 | -122.65192                |           |                                   |                     | 3400g |
|  |  | OF65-1        | 45° 34.952' N | 122° 39.116' W |                          |                           | 7         | 12                                |                     |       |
|  |  | OF65-3        | 45° 34.952' N | 122° 39.099' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF65-2        | 45° 34.955' N | 122° 39.126' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF65-4        | 45° 34.957' N | 122° 39.107' W |                          |                           | 7         | 12                                |                     |       |
|  |  | OF65-5        | 45° 34.959' N | 122° 39.118' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF65-6        | 45° 34.962' N | 122° 39.129' W |                          |                           | 7         | 12                                |                     |       |
| 35963  | <b>Columbia Slough at PEN 2 Levee</b>                      |               |               |                | 45.58338                 | -122.64805                |           |                                   |                     | 425g  |
|  |  | PEN2-1        | 45° 35.022' N | 122° 38.869' W |                          |                           | 5         | 2                                 |                     |       |
|  |  | PEN2-2        | 45° 35.015' N | 122° 38.875' W |                          |                           | 5         | 2                                 |                     |       |
|  |  | PEN2-3        | 45° 35.006' N | 122° 38.914' W |                          |                           | 5         | 2                                 |                     |       |
|  |  | PEN2-4        | 45° 34.994' N | 122° 38.888' W |                          |                           | 5         | 2                                 |                     |       |
|  |  | PEN2-5        | 45° 35.003' N | 122° 38.919' W |                          |                           | 7         | 2                                 |                     |       |
|  |  | PEN2-6        | 45° 34.998' N | 122° 38.922' W |                          |                           | 10        | 2                                 |                     |       |
| 35964  | <b>Columbia Slough at 13<sup>th</sup> Ave. pumpstation</b> |               |               |                | 45.58435                 | -122.65688                |           |                                   |                     | 425g  |
|  |  | 13thPS-1      | 45° 35.068' N | 122° 39.415' W |                          |                           | 5         | 2                                 |                     |       |
|  |  | 13thPS-2      | 45° 35.063' N | 122° 39.404' W |                          |                           | 10        | 2                                 |                     |       |
|  |  | 13thPS-3      | 45° 35.062' N | 122° 39.413' W |                          |                           | 7         | 2                                 |                     |       |
|  |  | 13thPS-4      | 45° 35.056' N | 122° 39.407' W |                          |                           | 10        | 2                                 |                     |       |
|  |  | 13thPS-5      | 45° 35.054' N | 122° 39.416' W |                          |                           | 10        | 2                                 |                     |       |
|  |  | 13thPS-6      | 45° 35.063' N | 122° 39.423' W |                          |                           | 5         | 2                                 |                     |       |
| 35965  | <b>Columbia Slough at City of Portland Outfall 63</b>      |               |               |                | 45.58485                 | -122.66749                |           |                                   |                     | 3400g |
|  |  | OF63-1        | 45° 35.093' N | 122° 40.068' W |                          |                           | 5         | 10                                |                     |       |
|  |  | OF63-2        | 45° 35.100' N | 122° 40.073' W |                          |                           | 10        | 10                                |                     |       |
|  |  | OF63-3        | 45° 35.099' N | 122° 40.061' W |                          |                           | 7         | 10                                |                     |       |
|  |  | OF63-4        | 45° 35.096' N | 122° 40.048' W |                          |                           | 10        | 10                                |                     |       |
|  |  | OF63-5        | 45° 35.099' N | 122° 40.034' W |                          |                           | 7         | 10                                |                     |       |
|  |  | OF63-6        | 45° 35.096' N | 122° 40.033' W |                          |                           | 10        | 10                                |                     |       |
|  |  | OF63-7        | 45° 35.095' N | 122° 40.020' W |                          |                           | 10        | 10                                |                     |       |
| 35966  | <b>Columbia Slough at City of Portland Outfall 64</b>      |               |               |                | 45.58548                 | -122.66557                |           |                                   |                     | 3400g |
|  |  | OF64-1        | 45° 35.132' N | 122° 39.928' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF64-2        | 45° 35.126' N | 122° 39.932' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF64-3        | 45° 35.122' N | 122° 39.938' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF64-4        | 45° 35.129' N | 122° 39.936' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF64-5        | 45° 35.132' N | 122° 39.941' W |                          |                           | 10        | 12                                |                     |       |
|  |  | OF64-6        | 45° 35.126' N | 122° 39.940' W |                          |                           | 10        | 12                                |                     |       |
| 35967  | <b>Columbia Slough at Prec+BR</b>                          |               |               |                | 45.58627                 | -122.67116                |           |                                   |                     | 3400g |
|  |  | Prec+BR-1     | 45° 35.197' N | 122° 40.305' W |                          |                           | 10        | 12                                |                     |       |
|  |  | Prec+BR-2     | 45° 35.188' N | 122° 40.290' W |                          |                           | 10        | 12                                |                     |       |
|  |  | Prec+BR-3     | 45° 35.181' N | 122° 40.284' W |                          |                           | 10        | 12                                |                     |       |
|  |  | Prec+BR-4     | 45° 35.171' N | 122° 40.261' W |                          |                           | 10        | 12                                |                     |       |
|  |  | Prec+BR-5     | 45° 35.161' N | 122° 40.240' W |                          |                           | 10        | 12                                |                     |       |
|  |  | Prec+BR-6     | 45° 35.147' N | 122° 40.213' W |                          |                           | 10        | 12                                |                     |       |

| Lasar # | Site Name  | Site Location | Latitude      | Longitude      | latitude(dd) | Longitude(dd) | depth(cm) | Number of Scoops | Volume    |  |
|---------|--|---------------|---------------|----------------|--------------|---------------|-----------|------------------|-----------|--|
|         |  |               |               |                | Centroid     | Centroid      |           | 50mg= 1 scoop    | requested |  |
| 35968   | <b>Columbia Slough at Wastech</b>                      |               |               |                | 45.58684     | -122.67288    |           |                  | 3400g     |  |
|         |  | Wastech-1     | 45° 35.199' N | 122° 40.317' W |              |               | 10        | 12               |           |  |
|         |  | Wastech-2     | 45° 35.203' N | 122° 40.332' W |              |               | 10        | 12               |           |  |
|         |  | Wastech-3     | 45° 35.208' N | 122° 40.359' W |              |               | 7         | 12               |           |  |
|         |  | Wastech-4     | 45° 35.204' N | 122° 40.391' W |              |               | 10        | 12               |           |  |
|         |  | Wastech-5     | 45° 35.198' N | 122° 40.414' W |              |               | 10        | 12               |           |  |
|         |  | Wastech-6     | 45° 35.192' N | 122° 40.432' W |              |               | 10        | 12               |           |  |
| 35969   | <b>Columbia Slough at City of Portland Outfall 62</b>  |               |               |                | 45.5864      | -122.67433    |           |                  | 3400g     |  |
|         |  | OF62-1        | 45° 35.182' N | 122° 40.456' W |              |               | 10        | 12               |           |  |
|         |  | OF62-2        | 45° 35.181' N | 122° 40.463' W |              |               | 10        | 12               |           |  |
|         |  | OF62-3        | 45° 35.184' N | 122° 40.471' W |              |               | 10        | 12               |           |  |
|         |  | OF62-4        | 45° 35.187' N | 122° 40.463' W |              |               | 10        | 12               |           |  |
|         |  | OF62-5        | 45° 35.189' N | 122° 40.457' W |              |               | 10        | 12               |           |  |
|         |  | OF62-6        | 45° 35.188' N | 122° 40.450' W |              |               | 10        | 12               |           |  |
| 35970   | <b>Columbia Slough at ODOT outfall</b>                 |               |               |                | 45.58677     | -122.67637    |           |                  | 425g      |  |
|         |  | ODOT-1        | 45° 35.232' N | 122° 40.633' W |              |               | 10        | 2                |           |  |
|         |  | ODOT-2        | 45° 35.218' N | 122° 40.607' W |              |               | 10        | 2                |           |  |
|         |  | ODOT-3        | 45° 35.207' N | 122° 40.588' W |              |               | 10        | 2                |           |  |
|         |  | ODOT-4        | 45° 35.206' N | 122° 40.573' W |              |               | 5         | 2                |           |  |
|         |  | ODOT-5        | 45° 35.197' N | 122° 40.570' W |              |               | 10        | 2                |           |  |
|         |  | ODOT-6        | 45° 35.187' N | 122° 40.547' W |              |               | 5         | 2                |           |  |
| 35971   | <b>Columbia Slough at City of Portland Outfall 61A</b> |               |               |                | 45.58774     | -122.67815    |           |                  | 3400g     |  |
|         |  | OF61A-1       | 45° 35.262' N | 122° 40.692' W |              |               | 10        | 12               |           |  |
|         |  | OF61A-2       | 45° 35.268' N | 122° 40.697' W |              |               | 10        | 12               |           |  |
|         |  | OF61A-3       | 45° 35.268' N | 122° 40.688' W |              |               | 10        | 12               |           |  |
|         |  | OF61A-4       | 45° 35.273' N | 122° 40.693' W |              |               | 5         | 12               |           |  |
|         |  | OF61A-5       | 45° 35.269' N | 122° 40.681' W |              |               | 10        | 12               |           |  |
|         |  | OF61A-6       | 45° 35.260' N | 122° 40.678' W |              |               | 10        | 12               |           |  |
| 35972   | <b>Columbia Slough at Blasen</b>                       |               |               |                | 45.58805     | -122.68234    |           |                  | 425g      |  |
|         |  | Blasen-1      | 45° 35.287' N | 122° 40.973' W |              |               | 10        | 2                |           |  |
|         |  | Blasen-2      | 45° 35.283' N | 122° 40.951' W |              |               | 10        | 2                |           |  |
|         |  | Blasen-3      | 45° 35.282' N | 122° 40.930' W |              |               | 10        | 2                |           |  |
|         |  | Blasen-4      | 45° 35.280' N | 122° 40.906' W |              |               | 10        | 2                |           |  |
| 35973   | <b>Columbia Slough at Schmeer Pumpstation</b>          |               |               |                | 45.58849     | -122.6825     |           |                  | 425g      |  |
|         |  | SchmeerPS-1   | 45° 35.310' N | 122° 40.945' W |              |               | 10        | 2                |           |  |
|         |  | SchmeerPS-2   | 45° 35.308' N | 122° 40.951' W |              |               | 10        | 2                |           |  |
|         |  | SchmeerPS-3   | 45° 35.307' N | 122° 40.957' W |              |               | 10        | 2                |           |  |
|         |  | SchmeerPS-4   | 45° 35.305' N | 122° 40.948' W |              |               | 10        | 2                |           |  |
|         |  | SchmeerPS-5   | 45° 35.312' N | 122° 40.952' W |              |               | 10        | 2                |           |  |
| 35974   | <b>Columbia Slough at City of Portland Outfall 61A</b> |               |               |                | 45.58838     | -122.6836     |           |                  | 3400g     |  |
|         |  | OF61-1        | 45° 35.295' N | 122° 41.014' W |              |               | 10        | 12               |           |  |
|         |  | OF61-2        | 45° 35.300' N | 122° 41.020' W |              |               | 10        | 12               |           |  |
|         |  | OF61-3        | 45° 35.302' N | 122° 41.011' W |              |               | 10        | 12               |           |  |
|         |  | OF61-4        | 45° 35.306' N | 122° 41.018' W |              |               | 7         | 12               |           |  |
|         |  | OF61-5        | 45° 35.310' N | 122° 41.011' W |              |               | 10        | 12               |           |  |
|         |  | OF61-6        | 45° 35.313' N | 122° 41.022' W |              |               | 10        | 12               |           |  |
| 35975   | <b>Columbia Slough at Simpson Cove</b>                 |               |               |                | 45.59103     | -122.68738    |           |                  | 425g      |  |
|         |  | SimpsonCove-1 | 45° 35.475' N | 122° 41.245' W |              |               | 10        | 3                |           |  |
|         |  | SimpsonCove-2 | 45° 35.478' N | 122° 41.241' W |              |               | 10        | 3                |           |  |
|         |  | SimpsonCove-3 | 45° 35.485' N | 122° 41.247' W |              |               | 10        | 3                |           |  |

Appendix B: Table of Targeted Sample Locations

| Lasar # | Site Name   | Site Location | Lattitude     | Longitude      | latitude(dd)<br>Centroid | Longitude(dd)<br>Centroid | depth(cm) | Number of Scoops<br>50mg= 1 scoop | Volume<br>requested |  |
|---------|---|---------------|---------------|----------------|--------------------------|---------------------------|-----------|-----------------------------------|---------------------|--|
| 35976   | <b>Columbia Slough at City of Portland Outfall 60</b> |               |               |                | 45.59096                 | -122.69197                |           |                                   | 3400g               |  |
|         |   | OF60-1        | 45° 35.454' N | 122° 41.515' W |                          |                           | 10        | 12                                |                     |  |
|         |   | OF60-2        | 45° 35.457' N | 122° 41.520' W |                          |                           | 10        | 12                                |                     |  |
|         |   | OF60-3        | 45° 35.458' N | 122° 41.510' W |                          |                           | 10        | 12                                |                     |  |
|         |   | OF60-4        | 45° 35.460' N | 122° 41.516' W |                          |                           | 10        | 12                                |                     |  |
|         |   | OF60-5        | 45° 35.461' N | 122° 41.524' W |                          |                           | 10        | 12                                |                     |  |
|         |   | OF60-6        | 45° 35.463' N | 122° 41.529' W |                          |                           | 10        | 12                                |                     |  |
| 35977   | <b>Columbia Slough at PacificMeats outfall 1</b>      |               |               |                | 45.59131                 | -122.69317                |           |                                   | 3400g               |  |
|         |   | PacMeatOF1-1  | 45° 35.474' N | 122° 41.580' W |                          |                           | 10        | 14                                |                     |  |
|         |   | PacMeatOF1-2  | 45° 35.478' N | 122° 41.585' W |                          |                           | 10        | 14                                |                     |  |
|         |   | PacMeatOF1-3  | 45° 35.482' N | 122° 41.587' W |                          |                           | 10        | 14                                |                     |  |
|         |   | PacMeatOF1-4  | 45° 35.480' N | 122° 41.593' W |                          |                           | 10        | 14                                |                     |  |
|         |   | PacMeatOF1-5  | 45° 35.482' N | 122° 41.599' W |                          |                           | 10        | 14                                |                     |  |
| 35977   | <b>Columbia Slough at PacificMeats outfall 1</b>      |               |               |                | 45.59131                 | -122.69317                |           |                                   | 425g                |  |
|         |   | PacMeatOF1-1  | 45° 35.474' N | 122° 41.580' W |                          |                           | 5         | 4                                 |                     |  |
|         |   | PacMeatOF1-2  | 45° 35.478' N | 122° 41.585' W |                          |                           | 5         | 4                                 |                     |  |
|         |   | PacMeatOF1-3  | 45° 35.482' N | 122° 41.587' W |                          |                           | 5         | 4                                 |                     |  |
|         |   | PacMeatOF1-4  | 45° 35.480' N | 122° 41.593' W |                          |                           | 5         | 4                                 |                     |  |
|         |   | PacMeatOF1-5  | 45° 35.482' N | 122° 41.599' W |                          |                           | 5         | 4                                 |                     |  |
| 35978   | <b>Columbia Slough at PacificMeats outfall 2</b>      |               |               |                | 45.59154                 | -122.69378                |           |                                   | 3400g               |  |
|         |   | PacMeatOF2-1  | 45° 35.501' N | 122° 41.648' W |                          |                           | 10        | 12                                |                     |  |
|         |   | PacMeatOF2-2  | 45° 35.496' N | 122° 41.638' W |                          |                           | 10        | 12                                |                     |  |
|         |   | PacMeatOF2-3  | 45° 35.495' N | 122° 41.629' W |                          |                           | 10        | 12                                |                     |  |
|         |   | PacMeatOF2-4  | 45° 35.495' N | 122° 41.622' W |                          |                           | 10        | 12                                |                     |  |
|         |   | PacMeatOF2-5  | 45° 35.490' N | 122° 41.621' W |                          |                           | 10        | 12                                |                     |  |
|         |   | PacMeatOF2-6  | 45° 35.488' N | 122° 41.614' W |                          |                           | 10        | 12                                |                     |  |
| 35978   | <b>Columbia Slough at PacificMeats outfall 2</b>      |               |               |                | 45.59154                 | -122.69378                |           |                                   | 425g                |  |
|         |   | PacMeatOF2-1  | 45° 35.501' N | 122° 41.648' W |                          |                           | 5         | 3                                 |                     |  |
|         |   | PacMeatOF2-2  | 45° 35.496' N | 122° 41.638' W |                          |                           | 5         | 3                                 |                     |  |
|         |   | PacMeatOF2-3  | 45° 35.495' N | 122° 41.629' W |                          |                           | 5         | 3                                 |                     |  |
|         |   | PacMeatOF2-4  | 45° 35.495' N | 122° 41.622' W |                          |                           | 5         | 3                                 |                     |  |
|         |   | PacMeatOF2-5  | 45° 35.490' N | 122° 41.621' W |                          |                           | 5         | 3                                 |                     |  |
|         |   | PacMeatOF2-6  | 45° 35.488' N | 122° 41.614' W |                          |                           | 5         | 3                                 |                     |  |
| 35979   | <b>Columbia Slough at PacificMeats outfall 3</b>      |               |               |                | 45.59194                 | -122.69536                |           |                                   | 425g                |  |
|         |   | PacMeatOF3-1  | 45. 35. 538 N | 122° 41.748' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | PacMeatOF3-2  | 45. 35. 537 N | 122° 41.738' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | PacMeatOF3-3  | 45. 35. 533 N | 122° 41.725' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | PacMeatOF3-4  | 45. 35. 529 N | 122. 41. 722 W |                          |                           | 10        | 2                                 |                     |  |
|         |   | PacMeatOF3-5  | 45° 35.519' N | 122° 41.707' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | PacMeatOF3-6  | 45° 35.514' N | 122° 41.695' W |                          |                           | 10        | 2                                 |                     |  |
| 35980   | <b>Columbia Slough at City of Portland Outfall 59</b> |               |               |                | 45.59362                 | -122.69936                |           |                                   | 3400g               |  |
|         |   | OF59-1        | 45° 35.609' N | 122° 41.961' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | OF59-2        | 45° 35.612' N | 122° 41.971' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | OF59-3        | 45° 35.619' N | 122° 41.970' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | OF59-4        | 45° 35.615' N | 122° 41.953' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | OF59-5        | 45° 35.625' N | 122° 41.959' W |                          |                           | 10        | 2                                 |                     |  |
|         |   | OF59-6        | 45° 35.622' N | 122° 41.950' W |                          |                           | 10        | 2                                 |                     |  |

## Appendix B: Table of Sample Locations

| LASAR # | Incremental Sampling Locations   |            |   | Latitude           | Longitude     | Actual Latitude* | Actual Longitude* | Sample Depth |   |
|---------|--|------------|---|--------------------|---------------|------------------|-------------------|--------------|---|
|         | Site Name  | Triplicate |   |                    |               |                  |                   |              |   |
|         |  | A          | B |                    |               |                  |                   |              | C |
| 35981   | Columbia Slough btwn between River Mile 5.9 and 8.7. Integrated Sample(IS) |            |   | 45.58660 -122.6746 |               |                  |                   | 8.6          |   |
|         | IS-LS001   | A          | B |                    | 45° 35.004' N | 122° 38.897' W   |                   | 5            |   |
|         | IS-LS002   | A          | B |                    | 45° 34.962' N | 122° 38.992' W   |                   | 5            |   |
|         | IS-LS003   |            | B | C                  | 45° 34.961' N | 122° 39.056' W   |                   | 5            |   |
|         | IS-LS004   | A          | B |                    | 45° 34.960' N | 122° 39.121' W   |                   | N/a          |   |
|         | IS-LS005   | A          | B |                    | 45° 34.997' N | 122° 39.219' W   |                   | 7            |   |
|         | IS-LS006   |            | B | C                  | 45° 34.996' N | 122° 39.283' W   |                   | 10           |   |
|         | IS-LS007   | A          |   |                    | 45° 35.035' N | 122° 39.317' W   |                   | 7            |   |
|         | IS-LS008   | A          |   |                    | 45° 35.034' N | 122° 39.381' W   |                   | 7            |   |
|         | IS-LS009   | A          | B |                    | 45° 35.071' N | 122° 39.480' W   |                   | 7            |   |
|         | IS-LS010   |            | B | C                  | 45° 35.108' N | 122° 39.578' W   |                   | 7            |   |
|         | IS-LS011   |            | B | C                  | 45° 35.107' N | 122° 39.642' W   |                   | 10           |   |
|         | IS-LS012   | A          | B |                    | 45° 35.146' N | 122° 39.676' W   |                   | 5            |   |
|         | IS-LS013   | A          |   | C                  | 45° 35.145' N | 122° 39.740' W   |                   | 7            |   |
|         | IS-LS014   |            | B | C                  | 45° 35.183' N | 122° 39.774' W   |                   | 10           |   |
|         | IS-LS015   | A          |   | C                  | 45° 35.182' N | 122° 39.838' W   |                   | 10           |   |
|         | IS-LS016   |            | B | C                  | 45° 35.181' N | 122° 39.903' W   |                   | 10           |   |
|         | IS-LS017   | A          |   | C                  | 45° 35.141' N | 122° 39.934' W   |                   | 10           |   |
|         | IS-LS018   | A          |   | C                  | 45° 35.100' N | 122° 40.029' W   |                   | 10           |   |
|         | IS-LS019   |            | B | C                  | 45° 35.099' N | 122° 40.093' W   |                   | 10           |   |
|         | IS-LS020   | A          |   | C                  | 45° 35.136' N | 122° 40.191' W   |                   | 10           |   |
|         | IS-LS021   | A          | B |                    | 45° 35.175' N | 122° 40.225' W   |                   | 10           |   |
|         | IS-LS022   | A          |   | C                  | 45° 35.212' N | 122° 40.323' W   |                   | N/a          |   |
|         | IS-LS023   | A          |   | C                  | 45° 35.211' N | 122° 40.388' W   |                   | 10           |   |
|         | IS-LS024   |            |   | C                  | 45° 35.210' N | 122° 40.452' W   |                   | 10           |   |
|         | IS-LS025   | A          |   | C                  | 45° 35.208' N | 122° 40.516' W   |                   | 10           |   |
|         | IS-LS026   |            | B | C                  | 45° 35.207' N | 122° 40.581' W   |                   | 5            |   |
|         | IS-LS027   |            | B |                    | 45° 35.246' N | 122° 40.615' W   |                   | 10           |   |
|         | IS-LS028   | A          | B |                    | 45° 35.283' N | 122° 40.713' W   |                   | 10           |   |
|         | IS-LS029   | A          |   | C                  | 45° 35.282' N | 122° 40.777' W   |                   | 7            |   |
|         | IS-LS030   |            | B |                    | 45° 35.279' N | 122° 40.906' W   |                   | 7            |   |
|         | IS-LS031   | A          | B |                    | 45° 35.317' N | 122° 41.004' W   |                   | 10           |   |
|         | IS-LS032   |            | B | C                  | 45° 35.316' N | 122° 41.069' W   |                   | 10           |   |
|         | IS-LS033   |            |   | C                  | 45° 35.354' N | 122° 41.102' W   |                   | 10           |   |
|         | IS-LS034   | A          |   | C                  | 45° 35.391' N | 122° 41.200' W   |                   | 7            |   |
|         | IS-LS035   | A          | B |                    | 45° 35.390' N | 122° 41.265' W   |                   | 10           |   |
|         | IS-LS036   |            | B | C                  | 45° 35.389' N | 122° 41.329' W   |                   | 7            |   |
|         | IS-LS037   | A          | B |                    | 45° 35.428' N | 122° 41.363' W   |                   | 10           |   |
|         | IS-LS038   |            |   | C                  | 45° 35.426' N | 122° 41.427' W   |                   | 7            |   |
|         | IS-LS039   | A          |   | C                  | 45° 35.465' N | 122° 41.461' W   |                   | 10           |   |
|         | IS-LS040   |            | B | C                  | 45° 35.464' N | 122° 41.526' W   |                   | 5            |   |
|         | IS-LS041   |            | B |                    | 45° 35.502' N | 122° 41.559' W   |                   | 10           |   |
|         | IS-LS042   | A          | B |                    | 45° 35.501' N | 122° 41.624' W   |                   | 7            |   |

Appendix B: Table of Sample Locations

| LASAR # | Incremental Sampling Locations |            |   |   | Latitude      | Longitude      | Actual        | Actual         | Sample |
|---------|--------------------------------|------------|---|---|---------------|----------------|---------------|----------------|--------|
|         | Site Name                      | Triplicate |   |   |               |                | Latitude*     | Longitude*     | Depth  |
|         |                                | A          | B | C |               |                |               |                |        |
|         | IS-LS043                       |            |   | C | 45° 35.538' N | 122° 41.722' W |               |                | 10     |
|         | IS-LS044                       | A          |   |   | 45° 35.576' N | 122° 41.820' W |               |                | 10     |
|         | IS-LS045                       |            | B | C | 45° 35.613' N | 122° 41.918' W |               |                | 7      |
|         | IS-LS046                       | A          | B |   | 45° 35.650' N | 122° 42.016' W |               |                | 10     |
|         | IS-LS047                       | A          |   | C | 45° 35.649' N | 122° 42.081' W |               |                | 10     |
|         | IS-SC048                       | A          |   | C | 45° 35.468' N | 122° 41.268' W | 45. 35. 490 N | 122. 41. 294 W | 10     |
|         | IS-SC049                       |            | B | C | 45° 35.467' N | 122° 41.332' W | 45. 35.482 N  | 122. 41. 351 W | 10     |
|         | IS-SC050                       | A          | B |   | 45° 35.466' N | 122° 41.397' W | 45° 35.466' N | 122. 41.424 W  | 10     |

\* unable to get to exact locations due to very soft sediments and Simpson Cove not boatable-even at high tide(6in of water).

**Appendix B: Sample Location Field Notes**

| Lasar # | FaceLab ID #                        | Site Name  | Abrev. Name | latitude(dd) Centroid | Longitude(dd) Centroid | Date Sampled | Depth (cm-ave) | How Sampled          | Field Comments(numbers refer to specific sample locations)  |
|---------|-------------------------------------|--|-------------|-----------------------|------------------------|--------------|----------------|----------------------|---|
| 35962   | 20090827-35962                      | Columbia Slough at City of Portland Outfall 65       | OF 65       | 45.58259              | -122.65192             | 9/8/2009     | 8              | glass scoop & Dredge | moved site 5 due to gravel bottom   |
| 35963   | 20090823-35963                      | Columbia Slough at PEN 2 Levee                       | PEN 2       | 45.58338              | -122.64805             | 9/2/2009     | 7              | dredge               | a lot of gravel/cobble- changed sample points-within 15ft   |
| 35964   | 20090824-35964                      | Columbia Slough at 13 <sup>th</sup> Ave. pumpstation | 13th        | 45.58435              | -122.65688             | 9/2/2009     | 8              | dredge               |   |
| 35965   | 20090829-35965                      | Columbia Slough at City of Portland Outfall 63       | OF 63       | 45.58485              | -122.66749             | 9/8/2009     | 8              | dredge               | #4 all rocky-moved site; #1 gravel-fair sample  |
| 35966   | 20090828-35966                      | Columbia Slough at City of Portland Outfall 64       | OF 64       | 45.58548              | -122.66557             | 9/8/2009     | 10             | dredge               | #6 oil sheen, #1&3 macrophytes  |
| 35967   | 20090834 FP-35967;20090834 FD-35967 | Columbia Slough at Prec+BR                           | Prec+BR     | 45.58627              | -122.67116             | 9/3/2009     | 10             | glass scoop          | iron seap near #6; Field Duplicate  |
| 35968   | 20090835-35968                      | Columbia Slough at Wastech                           | Wastech     | 45.58684              | -122.67288             | 9/3/2009     | 10             | ss spoon             | #4 sample smelled 'sweet' woody debris present  |
| 35969   | 20090830-35969                      | Columbia Slough at City of Portland Outfall 62       | OF 62       | 45.5864               | -122.67433             | 9/9/2009     | 7              | ss spoon & dredge    | #3, 4 gravel-all others mud   |
| 35970   | 20090825-35970                      | Columbia Slough at ODOT outfall                      | ODOT        | 45.58677              | -122.67637             | 9/3/2009     | 10             | ss spoon             | #1 chemical smell, #6 oil sheen, woody layer at 10-15cm throughout area                                 |
| 35971   | 20090831-35971                      | Columbia Slough at City of Portland Outfall 61A      | OF 61A      | 45.58774              | -122.67815             | 9/9/2009     | 9              | ss spoon             | some woody debris   |
| 35972   | 20090826-35972                      | Columbia Slough at Blasen                            | Blasen      | 45.58805              | -122.68234             | 9/4/2009     | 10             | ss spoon             |   |
| 35973   | 20090832-35973                      | Columbia Slough at Schmeer Pumpstation               | Schmeer     | 45.58849              | -122.6825              | 9/9/2009     | 10             | dredge               |   |
| 35974   | 20090853 FP-35974;20090853 FD-35974 | Columbia Slough at City of Portland Outfall 61       | OF 61       | 45.58838              | -122.6836              | 9/14/2009    | 9              | ss spoon & dredge    | silt and gravel; Field Duplicate  |
| 35975   | 20090833-35975                      | Columbia Slough at Simpson Cove                      | SMPCV       | 45.59103              | -122.68738             | 9/9/2009     | 10             | ss spoon             | moved sample points due to soft mud and unboatable(6in of water at high tide) iron precipitate in cove  |
| 35976   | 20090854-35976                      | Columbia Slough at City of Portland Outfall 60       | OF 60       | 45.59096              | -122.69197             | 9/14/2009    | 10             | ss spoon & dredge    | Dyena outfall flowing; #6 woody debris  |
| 35977   | 20090860-35977 10cm                 | Columbia Slough at PacificMeats outfall 1            | PM 1-10     | 45.59131              | -122.69317             | 9/15/2009    | 10             | ss spoon & dredge    | sweet smell?  |
| 35977   | 20090859-35977 5cm                  | Columbia Slough at PacificMeats outfall 1            | PM 1-5      | 45.59131              | -122.69317             | 9/15/2009    | 5              | ss spoon & dredge    | sweet smell?  |
| 35978   | 20090857-35978 10cm                 | Columbia Slough at PacificMeats outfall 2            | PM 2-10     | 45.59154              | -122.69378             | 9/15/2009    | 10             | ss spoon & dredge    | oil sheen at about 10cm depth? Not on surface-below foot prints, etc                                    |
| 35978   | 20090858-35978 5cm                  | Columbia Slough at PacificMeats outfall 2            | PM 2-5      | 45.59154              | -122.69378             | 9/15/2009    | 5              | ss spoon & dredge    |   |
| 35979   | 20090855-35979                      | Columbia Slough at PacificMeats outfall 3            | PM 3        | 45.59194              | -122.69536             | 9/14/2009    | 10             | dredge               | very soft sediment, unable to get exact lat due to tide level and too soft to walk on sediment          |
| 35980   | 20090856-35980                      | Columbia Slough at City of Portland Outfall 59       | OF 59       | 45.59362              | -122.69936             | 9/15/2009    | 10             | ss spoon & dredge    | oily sheen; asphalt shingle 'gravel' at outfall pipe continues ~40ft to a fan at edge of low tide mark. |

Appendix C:  
Lower Columbia Slough between River Mile 5.9 and 8.7 Summary Statistics

|                        |   |                 | % of total       | % of total   |                   |             |              |
|------------------------|---|-----------------|------------------|--------------|-------------------|-------------|--------------|
|                        | Columbia Slough RM 5.9 and 8.7 Sampling area        |                 | exceeding        | exceeding    |                   |             |              |
|                        |   | % Detect        | Risk base Values | 2009 IS Mean | %Estimates        | Mean        | Range        |
| <b>Metals</b>          | Total Aluminum                                      | 100%            | N/A              | 38%          | 0%                | 15765.4     | 10800-22000  |
| (ppm)                  | Total Antimony                                      | 46%             | 0%               | 46%          | 27%               | 0.9         | <0.50-3.05   |
|                        | Total Arsenic                                       | 100%            | 27%              | 58%          | 0%                | 6.6         | 2.5-22.6     |
|                        | Total Barium  | 100%            | N/A              | 42%          | 0%                | 120.9       | 80-160       |
|                        | Total Cadmium                                       | 81%             | 35%              | 54%          | 0%                | 0.9         | <0.5-1.9     |
|                        | Total Chromium                                      | 100%            | 50%              | 46%          | 0%                | 59.8        | 19-481       |
|                        | Total Cobalt  | 100%            | N/A              | 69%          | 0%                | 11.0        | 8.05-19.5    |
|                        | Total Copper  | 100%            | 62%              | 65%          | 0%                | 56.6        | 17.2-148     |
|                        | Total Lead  | 100%            | 92%              | 62%          | 0%                | 65.1        | 13.4-212     |
|                        | Total Manganese                                     | 100%            | 0%               | 46%          | 0%                | 425.5       | 203-732      |
|                        | Mercury   | 100%            | 81%              | 54%          | 4%                | 0.1         | 0.02-0.63    |
|                        | Total Nickel  | 100%            | 12%              | 65%          | 0%                | 32.8        | 15-255       |
|                        | Total Selenium                                      | 0%              | N/A              | N/A          | 0%                | 1.0         | <1           |
|                        | Total Silver  | 58%             | 0%               | 0%           | 0%                | 0.2         | <0.2-0.49    |
|                        | Total Thallium                                      | 0%              | N/A              | N/A          | 0%                | 0.2         | <0.2         |
|                        | Total Zinc  | 100%            | 85%              | 54%          | 0%                | 300.3       | 98-875       |
|                        |   |                 | % Exceeding      |              |                   | ug/kg       | ug/kg        |
| <b>Columbia Slough</b> | <b>Columbia Slough RM 5.9 and 8.7 Sampling area</b> | <b>% Detect</b> | <b>Risk base</b> |              | <b>%Estimates</b> | <b>Mean</b> | <b>Range</b> |
| <b>Semi-Vols</b>       | 1,2,4,5-Tetrachlorobenzene                          | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
| (ppb)                  | 1,2,4-Trichlorobenzene                              | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 1,2-Dichlorobenzene                                 | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 1,3-Dichlorobenzene                                 | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 1,4-Dichlorobenzene                                 | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,3,4,6-Tetrachlorophenol                           | 4%              | N/A              | 0%           | 4%                | <66.0       | <66.0- 65    |
|                        | 2,3,5,6-Tetrachlorophenol                           | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,4,5-Trichlorophenol                               | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,4,6-Trichlorophenol                               | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,4-Dichlorophenol                                  | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,4-Dimethylphenol                                  | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,4-Dinitrophenol                                   | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,4-Dinitrotoluene                                  | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,6-Dichlorophenol                                  | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2,6-Dinitrotoluene                                  | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2-Chloronaphthalene                                 | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2-Chlorophenol                                      | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2-Methylphenol                                      | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 2-Nitrophenol                                       | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 4,6-Dinitro-2-methylphenol                          | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 4-Bromophenyl phenyl ether                          | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 4-Chloro-3-methylphenol                             | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | 4-Methylphenol (p-Cresol)                           | 4%              | N/A              | 0%           | 0%                | 49.2        | <66.0-95     |
|                        | Acenaphthene  | 8%              | N/A              | 0%           | 8%                | <66.0       | <66.0        |
|                        | Acenaphthylene                                      | 0%              | N/A              | N/A          | 0%                | <66.0       | <66.0        |
|                        | Anthracene  | 23%             | 15%              | 15%          | 12%               | 76.2        | <33 -620     |
|                        | Benzo[a]anthracene                                  | 85%             | 88%              | 62%          | 23%               | 167.2       | <33 -695     |
|                        | Benzo[a]pyrene                                      | 85%             | 85%              | 65%          | 23%               | 183.0       | <33 -717     |
|                        | Benzo[b]fluoranthene                                | 88%             | 88%              | 54%          | 0%                | 324.2       | <33 -1380    |
|                        | Benzo[g,h,i]perylene                                | 31%             | 19%              | 46%          | 15%               | 176.4       | <33 -457     |
|                        | Benzo[k]fluoranthene                                | 54%             | 54%              | 42%          | 23%               | 96.6        | <33 -465     |
|                        | Bis(2-Chloroethoxy) methane                         | 0%              | N/A              | N/A          | 0                 | <66.0       | <66.0        |
|                        | Bis(2-Chloroethyl) ether                            | 0%              | N/A              | N/A          | 0                 | <66.0       | <66.0        |
|                        | Bis(2-Chloroisopropyl) ether                        | 0%              | N/A              | N/A          | 0                 | <66.0       | <66.0        |
|                        | Bis(2-ethylhexyl)adipate                            | 19%             | N/A              | 15%          | 0%                | 61.6        | <33.0 - 177  |
|                        | Bis(2-ethylhexyl)phthalate                          | 100%            | 35%              | 58%          | 12%               | 1505.4      | 164-5349     |
|                        | Butylbenzylphthalate                                | 46%             | N/A              | 31%          | 23%               | 149.9       | <49-1159     |

Appendix C:  
Lower Columbia Slough between River Mile 5.9 and 8.7 Summary Statistics

| Columbia Slough | Columbia Slough RM 5.9 and 8.7 Sampling at | % Detect | % Exceeding |     | %Estimates | ug/kg |              |
|-----------------|--|----------|-------------|-----|------------|-------|--------------|
|                 |  |          | Risk base   |     |            | Mean  | Range        |
|                 | Chrysene                                   | 92%      | 88%         | 62% | 15%        | 243.9 | <33.0 - 1050 |
|                 | di-n-Butylphthalate                        | 23%      | N/A         | 15% | 23%        | 91.4  | <49-392      |
|                 | Di-n-octylphthalate                        | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Dibenz[a,h]anthracene                      | 12%      | 8%          | 4%  | 8%         | 52.8  | <33-140      |
|                 | Dibenzofuran                               | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Diethylphthalate                           | 0%       | N/A         | N/A | 0%         | <100  | <100         |
|                 | Dimethylphthalate                          | 12%      | N/A         | 0%  | 0%         | 93.4  | <49-368      |
|                 | Fluoranthene                               | 96%      | 88%         | 65% | 4%         | 513.8 | <33-2302     |
|                 | Fluorene                                   | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Hexachloro-1,3-Butadiene                   | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Hexachlorocyclopentadiene                  | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Hexachloroethane                           | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Indeno[1,2,3-cd]pyrene                     | 81%      | 81%         | 42% | 19%        | 140.7 | <66-597      |
|                 | Isophorone                                 | 15%      | N/A         | 0%  | 15%        | 55.8  | <66-139      |
|                 | n-Nitroso-di-n-dipropylamine               | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | n-Nitrosodiphenylamine                     | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Naphthalene                                | 4%       | 4%          | 4%  | 4%         | 54.0  | <66-183      |
|                 | Nitrobenzene                               | 0%       | N/A         | N/A | 0%         | <66   | <66          |
|                 | Pentachlorophenol                          | 27%      | 27%         | 27% | 4%         | 98.1  | <66-426      |
|                 | Phenanthrene                               | 62%      | 62%         | 58% | 17%        | 184.9 | <66-870      |
|                 | Phenol                                     | 0%       | N/A         | N/A | 0%         | 48.2  | <66          |
|                 | Pyrene                                     | 96%      | 0%          | 58% | 4%         | 542.3 | <33-2020     |

|  |  | % Exceeding |           |            | ug/kg  |             |                |
|--|--|-------------|-----------|------------|--------|-------------|----------------|
| Columbia Slough                        | Columbia Slough RM 5.9 and 8.7 Sampling at | % Detect    | Risk base | %Estimates | Mean   | Range       |                |
|  |  | % Exceeding |           |            | ug/kg  |             |                |
| Columbia Slough OF 59-65 Sampling area |  | % Detect    | Risk base | %Estimates | Mean   | Range       |                |
| Pesticides                             | Aldrin                                     | 0%          | 0%        | 0%         | <6.3   | 0.396 -11   |                |
| Method                                 | alpha-BHC                                  | 0%          | 0%        | 0%         | <6.9   | 0.43 -12    |                |
| 8081                                   | beta-BHC                                   | 0%          | 0%        | 0%         | <8.1   | 0.504 -14   |                |
| (ppb)                                  | delta-BHC                                  | 15%         | N/A       | 15%        | 7.3    | 0.353 -15.3 |                |
|  | gamma-BHC (Lindane)                        | 0%          | 0%        | 0%         | <9.5   | 0.594 -16.5 |                |
|  | Chlordane                                  | 0%          | 0%        | 0%         | <36.2  | 2.26 -62.8  |                |
|  | alpha-Chlordane                            | 19%         | N/A       | 15%        | 6.1    | 0.563 -15.6 |                |
|  | gamma-Chlordane                            | 23%         | N/A       | 23%        | 9.9    | 0.603 -30.2 |                |
|  | 4,4'-DDD                                   | 15%         | 15%       | 8%         | 15%    | 3.6         | 0.927 -25.6    |
|  | 4,4'-DDE                                   | 46%         | 46%       | 27%        | 31%    | 6.6         | 0.4 -12.9      |
|  | 4,4'-DDT                                   | 19%         | 19%       | 12%        | 19%    | 23.5        | 1.98 -63.8     |
|  | Dieldrin                                   | 19%         | 15%       | 15%        | 19%    | 3.7         | 0.354 -9.79    |
|  | Endosulfan I                               | 0%          | N/A       | 0%         | <8.1   | 0.502 -13.9 |                |
|  | Endosulfan II                              | 4%          | N/A       | 4%         | 4%     | 0.5         | 0.436 -12.1    |
|  | Endosulfan sulfate                         | 12%         | N/A       | 12%        | 12%    | 6.8         | 0.567 -18.4    |
|  | Endrin                                     | 0%          | N/A       | 0%         | <7.7   | 0.475 -13.2 |                |
|  | Endrin aldehyde                            | 0%          | 0%        | 0%         | <7.9   | 0.49 -13.6  |                |
|  | Endrin ketone                              | 0%          | N/A       | 0%         | <18.8  | 1.17 -32.6  |                |
|  | Heptachlor                                 | 12%         | 12%       | 12%        | 12%    | 5.5         | 0.436 -12      |
|  | Heptachlor epoxide                         | 0%          | N/A       | 0%         | <6.8   | 0.426 -11.8 |                |
|  | Methoxychlor                               | 0%          | N/A       | 0%         | <18.8  | 1.1 -30.6   |                |
|  | Toxaphene                                  | 0%          | N/A       | 0%         | <451.2 | 28.1 -779   |                |
|  |  | % Exceeding |           |            | ug/kg  |             |                |
| Columbia Slough IS samples- 3 samples  |  | % Detect    | Risk base | %Estimates | Mean   | Range       |                |
| Pesticides                             | 2,4'-DDD                                   | 100%        | N/A       | N/A        | 0%     | 0.91        | 0.818 -1.07    |
| ug/kg                                  | 2,4'-DDE                                   | 100%        | N/A       | N/A        | 0%     | 0.31        | 0.271 -0.381   |
| Method                                 | 2,4'-DDT                                   | 67%         | N/A       | N/A        | 0%     | <0.12       | 0.0662 -0.172  |
| 1669                                   | 4,4'-DDD                                   | 100%        | 100%      | N/A        | 0%     | 3.70        | 3.14 -4.43     |
|  | 4,4'-DDE                                   | 100%        | 100%      | N/A        | 100%   | 6.31        | 5.93 -6.86     |
|  | 4,4'-DDT                                   | 100%        | 0%        | N/A        | 0%     | 0.47        | 0.405 -0.53    |
|  | 4,4'-Methoxychlor                          | 0%          | N/A       | N/A        | 0%     | <0.55       | <0.628         |
|  | Aldrin                                     | 67%         | 0%        | N/A        | 0%     | 0.18        | 0.172 -0.208   |
|  | alpha-BHC                                  | 33%         | 0%        | N/A        | 0%     | <0.02       | 0.0168 -0.0194 |
|  | beta-BHC                                   | 33%         | 0%        | N/A        | 0%     | 0.02        | 0.017 -0.022   |
|  | Lindane (gamma-BHC)                        | 0%          | 0%        | N/A        | 0%     | <0.03       | <0.0289        |
|  | cis-Chlordane (alpha)                      | 100%        | N/A       | N/A        | 0%     | 1.01        | 0.95 -1.09     |
|  | trans-Chlordane (gamma)                    | 100%        | N/A       | N/A        | 0%     | 0.97        | 0.927 -1.06    |
|  | Oxychlordane                               | 0%          | N/A       | N/A        | 0%     | <0.05       | <0.0635        |
|  | cis-Heptachlor Epoxide                     | 0%          | N/A       | N/A        | 0%     | <0.05       | <0.0593        |
|  | cis-Nonachlor                              | 100%        | N/A       | N/A        | 0%     | 0.32        | 0.295 -0.346   |
|  | delta-BHC                                  | 0%          | N/A       | N/A        | 0%     | <0.02       | <0.0166        |
|  | Dieldrin                                   | 100%        | 0%        | N/A        | 0%     | 0.70        | 0.644 -0.763   |
|  | Endosulfan I (alpha)                       | 0%          | N/A       | N/A        | 0%     | <0.07       | <0.0891        |
|  | Endosulfan II (beta)                       | 0%          | N/A       | N/A        | 0%     | <0.09       | <0.174         |
|  | Endosulfan Sulfate                         | 67%         | N/A       | N/A        | 0%     | 0.38        | 0.281 -0.485   |
|  | Endrin                                     | 0%          | N/A       | N/A        | 0%     | <0.15       | <0.239         |
|  | Endrin Aldehyde                            | 0%          | 0%        | N/A        | 0%     | <0.29       | <0.407         |
|  | Endrin Ketone                              | 0%          | N/A       | N/A        | 0%     | <0.36       | <0.439         |
|  | Heptachlor                                 | 0%          | 0%        | N/A        | 0%     | <0.08       | <0.0817        |
|  | Hexachlorobenzene                          | 100%        | N/A       | N/A        | 0%     | 0.37        | 0.192 -0.71    |
|  | Hexachlorobutadiene                        | 0%          | N/A       | N/A        | 0%     | <0.02       | <0.0169        |
|  | Mirex                                      | 0%          | N/A       | N/A        | 0%     | <0.11       | <0.135         |
|  | trans-Heptachlor Epoxide                   | 0%          | N/A       | N/A        | 0%     | <0.14       | <0.242         |
|  | trans-Nonachlor                            | 100%        | N/A       | N/A        | 0%     | 0.71        | 0.7 -0.721     |

|                 |  | % Exceeding |           |       | ug/kg      |       | ug/kg      |
|-----------------|--|-------------|-----------|-------|------------|-------|------------|
| Columbia Slough | Columbia Slough RM 5.9 and 8.7 Sampling area | % Detect    | Risk base |       | %Estimates | Mean  | Range      |
|                 |  | % Exceeding |           |       | ug/kg      |       | ug/kg      |
|                 |  | % Exceeding |           |       | ug/kg      |       | ug/kg      |
|                 |  | % Exceeding |           |       | ug/kg      |       | ug/kg      |
| <b>PCBs</b>     | Aroclor-1016                                 | 0%          | N/A       | N/A   | 0%         | <6.1  | <6.1       |
| (ppb)           | Aroclor-1242                                 | 0%          | N/A       | N/A   | 0%         | <6.1  | <6.1       |
|                 | Aroclor 1221                                 | 0%          | N/A       | N/A   | 0%         | <8.5  | <8.5       |
|                 | Aroclor 1232                                 | 0%          | N/A       | N/A   | 0%         | <11.3 | <11.3      |
|                 | Aroclor 1248                                 | 8%          | 8%        | 8%    | 0%         | 13.5  | <4.4-230   |
|                 | Aroclor 1254                                 | 54%         | 54%       | 46%   | 54%        | 58.0  | <2.5-333   |
|                 | Aroclor 1260                                 | 73%         | 62%       | 54%   | 62%        | 86.9  | <5.4-855   |
|                 | Total Congeners                              | 100%        | N/a       | N/A   | 0%         | 116.0 | 18-904     |
|                 |  | % Exceeding |           |       |            |       |            |
|                 |  | % Exceeding |           |       |            |       |            |
|                 |  | % Exceeding |           |       |            |       |            |
|                 |  | % Exceeding |           |       |            |       |            |
| <b>PBDE</b>     | Total PBDE                                   | 100         | N/A       | 58%   | 100%       | 15    | 3.3-36.9   |
| (ppb)           | BDE-47                                       | 100         | N/A       | 67%   | 100%       | 3.0   | 0.62-11.8  |
| <b>TBT(ppb)</b> | Tributyltin                                  | 62.5%       | 6.3%      | 15.4% | 25.0%      | 32.0  | <5-285     |
| <b>General</b>  | Total Organic Carbon(ppm)                    | 100%        | N/A       | N/A   | 0%         | 13436 | 8500-22000 |
|                 | Percent Solids(ppm)                          | N/A         | N/A       | N/A   | 0%         | 84%   | 72-89%     |
|                 | Clay(%)                                      | 100%        | N/A       | N/A   | 0%         | 7%    | 1-20%      |
|                 | Silt(%)                                      | 100%        | N/A       | N/A   | 0%         | 34%   | 1-77%      |
|                 | Sand(%)                                      | 100%        | N/A       | N/A   | 0%         | 52%   | 9-87%      |
|                 | Gravel(%)                                    | 62%         | N/A       | N/A   | 0%         | 7%    | 0-41%      |
|                 | Field Sample Depth(m)                        | N/A         | N/A       | N/A   | N/A        | 0.1   | 0.05-0.1   |
|                 | Final Weight(g)                              | N/A         | N/A       | N/A   | N/A        | 2611  | 2315-3290  |
|                 | Initial Weight(g)                            | N/A         | N/A       | N/A   | N/A        | 3112  | 2676-3848  |









Appendix E Lower Columbia Slough Sediment between River Mile 5.9 and 8.7.  
Duplicate QA/QC Review

| Columbia Slough OF 59-65     |                              |           | Precision acceptable when <20% |       |         |                                       |         |        |         |                                       |         |
|------------------------------|------------------------------|-----------|--------------------------------|-------|---------|---------------------------------------|---------|--------|---------|---------------------------------------|---------|
|                              |                              |           | OF 61 1                        |       | OF 61 2 | Precision relative percent difference |         | P&BR 1 | P&BR 2  | Precision relative percent difference |         |
| Sampling 2009                |                              |           | 9/14/09                        |       | 9/14/09 |                                       | 9/10/09 |        | 9/10/09 |                                       |         |
| Metals                       | Total Aluminum               | mg/Kg dry | 14700                          |       | 13800   | 6.3%                                  |         |        | 15900   | 16700                                 | -4.9%   |
|                              | Total Antimony               | mg/Kg dry | 1.67                           |       | 2.56    | -42.1%                                |         |        | 0.73    | 0.82                                  | -11.6%  |
|                              | Total Arsenic                | mg/Kg dry | 9.8                            |       | 22.6    | -79.0%                                |         |        | 4.2     | 4.2                                   | 0.0%    |
|                              | Total Barium                 | mg/Kg dry | 114                            |       | 116     | -1.7%                                 |         |        | 119     | 127                                   | -6.5%   |
|                              | Total Cadmium                | mg/Kg dry | 1.37                           |       | 1.2     | 13.2%                                 |         | 0.5 U  | 0.5 U   | 0.0%                                  |         |
|                              | Total Chromium               | mg/Kg dry | 79.8                           |       | 67.3    | 17.0%                                 |         |        | 37.2    | 40.2                                  | -7.8%   |
|                              | Total Cobalt                 | mg/Kg dry | 10.9                           |       | 11      | -0.9%                                 |         |        | 9.78    | 10.4                                  | -6.1%   |
|                              | Total Copper                 | mg/Kg dry | 79.4                           |       | 147     | -59.7%                                |         |        | 25.8    | 24.1                                  | 6.8%    |
|                              | Total Lead                   | mg/Kg dry | 212                            |       | 116     | 58.5%                                 |         |        | 129     | 139                                   | -7.5%   |
|                              | Total Manganese              | mg/Kg dry | 473                            |       | 495     | -4.5%                                 |         |        | 374     | 379                                   | -1.3%   |
|                              | Mercury                      | mg/Kg dry | 0.12                           |       | 0.13    | -8.0%                                 |         |        | 0.1     | 0.11                                  | -9.5%   |
|                              | Total Nickel                 | mg/Kg dry | 47.4                           |       | 37.1    | 24.4%                                 |         |        | 17.8    | 18.3                                  | -2.8%   |
|                              | Total Selenium               | mg/Kg dry | 1                              | U     | 1       | 0.0%                                  |         | 1 U    | 1 U     | 0.0%                                  |         |
| Total Silver                 | mg/Kg dry                    | 0.26      |                                | 0.3   | -14.3%  |                                       | 0.2 U   | 0.2 U  | 0.0%    |                                       |         |
| Total Thallium               | mg/Kg dry                    | 0.2       | U                              | 0.2   | 0.0%    |                                       | 0.2 U   | 0.2 U  | 0.0%    |                                       |         |
| Total Zinc                   | mg/Kg dry                    | 487       |                                | 622   | -24.3%  |                                       |         | 109    | 123     | -12.1%                                |         |
| Semi-Vols                    | 1,2,4,5-Tetrachlorobenzene   | µg/Kg dry | 66 U,E                         |       | 66 U,E  | 0.0%                                  |         |        | 33 U,E  | 33 U,E                                | 0.0%    |
|                              | 1,2,4-Trichlorobenzene       | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 1,2-Dichlorobenzene          | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 1,3-Dichlorobenzene          | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 1,4-Dichlorobenzene          | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,3,4,6-Tetrachlorophenol    | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,3,5,6-Tetrachlorophenol    | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,4,5-Trichlorophenol        | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,4,6-Trichlorophenol        | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,4-Dichlorophenol           | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,4-Dimethylphenol           | µg/Kg dry | 130 U                          |       | 130 U   | 0.0%                                  |         |        | 67 U    | 67 U,E                                | 0.0%    |
|                              | 2,4-Dinitrophenol            | µg/Kg dry | Void                           |       | Void    | #VALUE!                               |         |        | Void    | Void                                  | #VALUE! |
|                              | 2,4-Dinitrotoluene           | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,6-Dichlorophenol           | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2,6-Dinitrotoluene           | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2-Chloronaphthalene          | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2-Chlorophenol               | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2-Methylphenol               | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 2-Nitrophenol                | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 4,6-Dinitro-2-methylphenol   | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | Void    | Void                                  | #VALUE! |
|                              | 4-Bromophenyl phenyl ether   | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 4-Chloro-3-methylphenol      | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | 4-Methylphenol (p-Cresol)    | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Acenaphthene                 | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 37.4 E  | 52.5 E                                | -33.6%  |
|                              | Acenaphthylene               | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Anthracene                   | µg/Kg dry | 620.3                          |       | 66      | 161.5%                                |         |        | 49.9    | 82.5                                  | -49.2%  |
|                              | Benzo[a]anthracene           | µg/Kg dry | 198.5                          |       | 134.7   | 38.3%                                 |         |        | 232.2   | 282.4                                 | -19.5%  |
|                              | Benzo[a]pyrene               | µg/Kg dry | 332.5                          |       | 139.7   | 81.7%                                 |         |        | 189.7   | 234.9                                 | -21.3%  |
|                              | Benzo[b]fluoranthene         | µg/Kg dry | 620.3                          |       | 299.4   | 69.8%                                 |         |        | 309.5   | 329.8                                 | -6.4%   |
|                              | Benzo[g,h,i]perylene         | µg/Kg dry | 397                            |       | 194.6   | 68.4%                                 |         |        | 104.8   | 119.9                                 | -13.4%  |
|                              | Benzo[k]fluoranthene         | µg/Kg dry | 158.8                          |       | 66 U    | 82.6%                                 |         |        | 109.8   | 122.4                                 | -10.9%  |
|                              | Bis(2-Chloroethoxy) methane  | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Bis(2-Chloroethyl) ether     | µg/Kg dry | 66 U,E                         |       | 67 U,E  | -1.5%                                 |         |        | 33 U,E  | 33 U,E                                | 0.0%    |
|                              | Bis(2-Chloroisopropyl) ether | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Bis(2-ethylhexyl)adipate     | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Bis(2-ethylhexyl)phthalate   | µg/Kg dry | 4387.1                         |       | 5349    | -19.8%                                |         |        | 626.6   | 842.1                                 | -29.3%  |
|                              | Butylbenzylphthalate         | µg/Kg dry | 213.4                          |       | 99.8 U  | 72.5%                                 |         |        | 49.9 U  | 50 U                                  | -0.2%   |
|                              | Chrysene                     | µg/Kg dry | 466.5                          |       | 189.6   | 84.4%                                 |         |        | 327     | 369.8                                 | -12.3%  |
|                              | di-n-Butylphthalate          | µg/Kg dry | 392.1 E                        |       | 99.8 U  | 118.8%                                |         |        | 49.9 U  | 50 U                                  | -0.2%   |
|                              | Di-n-octylphthalate          | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Dibenz[a,h]anthracene        | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Dibenzofuran                 | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Diethylphthalate             | µg/Kg dry | 99.2 U                         |       | 99.8 U  | -0.6%                                 |         |        | 49.9 U  | 50 U                                  | -0.2%   |
|                              | Dimethylphthalate            | µg/Kg dry | 99.2 U                         |       | 99.8 U  | -0.6%                                 |         |        | 49.9 U  | 50 U                                  | -0.2%   |
|                              | Fluoranthene                 | µg/Kg dry | 1206                           |       | 334.3   | 113.2%                                |         |        | 748.9   | 924.5                                 | -21.0%  |
|                              | Fluorene                     | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
|                              | Hexachloro-1,3-Butadiene     | µg/Kg dry | 66 U                           |       | 66 U    | 0.0%                                  |         |        | 33 U    | 33 U                                  | 0.0%    |
| Hexachlorocyclopentadiene    | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| Hexachloroethane             | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| Indeno[1,2,3-cd]pyrene       | µg/Kg dry                    | 297.8     |                                | 134.7 | 75.4%   |                                       |         | 99.9   | 114.9   | -14.0%                                |         |
| Isophorone                   | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 39.9 E | 72      | -57.4%                                |         |
| n-Nitroso-di-n-dipropylamine | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| n-Nitrosodiphenylamine       | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| Naphthalene                  | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| Nitrobenzene                 | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| Pentachlorophenol            | µg/Kg dry                    | 426.8     |                                | 254.5 | 50.6%   |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| Phenanthrene                 | µg/Kg dry                    | 645.2     |                                | 129.7 | 133.0%  |                                       |         | 297.1  | 397.3   | -28.9%                                |         |
| Phenol                       | µg/Kg dry                    | 66 U      |                                | 66 U  | 0.0%    |                                       |         | 33 U   | 33 U    | 0.0%                                  |         |
| Pyrene                       | µg/Kg dry                    | 1429.3    |                                | 364.3 | 118.8%  |                                       |         | 773.8  | 949.5   | -20.4%                                |         |

Appendix E Lower Columbia Slough Sediment between River Mile 5.9 and 8.7.  
Duplicate QA/QC Review

| Columbia Slough OF 59-65<br><br>Sampling 2009 |                         |   | Precision acceptable when <20% |           |                |           |                             |                |   |                |   |                             |  |
|---|-------------------------|---|--------------------------------|-----------|----------------|-----------|-----------------------------|----------------|---|----------------|---|-----------------------------|--|
|   |                         |   | OF 61 1                        |           | OF 61 2        |           | Precision                   | P&BR 1         |   | P&BR 2         |   | Precision                   |  |
|   |                         |   | 9/14/09                        |           | 9/14/09        |           | relative percent difference | 9/10/09        |   | 9/10/09        |   | relative percent difference |  |
| <b>Pesticides</b>                             | Aldrin                  | ug/kg                                     | 8.6                            | U         | 8.35           | U         | 2.9%                        | 0.399          | U | 1.7            | U | -124.0%                     |  |
|   | alpha-BHC               | ug/kg                                     | 9.35                           | U         | 9.08           | U         | 2.9%                        | 0.434          | U | 1.7            | U | -118.7%                     |  |
| Method  | beta-BHC                | ug/kg                                     | 10.9                           | U         | 10.6           | U         | 2.8%                        | 0.508          | U | 1.7            | U | -108.0%                     |  |
| 8081  | delta-BHC               | ug/kg                                     | 7.63                           | U         | 7.41           | U         | 2.9%                        | <b>2.86</b>    |   | 1.7            | U | 50.9%                       |  |
|   | gamma-BHC (Lindane)     | ug/kg                                     | 12.9                           | U         | 12.5           | U         | 3.1%                        | 0.598          | U | 1.7            | U | -95.9%                      |  |
|   | Chlordane               | ug/kg                                     | 49.1                           | U         | 47.7           | U         | 2.9%                        | 3.36           | U | 3.32           | U | 1.2%                        |  |
|   | alpha-Chlordane         | ug/kg                                     | 12.2                           | U         | 11.9           | U         | 2.5%                        | <b>7.92</b>    |   | <b>7.44</b>    |   | 6.2%                        |  |
|   | gamma-Chlordane         | ug/kg                                     | 13.1                           | U         | 12.7           | U         | 3.1%                        | <b>5.74</b>    |   | <b>5.44</b>    |   | 5.4%                        |  |
|   | 4,4'-DDD                | ug/kg                                     | 20                             | U         | 19.5           | U         | 2.5%                        | <b>5.58</b>    |   | <b>4.11</b>    |   | 30.3%                       |  |
|   | 4,4'-DDE                | ug/kg                                     | <b>9.6</b>                     | J         | <b>12.7</b>    | J         | -27.8%                      | <b>4.51</b>    |   | 3.32           | U | 30.4%                       |  |
|   | 4,4'-DDT                | ug/kg                                     | 43                             | U         | 41.7           | U         | 3.1%                        | 1.99           | U | 3.32           | U | -50.1%                      |  |
|   | Dieldrin                | ug/kg                                     | 7.65                           | U         | 7.43           | U         | 2.9%                        | 0.355          | U | 3.32           | U | -161.4%                     |  |
|   | Endosulfan I            | ug/kg                                     | 10.9                           | U         | 10.6           | U         | 2.8%                        | 0.506          | U | 1.7            | U | -108.3%                     |  |
|   | Endosulfan II           | ug/kg                                     | 9.48                           | U         | 9.2            | U         | 3.0%                        | 0.44           | U | 3.32           | U | -153.2%                     |  |
|   | Endosulfan sulfate      | ug/kg                                     | 12.3                           | U         | 12             | U         | 2.5%                        | 0.572          | U | 3.32           | U | -141.2%                     |  |
|   | Endrin                  | ug/kg                                     | 10.3                           | U         | 10             | U         | 3.0%                        | 0.479          | U | 3.32           | U | -149.6%                     |  |
|   | Endrin aldehyde         | ug/kg                                     | 10.6                           | U         | 10.3           | U         | 2.9%                        | 0.494          | U | 3.32           | U | -148.2%                     |  |
|   | Endrin ketone           | ug/kg                                     | 25.5                           | U         | 24.7           | U         | 3.2%                        | 1.18           | U | 3.32           | U | -95.1%                      |  |
|   | Heptachlor              | ug/kg                                     | 9.41                           | U         | 9.14           | U         | 2.9%                        | 0.437          | U | 1.7            | U | -118.2%                     |  |
|   | Heptachlor epoxide      | ug/kg                                     | 9.26                           | U         | 8.99           | U         | 3.0%                        | 0.43           | U | 1.7            | U | -119.2%                     |  |
|   | Methoxychlor            | ug/kg                                     | 23.9                           | U         | 23.3           | U         | 2.5%                        | 16.8           | U | 16.7           | U | 0.6%                        |  |
|   | Toxaphene               | ug/kg                                     | 609                            | U         | 592            | U         | 2.8%                        | 67.3           | U | 66.5           | U | 1.2%                        |  |
| <b>PCBs</b>                                   | Aroclor-1016 (PCB-1016) | µg/Kg dry                                 | 5.1                            | U         | 5.3            | U         | -3.8%                       | 4.8            | U | 5.1            | U | -6.1%                       |  |
|   | Aroclor-1242 (PCB-1242) | µg/Kg dry                                 | 2.5                            | U         | 2.6            | U         | -3.9%                       | 2.4            | U | 2.6            | U | -8.0%                       |  |
|   | Aroclor 1221            | µg/Kg dry                                 | 3.5                            | U         | 3.7            | U         | -5.6%                       | 3.4            | U | 3.6            | U | -5.7%                       |  |
|   | Aroclor 1232            | µg/Kg dry                                 | 4.7                            | U         | 4.9            | U         | -4.2%                       | 4.4            | U | 4.7            | U | -6.6%                       |  |
|   | Aroclor 1248            | µg/Kg dry                                 | 4.5                            | U         | 4.7            | U         | -4.3%                       | 4.2            | U | 4.5            | U | -6.9%                       |  |
|   | Aroclor 1254            | µg/Kg dry                                 | 2.7                            | U         | 2.8            | U         | -3.6%                       | <b>71.5</b>    |   | <b>65</b>      |   | 9.5%                        |  |
|   | Aroclor 1260            | µg/Kg dry                                 | 5.4                            | U         | 5.7            | U         | -5.4%                       | <b>52.4</b>    |   | <b>53.2</b>    |   | -1.5%                       |  |
| <b>Congener</b>                               | PCB- 77                 | ng/Kg                                     | <b>186</b>                     |           | <b>171</b>     |           | 8.4%                        | <b>197</b>     |   | <b>228</b>     |   | -14.6%                      |  |
|   | PCB 81                  | ng/Kg                                     | <b>8.79</b>                    | J         | <b>6.48</b>    | J         | 30.3%                       | <b>5.58</b>    | J | <b>7.43</b>    | J | -28.4%                      |  |
|   | PCB 105                 | ng/Kg                                     | <b>1100</b>                    |           | <b>1180</b>    |           | -7.0%                       | <b>1240</b>    |   | <b>1390</b>    |   | -11.4%                      |  |
|   | PCB 114                 | ng/Kg                                     | <b>64.8</b>                    |           | <b>67.9</b>    |           | -4.7%                       | <b>72.1</b>    |   | <b>79.6</b>    |   | -9.9%                       |  |
|   | PCB 118                 | ng/Kg                                     | <b>2900</b>                    |           | <b>3250</b>    |           | -11.4%                      | <b>2990</b>    |   | <b>3410</b>    |   | -13.1%                      |  |
|   | PCB 123                 | ng/Kg                                     | <b>62.7</b>                    |           | <b>67</b>      |           | -6.6%                       | <b>64.1</b>    |   | <b>70.7</b>    |   | -9.8%                       |  |
|   | PCB 126                 | ng/Kg                                     | <b>22.5</b>                    |           | <b>13.1</b>    |           | 52.8%                       | <b>16.5</b>    |   | <b>13.9</b>    |   | 17.1%                       |  |
|   | PCB 156+157             | ng/Kg                                     | <b>553</b>                     |           | <b>609</b>     |           | -9.6%                       | <b>546</b>     |   | <b>653</b>     |   | -17.8%                      |  |
|   | PCB 167                 | ng/Kg                                     | <b>196</b>                     |           | <b>216</b>     |           | -9.7%                       | <b>197</b>     |   | <b>239</b>     |   | -19.3%                      |  |
|   | PCB 169                 | ng/Kg                                     | <b>4.27</b>                    | J         | <b>5.49</b>    | J         | -25.0%                      | <b>5.09</b>    | J | <b>6.46</b>    | J | -23.7%                      |  |
|   | PCB 189                 | ng/Kg                                     | <b>54.2</b>                    |           | <b>52.4</b>    |           | 3.4%                        | <b>55.1</b>    |   | <b>71.1</b>    |   | -25.4%                      |  |
| <b>PBDE</b>                                   | Total PBDE              | ug/kg                                     | <b>9.1</b>                     |           | <b>12.1</b>    |           | -28.7%                      | <b>4.9</b>     |   | <b>3.3</b>     |   | 38.6%                       |  |
| <b>TributylTin</b>                            | Tributyltin             | µg/Kg                                     | <b>20.8</b>                    |           | <b>28</b>      |           | -29.5%                      | <b>5</b>       |   | <b>5</b>       |   | 0.0%                        |  |
| <b>Bioassay</b>                               | Bioassay                | Toxic/not toxic                           |                                | Not Toxic |                | Not Toxic |                             | Toxic          |   | Toxic          |   |                             |  |
| <b>General</b>                                | Total Organic Carbon    | mg/Kg dry                                 | <b>13000</b>                   |           | <b>12000</b>   |           | 8.0%                        | <b>8800</b>    |   | <b>8600</b>    |   | 2.3%                        |  |
|   | Percent Solids          | %   | <b>86.02</b>                   |           | <b>85.51</b>   |           | 0.6%                        | <b>89.44</b>   |   | <b>89.92</b>   |   | -0.5%                       |  |
|   | Clay                    | %   | <b>4</b>                       |           | <b>2</b>       |           | 66.7%                       | <b>10</b>      |   | <b>7</b>       |   | 35.3%                       |  |
|   | Silt                    | %   | <b>10</b>                      |           | <b>9</b>       |           | 10.5%                       | <b>35</b>      |   | <b>35</b>      |   | 0.0%                        |  |
|   | Sand                    | %   | <b>45</b>                      |           | <b>49</b>      |           | -8.5%                       | <b>54</b>      |   | <b>53</b>      |   | 1.9%                        |  |
|   | Gravel                  | %   | <b>41</b>                      |           | <b>40</b>      |           | 2.5%                        | <b>1</b>       |   | <b>5</b>       |   | -133.3%                     |  |
|   | Start Date              | Date                                      | <b>9/18/09</b>                 |           | <b>9/18/09</b> |           |                             | <b>9/18/09</b> |   | <b>9/18/09</b> |   |                             |  |
|   | Stop Date               | Date                                      | <b>9/28/09</b>                 |           | <b>9/28/09</b> |           |                             | <b>9/28/09</b> |   | <b>9/28/09</b> |   |                             |  |
|   | Field Sample Depth      | m   | <b>0.1</b>                     |           | <b>0.1</b>     |           | 0.0%                        | <b>0.1</b>     |   | <b>0.1</b>     |   | 0.0%                        |  |
|   | Final Weight            | g   | <b>2903.5</b>                  |           | <b>3291</b>    |           | -12.5%                      | <b>2742</b>    |   | <b>2794</b>    |   | -1.9%                       |  |
|   | Initial Weight          | g   | <b>3375.4</b>                  |           | <b>3849</b>    |           | -13.1%                      | <b>3066</b>    |   | <b>3107</b>    |   | -1.3%                       |  |
|   | U                       | Under method detection limit              |                                |           |                |           |                             |                |   |                |   |                             |  |
|   | J                       | Estimate below Limit of Quantitation(LOQ) |                                |           |                |           |                             |                |   |                |   |                             |  |
|   |                         | Precision >20%                            |                                |           |                |           |                             |                |   |                |   |                             |  |

Appendix E Columbia Slough Triplicate QA/QC Review  
 Decision Area: Columbia Slough between River Mile 5.9 and 8.7.

| Columbia Slough OF 59-65     |                              | IS A         | IS B         | IS C         | Simple       | Standard Deviation | Relative Standard Deviation | 95% UCL Factor |         |
|------------------------------|------------------------------|--------------|--------------|--------------|--------------|--------------------|-----------------------------|----------------|---------|
|                              |                              | Sample       |              |              | Mean         |                    |                             |                |         |
| Sampling 2009                |                              |              |              |              |              |                    |                             |                |         |
| <b>Metals</b>                | Total Aluminum               | mg/Kg dry    | <b>17100</b> | <b>16100</b> | <b>17400</b> | <b>16867</b>       | 680.7                       | 4.0            | 18014.2 |
|                              | Total Antimony               | mg/Kg dry    | 0.5          | 0.5          | 0.5          | <b>0.5</b>         | 0.0                         | 0.0            | 0.5     |
|                              | Total Arsenic                | mg/Kg dry    | <b>5.2</b>   | <b>5.2</b>   | <b>5.6</b>   | <b>5.3</b>         | 0.2                         | 4.3            | 5.7     |
|                              | Total Barium                 | mg/Kg dry    | <b>122</b>   | <b>122</b>   | <b>129</b>   | <b>124.3</b>       | 4.0                         | 3.3            | 131.1   |
|                              | Total Cadmium                | mg/Kg dry    | <b>0.92</b>  | <b>0.84</b>  | <b>0.96</b>  | <b>0.9</b>         | 0.1                         | 6.7            | 1.0     |
|                              | Total Chromium               | mg/Kg dry    | <b>45.4</b>  | <b>41.5</b>  | <b>45.5</b>  | <b>44.1</b>        | 2.3                         | 5.2            | 48.0    |
|                              | Total Cobalt                 | mg/Kg dry    | <b>9.7</b>   | <b>9.54</b>  | <b>10.2</b>  | <b>9.8</b>         | 0.3                         | 3.5            | 10.4    |
|                              | Total Copper                 | mg/Kg dry    | <b>37.4</b>  | <b>34.2</b>  | <b>41</b>    | <b>37.5</b>        | 3.4                         | 9.1            | 43.3    |
|                              | Total Lead                   | mg/Kg dry    | <b>40.8</b>  | <b>37.6</b>  | <b>45.5</b>  | <b>41.3</b>        | 4.0                         | 9.6            | 48.0    |
|                              | Total Manganese              | mg/Kg dry    | <b>409</b>   | <b>408</b>   | <b>436</b>   | <b>417.7</b>       | 15.9                        | 3.8            | 444.4   |
|                              | Mercury                      | mg/Kg dry    | <b>0.1</b>   | <b>0.09</b>  | <b>0.11</b>  | <b>0.1</b>         | 0.0                         | 10.0           | 0.1     |
|                              | Total Nickel                 | mg/Kg dry    | <b>19.8</b>  | <b>18.7</b>  | <b>20.7</b>  | <b>19.7</b>        | 1.0                         | 5.1            | 21.4    |
|                              | Total Selenium               | mg/Kg dry    | 1            | 1            | 1            | <b>1.0</b>         | 0.0                         | 0.0            | 1.0     |
|                              | Total Silver                 | mg/Kg dry    | <b>0.19</b>  | <b>0.49</b>  | <b>0.25</b>  | <b>0.3</b>         | 0.2                         | 51.2           | 0.6     |
| Total Thallium               | mg/Kg dry                    | 0.2          | 0.2          | 0.2          | <b>0.2</b>   | 0.0                | 0.0                         | 0.2            |         |
| Total Zinc                   | mg/Kg dry                    | <b>244</b>   | <b>227</b>   | <b>261</b>   | <b>244.0</b> | 17.0               | 7.0                         | 272.7          |         |
| <b>Semi-Vols</b>             | 1,2,4,5-Tetrachlorobenzene   | µg/Kg dry    | 227          | 240          | 281          | <b>249.3</b>       | 28.2                        | 11.3           | 296.8   |
|                              | 1,2,4-Trichlorobenzene       | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 1,2-Dichlorobenzene          | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 1,3-Dichlorobenzene          | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 1,4-Dichlorobenzene          | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,3,4,6-Tetrachlorophenol    | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,3,5,6-Tetrachlorophenol    | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,4,5-Trichlorophenol        | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,4,6-Trichlorophenol        | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,4-Dichlorophenol           | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,4-Dimethylphenol           | µg/Kg dry    | 130          | 130          | 130          | 130.0              | 0.0                         | 0.0            | 130.0   |
|                              | 2,4-Dinitrophenol            | µg/Kg dry    | Void         | Void         | Void         | Void               | Void                        | Void           | Void    |
|                              | 2,4-Dinitrotoluene           | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,6-Dichlorophenol           | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2,6-Dinitrotoluene           | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2-Chloronaphthalene          | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2-Chlorophenol               | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2-Methylphenol               | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 2-Nitrophenol                | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 4,6-Dinitro-2-methylphenol   | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 4-Bromophenyl phenyl ether   | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 4-Chloro-3-methylphenol      | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | 4-Methylphenol (p-Cresol)    | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | Acenaphthene                 | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | Acenaphthylene               | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | Anthracene                   | µg/Kg dry    | 66           | 66           | 66           | 66.0               | 0.0                         | 0.0            | 66.0    |
|                              | Benzo[a]anthracene           | µg/Kg dry    | <b>94.9</b>  | <b>119.6</b> | <b>124.8</b> | <b>113.1</b>       | 16.0                        | 14.1           | 140.0   |
|                              | Benzo[a]pyrene               | µg/Kg dry    | <b>79.9</b>  | <b>119.6</b> | <b>129.8</b> | <b>109.8</b>       | 26.4                        | 24.0           | 154.2   |
|                              | Benzo[b]fluoranthene         | µg/Kg dry    | <b>194.7</b> | <b>254.2</b> | <b>274.6</b> | <b>241.2</b>       | 41.5                        | 17.2           | 311.2   |
|                              | Benzo[g,h,i]perylene         | µg/Kg dry    | <b>114.8</b> | <b>154.5</b> | <b>159.8</b> | <b>143.0</b>       | 24.6                        | 17.2           | 184.5   |
|                              | Benzo[k]fluoranthene         | µg/Kg dry    | 66           | 66           | <b>69.9</b>  | <b>67.3</b>        | 2.3                         | 3.3            | 71.1    |
|                              | Bis(2-Chloroethoxy) methane  | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Bis(2-Chloroethyl) ether     | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Bis(2-Chloroisopropyl) ether | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Bis(2-ethylhexyl)adipate     | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Bis(2-ethylhexyl)phthalate   | µg/Kg dry    | <b>713.9</b> | <b>942.2</b> | <b>1113</b>  | <b>923.1</b>       | 200.4                       | 21.7           | 1260.9  |
|                              | Butylbenzylphthalate         | µg/Kg dry    | 99.8         | 99.8         | 99.8         | <b>99.8</b>        | 0.0                         | 0.0            | 99.8    |
|                              | Chrysene                     | µg/Kg dry    | <b>109.8</b> | <b>164.5</b> | <b>174.7</b> | <b>149.7</b>       | 34.9                        | 23.3           | 208.5   |
|                              | di-n-Butylphthalate          | µg/Kg dry    | 99.8         | 99.8         | 99.8         | <b>99.8</b>        | 0.0                         | 0.0            | 99.8    |
|                              | Di-n-octylphthalate          | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Dibenz[a,h]anthracene        | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Dibenzofuran                 | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Diethylphthalate             | µg/Kg dry    | 99.8         | 99.8         | 99.8         | <b>99.8</b>        | 0.0                         | 0.0            | 99.8    |
|                              | Dimethylphthalate            | µg/Kg dry    | 99.8         | 99.8         | 99.8         | <b>99.8</b>        | 0.0                         | 0.0            | 99.8    |
|                              | Fluoranthene                 | µg/Kg dry    | <b>234.6</b> | <b>319</b>   | <b>354.5</b> | <b>302.7</b>       | 61.6                        | 20.3           | 406.5   |
|                              | Fluorene                     | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Hexachloro-1,3-Butadiene     | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Hexachlorocyclopentadiene    | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Hexachloroethane             | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
|                              | Indeno[1,2,3-cd]pyrene       | µg/Kg dry    | <b>74.9</b>  | <b>109.7</b> | <b>119.8</b> | <b>101.5</b>       | 23.6                        | 23.2           | 141.2   |
|                              | Isophorone                   | µg/Kg dry    | 66           | 66           | 66           | <b>66.0</b>        | 0.0                         | 0.0            | 66.0    |
| n-Nitroso-di-n-dipropylamine | µg/Kg dry                    | 66           | 66           | 66           | <b>66.0</b>  | 0.0                | 0.0                         | 66.0           |         |
| n-Nitrosodiphenylamine       | µg/Kg dry                    | 66           | 66           | 66           | <b>66.0</b>  | 0.0                | 0.0                         | 66.0           |         |
| Naphthalene                  | µg/Kg dry                    | 66           | 66           | 66           | <b>66.0</b>  | 0.0                | 0.0                         | 66.0           |         |
| Nitrobenzene                 | µg/Kg dry                    | 66           | 66           | 66           | <b>66.0</b>  | 0.0                | 0.0                         | 66.0           |         |
| Pentachlorophenol            | µg/Kg dry                    | 66           | 66           | 66           | <b>66.0</b>  | 0.0                | 0.0                         | 66.0           |         |
| Phenanthrene                 | µg/Kg dry                    | 66           | <b>114.7</b> | <b>119.8</b> | <b>100.2</b> | 29.7               | 29.6                        | 150.2          |         |
| Phenol                       | µg/Kg dry                    | 66           | 66           | 66           | <b>66.0</b>  | 0.0                | 0.0                         | 66.0           |         |
| Pyrene                       | µg/Kg dry                    | <b>234.6</b> | <b>324</b>   | <b>339.5</b> | <b>299.4</b> | 56.6               | 18.9                        | 394.8          |         |

Appendix E Columbia Slough Triplicate QA/QC Review  
 Decision Area: Columbia Slough between River Mile 5.9 and 8.7.

| Columbia Slough OF 59-65    |                          | IS A      | IS B         | IS C         | Simple Mean  | Standard Deviation | Relative Standard Deviation | 95% UCL Factor |         |
|-----------------------------|--------------------------|-----------|--------------|--------------|--------------|--------------------|-----------------------------|----------------|---------|
|                             |                          |           |              |              |              |                    |                             |                | Sample  |
| Sampling 2009               |                          |           |              |              |              |                    |                             |                |         |
| <b>Pesticides</b>           | Aldrin                   | ug/kg     | 8.02         | 8.04         | 7.87         | <b>8.0</b>         | 0.1                         | 1.2            | 8.1     |
| Method                      | alpha-BHC                | ug/kg     | 8.71         | 8.73         | 8.55         | <b>8.7</b>         | 0.1                         | 1.1            | 8.8     |
| 8081                        | beta-BHC                 | ug/kg     | 10.2         | 10.2         | 10           | <b>10.1</b>        | 0.1                         | 1.1            | 10.3    |
|                             | delta-BHC                | ug/kg     | 7.12         | 7.13         | 6.98         | <b>7.1</b>         | 0.1                         | 1.2            | 7.2     |
|                             | gamma-BHC (Lindane)      | ug/kg     | 12           | 12           | 11.8         | <b>11.9</b>        | 0.1                         | 1.0            | 12.1    |
|                             | Chlordane                | ug/kg     | 45.8         | 45.9         | 44.9         | <b>45.5</b>        | 0.6                         | 1.2            | 46.5    |
|                             | alpha-Chlordane          | ug/kg     | 11.4         | 11.4         | 11.2         | <b>11.3</b>        | 0.1                         | 1.0            | 11.5    |
|                             | gamma-Chlordane          | ug/kg     | 12.2         | 12.2         | 12           | <b>12.1</b>        | 0.1                         | 1.0            | 12.3    |
|                             | 4,4'-DDD                 | ug/kg     | 18.7         | 18.7         | 18.3         | <b>18.6</b>        | 0.2                         | 1.2            | 19.0    |
|                             | 4,4'-DDE                 | ug/kg     | 5.4          | 5.41         | 5.29         | <b>5.4</b>         | 0.1                         | 1.2            | 5.5     |
|                             | 4,4'-DDT                 | ug/kg     | 40           | 40.1         | 39.3         | <b>39.8</b>        | 0.4                         | 1.1            | 40.5    |
|                             | Dieldrin                 | ug/kg     | 7.14         | 7.15         | 7            | <b>7.1</b>         | 0.1                         | 1.2            | 7.2     |
|                             | Endosulfan I             | ug/kg     | 10.2         | 10.2         | 9.97         | <b>10.1</b>        | 0.1                         | 1.3            | 10.3    |
|                             | Endosulfan II            | ug/kg     | 8.84         | 8.86         | 8.67         | <b>8.8</b>         | 0.1                         | 1.2            | 9.0     |
|                             | Endosulfan sulfate       | ug/kg     | 11.5         | 11.5         | 11.3         | <b>11.4</b>        | 0.1                         | 1.0            | 11.6    |
|                             | Endrin                   | ug/kg     | 9.62         | 9.64         | 9.43         | <b>9.6</b>         | 0.1                         | 1.2            | 9.8     |
|                             | Endrin aldehyde          | ug/kg     | 9.93         | 9.95         | 9.74         | <b>9.9</b>         | 0.1                         | 1.2            | 10.1    |
|                             | Endrin ketone            | ug/kg     | 23.7         | 23.8         | 23.3         | <b>23.6</b>        | 0.3                         | 1.1            | 24.0    |
|                             | Heptachlor               | ug/kg     | 8.77         | 8.79         | 8.61         | <b>8.7</b>         | 0.1                         | 1.1            | 8.9     |
|                             | Heptachlor epoxide       | ug/kg     | 8.63         | 8.65         | 8.47         | <b>8.6</b>         | 0.1                         | 1.1            | 8.7     |
|                             | Methoxychlor             | ug/kg     | 22.3         | 22.4         | 21.9         | <b>22.2</b>        | 0.3                         | 1.2            | 22.6    |
|                             | Toxaphene                | ug/kg     | 568          | 569          | 557          | <b>564.7</b>       | 6.7                         | 1.2            | 575.9   |
| <b>Pesticides</b>           | 2,4'-DDD                 |           | <b>0.818</b> | <b>1.070</b> | <b>0.836</b> | <b>0.91</b>        | 0.14                        | 15.5           | 1.1     |
| Method                      | 2,4'-DDE                 |           | <b>0.271</b> | <b>0.381</b> | <b>0.291</b> | <b>0.31</b>        | 0.06                        | 18.6           | 0.4     |
| 1669                        | 2,4'-DDT                 |           | <b>0.066</b> | <b>0.121</b> | <b>0.172</b> | <b>0.12</b>        | 0.05                        | <b>44.2</b>    | 0.2     |
|                             | 4,4'-DDD                 |           | <b>3.140</b> | <b>4.430</b> | <b>3.520</b> | <b>3.70</b>        | 0.66                        | 17.9           | 4.8     |
|                             | 4,4'-DDE                 |           | <b>5.930</b> | <b>6.860</b> | <b>6.140</b> | <b>6.31</b>        | 0.49                        | 7.7            | 7.1     |
|                             | 4,4'-DDT                 |           | <b>0.405</b> | <b>0.480</b> | <b>0.530</b> | <b>0.47</b>        | 0.06                        | 13.3           | 0.6     |
|                             | 4,4'-Methoxychlor        |           | 0.439        | 0.628        | 0.570        | <b>0.55</b>        | 0.10                        | 17.7           | 0.7     |
|                             | Aldrin                   |           | <b>0.208</b> | 0.172        | <b>0.172</b> | <b>0.18</b>        | 0.02                        | 11.3           | 0.2     |
|                             | alpha-BHC                |           | <b>0.019</b> | 0.017        | 0.017        | <b>0.02</b>        | 0.00                        | 8.5            | 0.0     |
|                             | beta-BHC                 |           | <b>0.022</b> | 0.017        | 0.017        | <b>0.02</b>        | 0.00                        | 15.5           | 0.0     |
|                             | Lindane (gamma-BHC)      |           | 0.029        | 0.029        | 0.029        | <b>0.03</b>        | 0.00                        | 0.0            | 0.0     |
|                             | delta-BHC                |           | 0.017        | 0.017        | 0.017        | <b>0.02</b>        | 0.00                        | 0.0            | 0.0     |
|                             | cis-Chlordane (alpha)    |           | <b>0.950</b> | <b>1.000</b> | <b>1.090</b> | <b>1.01</b>        | 0.07                        | 7.0            | 1.1     |
|                             | trans-Chlordane (gamma)  |           | <b>0.927</b> | <b>1.060</b> | <b>0.931</b> | <b>0.97</b>        | 0.08                        | 7.8            | 1.1     |
|                             | Oxchlordane              |           | 0.023        | 0.064        | 0.063        | <b>0.05</b>        | 0.02                        | 46.4           | 0.1     |
|                             | cis-Heptachlor Epoxide   |           | 0.050        | 0.051        | 0.059        | <b>0.05</b>        | 0.01                        | 9.4            | 0.1     |
|                             | cis-Nonachlor            |           | <b>0.346</b> | <b>0.295</b> | <b>0.318</b> | <b>0.32</b>        | 0.03                        | 8.0            | 0.4     |
|                             | Dieldrin                 |           | <b>0.644</b> | <b>0.763</b> | <b>0.694</b> | <b>0.70</b>        | 0.06                        | 8.5            | 0.8     |
|                             | Endosulfan I (alpha)     |           | 0.064        | 0.064        | 0.089        | <b>0.07</b>        | 0.01                        | 19.8           | 0.1     |
|                             | Endosulfan II (beta)     |           | 0.102        | 0.174        |              | <b>0.14</b>        | 0.05                        | 36.9           | 0.2     |
|                             | Endosulfan Sulfate       |           | <b>0.375</b> | <b>0.281</b> | 0.485        | <b>0.38</b>        | 0.10                        | 26.8           | 0.6     |
|                             | Endrin                   |           | 0.087        | 0.127        | 0.239        | <b>0.15</b>        | 0.08                        | 52.1           | 0.3     |
|                             | Endrin Aldehyde          |           | 0.226        | 0.226        | 0.407        | <b>0.29</b>        | 0.10                        | 36.5           | 0.5     |
|                             | Endrin Ketone            |           | 0.226        | 0.402        | 0.439        | <b>0.36</b>        | 0.11                        | 32.0           | 0.5     |
|                             | Heptachlor               |           | 0.082        | 0.082        | 0.082        | <b>0.08</b>        | 0.00                        | 0.0            | 0.1     |
|                             | Hexachlorobenzene        |           | <b>0.212</b> | <b>0.192</b> | <b>0.170</b> | <b>0.37</b>        | 0.29                        | <b>79.0</b>    | 0.9     |
|                             | Hexachlorobutadiene      |           | 0.017        | 0.017        | 0.017        | <b>0.02</b>        | 0.00                        | 0.0            | 0.0     |
|                             | Mirex                    |           | 0.070        | 0.121        | 0.135        | <b>0.11</b>        | 0.03                        | 31.3           | 0.2     |
|                             | trans-Heptachlor Epoxide |           | 0.087        | 0.087        | 0.242        | <b>0.14</b>        | 0.09                        | 64.7           | 0.3     |
|                             | trans-Nonachlor          |           | <b>0.712</b> | <b>0.700</b> | <b>0.721</b> | <b>0.71</b>        | 0.01                        | 1.5            | 0.7     |
| <b>TBT</b>                  | Tributyltin              | µg/Kg dry | 6            | <b>11</b>    | <b>51.7</b>  | <b>22.9</b>        | 25.1                        | <b>109.5</b>   | 65.2    |
| <b>PCBs</b>                 | Aroclor-1016 (PCB-1016)  | µg/Kg dry | 4.9          | 4.8          | 12.2         | 7.3                | 4.2                         | 58.1           | 14.5    |
|                             | Aroclor-1242 (PCB-1242)  | µg/Kg dry | 2.4          | 2.4          | 6.1          | 3.6                | 2.1                         | 58.8           | 7.2     |
|                             | Aroclor 1221             | µg/Kg dry | 3.4          | 3.4          | 8.5          | 5.1                | 2.9                         | 57.7           | 10.1    |
|                             | Aroclor 1232             | µg/Kg dry | 4.5          | 4.5          | 11.3         | 6.8                | 3.9                         | 58.0           | 13.4    |
|                             | Aroclor 1248             | µg/Kg dry | 4.3          | 4.3          | <b>10.8</b>  | 6.5                | 3.8                         | 58.0           | 12.8    |
|                             | Aroclor 1254             | µg/Kg dry | <b>27.3</b>  | <b>16.5</b>  | <b>36.6</b>  | <b>26.8</b>        | 10.1                        | <b>37.5</b>    | 43.8    |
|                             | Aroclor 1260             | µg/Kg dry | <b>20.2</b>  | <b>13.4</b>  | <b>22.5</b>  | <b>18.7</b>        | 4.7                         | 25.3           | 26.7    |
| <b>Dioxin-like Congener</b> | PCB- 77                  | ng/Kg     | <b>89</b>    | <b>79.9</b>  | <b>119</b>   | <b>96.0</b>        | 20.5                        | 21.3           | 130.5   |
|                             | PCB 81                   | ng/Kg     | <b>2.77</b>  | <b>2.71</b>  | 4.59         | 3.4                | 1.1                         | <b>31.8</b>    | 5.2     |
|                             | PCB 105                  | ng/Kg     | <b>540</b>   | <b>465</b>   | <b>671</b>   | <b>558.7</b>       | 104.3                       | 18.7           | 734.4   |
|                             | PCB 114                  | ng/Kg     | <b>32.4</b>  | <b>28.3</b>  | <b>41.3</b>  | <b>34.0</b>        | 6.6                         | 19.5           | 45.2    |
|                             | PCB 118                  | ng/Kg     | <b>1390</b>  | <b>1180</b>  | <b>1690</b>  | <b>1420.0</b>      | 256.3                       | 18.1           | 1852.1  |
|                             | PCB 123                  | ng/Kg     | <b>30</b>    | <b>23.2</b>  | <b>29.1</b>  | <b>27.4</b>        | 3.7                         | 13.5           | 33.7    |
|                             | PCB 126                  | ng/Kg     | <b>9.94</b>  | <b>5.77</b>  | <b>9.29</b>  | <b>8.3</b>         | 2.2                         | 26.9           | 12.1    |
|                             | PCB 156 + 157            | ng/Kg     | <b>279</b>   | <b>241</b>   | <b>351</b>   | <b>290.3</b>       | 55.9                        | 19.2           | 384.5   |
|                             | PCB 167                  | ng/Kg     | <b>101</b>   | <b>89.4</b>  | <b>126</b>   | <b>105.5</b>       | 18.7                        | 17.7           | 137.0   |
|                             | PCB 169                  | ng/Kg     | 3.54         | 3.51         | 6.9          | 4.7                | 1.9                         | 41.9           | 7.9     |
|                             | PCB 189                  | ng/Kg     | <b>28.8</b>  | <b>23.3</b>  | <b>39.8</b>  | <b>30.6</b>        | 8.4                         | 27.4           | 44.8    |
|                             | Total PCBs               | ng/Kg     | <b>37800</b> | <b>33500</b> | <b>42400</b> | <b>37900</b>       | 4450.8                      | 11.7           | 45403.5 |
| <b>PBDE</b>                 | BDE-(28/33)              | ng/Kg     | <b>160.0</b> | <b>154.0</b> |              | 157.0              | 4.2                         | 2.7            | 164.2   |
|                             | BDE-47                   | ng/Kg     | <b>1830</b>  | <b>2020</b>  |              | 1925.0             | 134.4                       | 7.0            | 2151.5  |
|                             | BDE-99                   | ng/Kg     | <b>2180</b>  | <b>5030</b>  |              | 3605.0             | 2015.3                      | 55.9           | 7002.4  |
|                             | BDE-100                  | ng/Kg     | <b>538</b>   | <b>1140</b>  |              | 839.0              | 425.7                       | 50.7           | 1556.6  |
|                             | BDE-153                  | ng/Kg     | <b>349</b>   | <b>898</b>   |              | 623.5              | 388.2                       | 62.3           | 1278.0  |
|                             | BDE-154                  | ng/Kg     | <b>270</b>   | <b>697</b>   |              | 483.5              | 301.9                       | 62.4           | 992.5   |
|                             | BDE-183                  | ng/Kg     | <b>166</b>   | <b>201</b>   |              | 183.5              | 24.7                        | 13.5           | 225.2   |
|                             | BDE-209                  | ng/Kg     | <b>4420</b>  | <b>3430</b>  |              | 3925.0             | 700.0                       | 17.8           | 5105.2  |
|                             | Total BDE                | ng/Kg     | <b>9913</b>  | <b>13570</b> |              | 11741.5            | 2585.9                      | 22.0           | 16101.0 |



Appendix F -- 2009 Lower Columbia Slough Sediment between River Mile 5.9 and 8.7.

Congener Data Table

| Lab ID #                         | 35980        | 35976  | 35974  | 35974          | 35971          | 35969  | 35968   | 35967          | 35967          | 35965  | 35966  | 35962  | 35981           | 35981             | 35981             |        |
|----------------------------------|--------------|--------|--------|----------------|----------------|--------|---------|----------------|----------------|--------|--------|--------|-----------------|-------------------|-------------------|--------|
| Site Name                        | OF 59 #59    | OF #60 | OF #61 | OF #61         | OF #61A        | OF62   | Wastech | Prec +BR       | Prec +BR       | OF 63  | OF 64  | OF 65  | IS Triplicate A | IS (Triplicate B) | IS (Triplicate C) |        |
| Parameter                        | Unit         | Sample | Sample | Sample - Field | Sample - Field | Sample | Sample  | Sample - Field | Sample - Field | Sample | Sample | Sample | Sample          | Sample            | Sample            |        |
| Total PCBs                       | ug/kg dry wt | 48.5   | 60.2   | 74.2           | 82.2           | 54     | 904     | 95.1           | 88.1           | 103    | 50.8   | 42.1   | 17.5            | 37.8              | 33.5              | 42.4   |
| Total Deca PCBs                  | ng/Kg dry wt | 197    | 221    | 200            | 353            | 80.9   | 321     | 337            | 192            | 207    | 187    | 115    | 53.9            | 151               | 148               | 205    |
| Total Di PCBs                    | ng/Kg dry wt | 628    | 1030   | 657            | 715            | 865    | 27400   | 641            | 625            | 855    | 688    | 729    | 248             | 443               | 408               | 1170   |
| Total Hepta PCBs                 | ng/Kg dry wt | 5170   | 5410   | 8570           | 7660           | 5520   | 40100   | 8290           | 10000          | 12800  | 6200   | 6140   | 2100            | 4420              | 4060              | 5270   |
| Total Hexa PCBs                  | ng/Kg dry wt | 12400  | 15700  | 21000          | 23700          | 12600  | 99300   | 28500          | 24800          | 30200  | 15500  | 12500  | 5290            | 11100             | 10000             | 11800  |
| Total Mono PCBs                  | ng/Kg dry wt | 62.0   | 74.4   | 98.4           | 140            | 80.9   | 523     | 64.8           | 66.3           | 78.6   | 90.2   | 97.4   | 30.5            | 56.9              | 60.5              | 196    |
| Total Nona PCBs                  | ng/Kg dry wt | 421    | 397    | 381            | 464            | 242    | 1720    | 403            | 452            | 496    | 370    | 397    | 118             | 314               | 299               | 404    |
| Total Octa PCBs                  | ng/Kg dry wt | 1700   | 1740   | 2400           | 2210           | 1480   | 10300   | 2050           | 2640           | 3250   | 1760   | 1710   | 562             | 1320              | 1320              | 1940   |
| Total PCB                        | ng/Kg dry wt | 48500  | 60200  | 74200          | 82200          | 54000  | 904000  | 95100          | 88100          | 103000 | 50800  | 42100  | 17500           | 37800             | 33500             | 42400  |
| Total Penta PCBs                 | ng/Kg dry wt | 13700  | 16900  | 24000          | 28400          | 15200  | 142000  | 34000          | 29600          | 31600  | 16400  | 10900  | 5920            | 11600             | 9740              | 10900  |
| Total Tetra PCBs                 | ng/Kg dry wt | 10400  | 12900  | 12900          | 14300          | 12500  | 387000  | 17200          | 14100          | 17100  | 7410   | 7110   | 2480            | 6300              | 5660              | 7210   |
| Total Tri PCBs                   | ng/Kg dry wt | 3870   | 5920   | 4020           | 4360           | 5460   | 196000  | 3630           | 5520           | 6800   | 2270   | 2320   | 677             | 2080              | 1810              | 3310   |
| PCB-1                            | ng/Kg dry wt | 17.4   | 23.9   | 27.7           | 33             | 20.6   | 307     | 21.8           | 21.8           | 27.7   | 26.2   | 35.3   | 13.8            | 15.4              | 18.3              | 52.8   |
| PCB-10                           | ng/Kg dry wt | 3.17 U | 3.36 J | 3.34 U         | 2.42 J         | 3.21 U | 272     | 3.72 J         | 5.17 J         | 5.58 J | 3.59 U | 4.39 J | 3.06 U          | 3.2 U             | 3.17 U            | 6.24 U |
| PCB-100 C-E 93+98+100+102        | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-101 C-E 90+101+113           | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-102 C-E 93+98+100+102        | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-103                          | ng/Kg dry wt | 34     | 29.1   | 56.1           | 69.1           | 22.5   | 262     | 429            | 38             | 37.4   | 22     | 16.5   | 8.01 J          | 21.1              | 19.3              | 16.9   |
| PCB-104                          | ng/Kg dry wt | 1.86 U | 1.87 U | 1.96 U         | 0.362 U        | 1.88 U | 1.88 U  | 1.92 U         | 1.87 U         | 1.85 U | 2.11 U | 2.29 U | 1.8 U           | 1.88 U            | 1.86 U            | 3.66 U |
| PCB-105                          | ng/Kg dry wt | 646    | 824    | 1100           | 1180           | 705    | 9500    | 820            | 1240           | 1390   | 908    | 606    | 346             | 540               | 465               | 671    |
| PCB-106                          | ng/Kg dry wt | 4.26 U | 4.29 U | 4.48 U         | 0.828 U        | 4.31 U | 4.3 U   | 4.39 U         | 4.29 U         | 4.22 U | 4.83 U | 5.23 U | 4.11 U          | 4.29 U            | 4.26 U            | 8.38 U |
| PCB-107 C-E 107+123              | ng/Kg dry wt | 67.3   | 84.5   | 111            | 130            | 69.8   | 942     | 98.5           | 128            | 148    | 97.3   | 68.2   | 34.6            | 59.8              | 52.4              | 64.7   |
| PCB-108 C-E 86+87+97+108+119+125 | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-109                          | ng/Kg dry wt | 117    | 136    | 200            | 248            | 93.3   | 1240    | 162            | 204            | 224    | 164    | 104    | 54.4            | 96.8              | 86.5              | 127    |
| PCB-11                           | ng/Kg dry wt | 215    | 234    | 183            | 210            | 267    | 541     | 137            | 127            | 144    | 275    | 251    | 130             | 172               | 162               | 692    |
| PCB-110 C-E 110+115              | ng/Kg dry wt | 2350   | 3040   | 4230           | 5090           | 2690   | 15400   | 4980           | 5870           | 5670   | 2890   | 1880   | 1080            | 1970              | 1660              | 1920   |
| PCB-111                          | ng/Kg dry wt | 2.24 U | 2.25 U | 4.76 J         | 0.434 U        | 2.26 U | 2.25 U  | 12.7           | 2.25 U         | 2.22 U | 2.53 U | 2.74 U | 2.16 U          | 2.25 U            | 2.23 U            | 4.4 U  |
| PCB-112                          | ng/Kg dry wt | 2.82 U | 2.84 U | 2.96 U         | 0.548 U        | 2.85 U | 2.84 U  | 2.91 U         | 2.83 U         | 2.79 U | 3.19 U | 3.46 U | 2.72 U          | 2.84 U            | 2.82 U            | 5.54 U |
| PCB-113 C-E 90+101+113           | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-114                          | ng/Kg dry wt | 41     | 46.8   | 64.8           | 67.9           | 41.5   | 613     | 46.8           | 72.1           | 79.6   | 50.4   | 34.6   | 15.8            | 32.4              | 28.3              | 41.3   |
| PCB-115 C-E 110+115              | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-116 C-E 85+116+117           | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-117 C-E 85+116+117           | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-118                          | ng/Kg dry wt | 1540   | 1950   | 2900           | 3250           | 1620   | 20300   | 2090           | 2990           | 3410   | 2290   | 1480   | 858             | 1390              | 1180              | 1690   |
| PCB-119 C-E 86+87+97+108+119+125 | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-12 C-E 12+13                 | ng/Kg dry wt | 38.3   | 60.2   | 40.8           | 46.3           | 41     | 730     | 32.2           | 33.1           | 44.1   | 35.5   | 31.1   | 16.5 J          | 29.7              | 25.9              | 66     |
| PCB-120                          | ng/Kg dry wt | 7.61 J | 9.84 J | 33.9           | 29.3           | 2.56 U | 8.88 J  | 33.9           | 16.4           | 12.6   | 8.95 J | 5.33 J | 3.29 J          | 6.32 J            | 6.29 J            | 5.13 J |
| PCB-121                          | ng/Kg dry wt | 2.52 U | 2.54 U | 2.65 U         | 0.49 U         | 2.55 U | 2.54 U  | 23.8           | 2.53 U         | 2.5 U  | 2.85 U | 3.09 U | 2.43 U          | 2.54 U            | 2.52 U            | 4.95 U |
| PCB-122                          | ng/Kg dry wt | 23.8   | 27.6   | 34.8           | 41.4           | 26.1   | 270     | 31.7           | 37.3           | 46.3   | 31.4   | 19.6   | 9.13 J          | 19.6              | 18.1              | 22.2   |
| PCB-123                          | ng/Kg dry wt | 35.7   | 42.9   | 62.7           | 67             | 26.4 J | 477     | 50.8           | 64.1           | 70.7   | 40.5   | 33.4   | 17.3            | 30                | 23.2              | 29.1   |
| PCB-124 C-E 107+123              | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-125 C-E 86+87+97+108+119+125 | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-126                          | ng/Kg dry wt | 7.5 J  | 11.6   | 22.5           | 13.1           | 37.4   | 257     | 5.46 J         | 16.5           | 13.9   | 16.9   | 7.47 J | 2.76 J          | 9.94 J            | 5.77 J            | 9.29 J |
| PCB-127                          | ng/Kg dry wt | 2.36 U | 2.83 J | 6.64 J         | 6.07           | 2.39 U | 34.6    | 11.5           | 5.29 J         | 6.89 J | 4.42 J | 3.19 J | 2.28 U          | 2.38 U            | 2.36 U            | 4.65 U |
| PCB-128 C-E 128+166              | ng/Kg dry wt | 342    | 454    | 678            | 749            | 390    | 3800    | 635            | 733            | 837    | 537    | 399    | 195             | 336               | 288               | 410    |
| PCB-129 C-E 129+138+163          | ng/Kg dry wt | 2580   | 3360   | 4700           | 5200           | 2660   | 26700   | 4870           | 5390           | 6410   | 3690   | 2860   | 1280            | 2430              | 2150              | 2910   |
| PCB-13 C-E 12+13                 | ng/Kg dry wt | C-E    | C-E    | C-E            | C-E            | C-E    | C-E     | C-E            | C-E            | C-E    | C-E    | C-E    | C-E             | C-E               | C-E               | C-E    |
| PCB-130                          | ng/Kg dry wt | 169    | 214    | 304            | 359            | 167    | 1780    | 306            | 337            | 378    | 240    | 170    | 85.3            | 165               | 144               | 176    |
| PCB-131                          | ng/Kg dry wt | 34.2   | 54.1   | 67.6           | 87.2           | 39.2   | 528     | 70.8           | 79.1           | 94.8   | 55.4   | 36.4   | 20.4            | 35.9              | 32.2              | 38.7   |
| PCB-132                          | ng/Kg dry wt | 911    | 1190   | 1540           | 1880           | 978    | 9000    | 1740           | 1730           | 2190   | 1180   | 899    | 405             | 829               | 754               | 811    |
| PCB-133                          | ng/Kg dry wt | 47.4   | 51.9   | 80.2           | 86.2           | 43.7   | 319     | 174            | 76.3           | 88.9   | 56.4   | 42.3   | 17.8            | 39.5              | 36.9              | 44.4   |

Appendix F -- 2009 Lower Columbia Slough Sediment between River Mile 5.9 and 8.7.

Congener Data Table

| Lab ID #                | 35980         | 35976         | 35974          | 35974          | 35971         | 35969         | 35968         | 35967          | 35967          | 35965         | 35966         | 35962         | 35981           | 35981             | 35981             |        |
|-------------------------|---------------|---------------|----------------|----------------|---------------|---------------|---------------|----------------|----------------|---------------|---------------|---------------|-----------------|-------------------|-------------------|--------|
| Site Name               | OF 59 #59     | OF #60        | OF #61         | OF #61         | OF #61A       | OF62          | Wastech       | Prec +BR       | Prec +BR       | OF 63         | OF 64         | OF 65         | IS Triplicate A | IS (Triplicate B) | IS (Triplicate C) |        |
| Parameter               | 09/15/2009 09 | 09/14/2009 10 | 09/14/2009 09  | 09/14/2009 09  | 09/09/2009 10 | 09/09/2009 09 | 09/10/2009 10 | 09/10/2009 09  | 09/10/2009 09  | 09/08/2009 11 | 09/08/2009 10 | 09/08/2009 09 | 09/18/2009      | 09/18/2009        | 09/18/2009        |        |
| Unit                    | Sample        | Sample        | Sample - Field | Sample - Field | Sample        | Sample        | Sample        | Sample - Field | Sample - Field | Sample        | Sample        | Sample        | Sample          | Sample            | Sample            |        |
| PCB-134 C-E 134+143     | ng/Kg dry wt  | 159           | 199            | 288            | 345           | 184           | 1690          | 369            | 325            | 379           | 205           | 156           | 68              | 147               | 125               | 146    |
| PCB-135 C-E 135+151     | ng/Kg dry wt  | 1100          | 1350           | 1660           | 1820          | 1130          | 4900          | 3400           | 2380           | 2790          | 1070          | 971           | 348             | 896               | 843               | 863    |
| PCB-136                 | ng/Kg dry wt  | 442           | 526            | 661            | 787           | 471           | 2170          | 1590           | 929            | 1080          | 437           | 371           | 145             | 367               | 332               | 303    |
| PCB-137                 | ng/Kg dry wt  | 129           | 135            | 224            | 265           | 114           | 1470          | 206            | 194            | 260           | 155           | 99.5          | 61.4            | 104               | 108               | 109    |
| PCB-138 C-E 129+138+163 | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-139 C-E 139+140     | ng/Kg dry wt  | 50.2          | 61             | 89             | 105           | 46.2          | 561           | 135            | 88.8           | 99.3          | 64.6          | 42            | 21.8            | 44.8              | 39.7              | 46.5   |
| PCB-14                  | ng/Kg dry wt  | 3.83 U        | 3.86 U         | 4.03 U         | 0.745 U       | 3.88 U        | 3.86 U        | 3.95 U         | 3.86 U         | 3.8 U         | 4.34 U        | 4.71 U        | 3.7 U           | 3.86 U            | 3.83 U            | 7.54 U |
| PCB-140 C-E 139+140     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-141                 | ng/Kg dry wt  | 460           | 616            | 796            | 877           | 520           | 5060          | 822            | 1050           | 1350          | 597           | 509           | 204             | 431               | 387               | 477    |
| PCB-142                 | ng/Kg dry wt  | 2.14 U        | 2.15 U         | 2.25 U         | 0.415 U       | 2.16 U        | 2.15 U        | 2.2 U          | 2.15 U         | 2.12 U        | 2.42 U        | 2.62 U        | 2.06 U          | 2.15 U            | 2.14 U            | 4.2 U  |
| PCB-143 C-E 134+143     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-144                 | ng/Kg dry wt  | 146           | 199            | 238            | 262           | 177           | 699           | 295            | 343            | 436           | 164           | 147           | 55.3            | 131               | 118               | 111    |
| PCB-145                 | ng/Kg dry wt  | 2.32 U        | 2.34 U         | 2.45 U         | 0.452 U       | 2.35 U        | 2.34 U        | 2.4 U          | 2.34 U         | 2.3 U         | 2.63 U        | 2.85          | 2.24 U          | 2.34 U            | 2.32 U            | 4.57 U |
| PCB-146                 | ng/Kg dry wt  | 409           | 486            | 715            | 779           | 372           | 3300          | 1070           | 726            | 864           | 489           | 402           | 165             | 358               | 326               | 389    |
| PCB-147 C-E 147+149     | ng/Kg dry wt  | 2340          | 2930           | 3690           | 4320          | 2340          | 8640          | 6990           | 4500           | 5630          | 2600          | 2250          | 862             | 2010              | 1840              | 1830   |
| PCB-148                 | ng/Kg dry wt  | 13.9          | 2.71 U         | 13.9           | 11.4          | 2.72 U        | 2.71 U        | 90.4           | 2.71 U         | 2.67 U        | 3.05 U        | 3.31 U        | 2.6 U           | 2.71 U            | 2.69 U            | 5.3 U  |
| PCB-149 C-E 147+149     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-15                  | ng/Kg dry wt  | 199           | 411            | 222            | 231           | 265           | 6970          | 184            | 217            | 299           | 174           | 150           | 48.4            | 124               | 114               | 188    |
| PCB-150                 | ng/Kg dry wt  | 9.93 J        | 2.09 U         | 14.3           | 13.8          | 2.1 U         | 19            | 76.6           | 7.11 J         | 9.07 J        | 2.35 U        | 2.55 U        | 2 U             | 2.09 U            | 2.08 U            | 4.08 U |
| PCB-151 C-E 135+151     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-152                 | ng/Kg dry wt  | 5.48 J        | 2.21 U         | 2.31 U         | 6.01          | 2.22 U        | 26.7          | 2.26 U         | 4.12 J         | 7.57 J        | 2.49          | 2.7 U         | 2.12            | 2.21 U            | 2.19 U            | 4.32 U |
| PCB-153 C-E 153+168     | ng/Kg dry wt  | 2170          | 2740           | 3650           | 4030          | 2030          | 19900         | 4000           | 4160           | 5250          | 2670          | 2230          | 909             | 1950              | 1790              | 2190   |
| PCB-154                 | ng/Kg dry wt  | 52.8          | 46.7           | 91.9           | 88.2          | 31.1          | 63.6          | 361            | 53.7           | 52.2          | 34.2          | 26.2          | 11.4            | 30.9              | 26.2              | 28.6   |
| PCB-155                 | ng/Kg dry wt  | 2.04 U        | 2.06 U         | 2.15 U         | 0.397 U       | 2.06 U        | 2.06 U        | 2.1 U          | 2.05 U         | 2.02 U        | 2.31 U        | 2.51 U        | 1.97 U          | 2.06 U            | 2.04 U            | 4.01 U |
| PCB-156 C-E 156+157     | ng/Kg dry wt  | 293           | 358            | 553            | 609           | 304           | 3280          | 413            | 546            | 653           | 460           | 317           | 166             | 279               | 241               | 351    |
| PCB-157 C-E 156+157     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-158                 | ng/Kg dry wt  | 215           | 292            | 429            | 462           | 257           | 2570          | 385            | 496            | 588           | 329           | 254           | 118             | 203               | 179               | 251    |
| PCB-159                 | ng/Kg dry wt  | 30.4          | 35.1           | 8.94 J         | 42.6          | 23.3          | 8.23 J        | 44.7           | 45.4           | 64.7          | 39.4          | 39.3          | 13.1            | 26                | 25.6              | 8.01 J |
| PCB-16                  | ng/Kg dry wt  | 118           | 182            | 138            | 133           | 253           | 14100         | 153            | 338            | 405           | 75.8          | 69.2          | 22.7            | 63.4              | 52.3              | 125    |
| PCB-160                 | ng/Kg dry wt  | 4.63 U        | 4.66 U         | 4.87 U         | 0.9 U         | 4.68 U        | 4.67 U        | 4.77 U         | 4.66 U         | 4.59 U        | 5.24 U        | 5.68 U        | 4.47 U          | 4.67 U            | 4.63 U            | 9.11 U |
| PCB-161                 | ng/Kg dry wt  | 2.22 U        | 2.24 U         | 2.34 U         | 0.432 U       | 2.25 U        | 2.24 U        | 2.29 U         | 2.23 U         | 2.2 U         | 2.51 U        | 2.73 U        | 2.14 U          | 2.24 U            | 2.22 U            | 4.37 U |
| PCB-162                 | ng/Kg dry wt  | 10            | 12.9           | 14             | 18.3          | 4.6 J         | 183           | 9.4 J          | 10.2           | 50.7          | 34.7          | 7.97 J        | 4.22 U          | 10.3              | 5.52 J            | 9.58 J |
| PCB-163 C-E 129+138+163 | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-164                 | ng/Kg dry wt  | 149           | 217            | 293            | 286           | 182           | 1600          | 275            | 373            | 402           | 231           | 192           | 78              | 147               | 121               | 197    |
| PCB-165                 | ng/Kg dry wt  | 1.89 U        | 1.9 U          | 2.68 J         | 0.368 U       | 1.91 U        | 1.91 U        | 19.3           | 1.9 U          | 1.88 U        | 2.14 U        | 2.32 U        | 1.83 U          | 1.91 U            | 1.89 U            | 3.72 U |
| PCB-166 C-E 128+166     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-167                 | ng/Kg dry wt  | 105           | 123            | 196            | 216           | 103           | 1090          | 167            | 197            | 239           | 166           | 126           | 59.7            | 101               | 89.4              | 126    |
| PCB-168 C-E 153+168     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-169                 | ng/Kg dry wt  | 3.92 J        | 3.53 U         | 4.27 J         | 5.49 J        | 3.55 U        | 16.8 J        | 4.54 J         | 5.09 J         | 6.46 J        | 5.52 J        | 4.31 U        | 3.39 U          | 3.54 U            | 3.51 U            | 6.9 U  |
| PCB-17                  | ng/Kg dry wt  | 155           | 241            | 170            | 176           | 260           | 14500         | 173            | 305            | 379           | 98.3          | 104           | 27.4            | 80.6              | 67.9              | 159    |
| PCB-170                 | ng/Kg dry wt  | 514           | 600            | 1070           | 868           | 579           | 4720          | 887            | 1110           | 1430          | 736           | 718           | 253             | 476               | 422               | 766    |
| PCB-171 C-E 171+173     | ng/Kg dry wt  | 183           | 220            | 330            | 302           | 208           | 1650          | 290            | 373            | 473           | 240           | 234           | 84.2            | 161               | 147               | 182    |
| PCB-172                 | ng/Kg dry wt  | 103           | 126            | 187            | 160           | 117           | 919           | 162            | 213            | 270           | 144           | 141           | 48.5            | 99                | 87.6              | 126    |
| PCB-173 C-E 171+173     | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-174                 | ng/Kg dry wt  | 638           | 756            | 947            | 929           | 683           | 5170          | 934            | 1200           | 1550          | 774           | 752           | 259             | 577               | 516               | 547    |
| PCB-175                 | ng/Kg dry wt  | 32.1          | 37.1           | 49.6           | 46.2          | 41.6          | 231           | 47             | 61.9           | 78.4          | 34.9          | 36.9          | 12.8            | 27.2              | 27.7              | 28.6   |
| PCB-176                 | ng/Kg dry wt  | 102           | 123            | 150            | 154           | 118           | 811           | 173            | 208            | 261           | 105           | 102           | 34.7            | 88.1              | 84.4              | 83.1   |
| PCB-177                 | ng/Kg dry wt  | 371           | 439            | 605            | 549           | 396           | 2890          | 604            | 696            | 887           | 460           | 468           | 154             | 319               | 283               | 353    |
| PCB-178                 | ng/Kg dry wt  | 169           | 178            | 236            | 219           | 165           | 1150          | 303            | 294            | 360           | 169           | 158           | 53.6            | 130               | 126               | 132    |
| PCB-179                 | ng/Kg dry wt  | 374           | 413            | 490            | 494           | 387           | 2230          | 665            | 641            | 826           | 333           | 323           | 109             | 290               | 274               | 260    |
| PCB-18 C-E 18+30        | ng/Kg dry wt  | 283           | 436            | 308            | 314           | 502           | 29700         | 343            | 683            | 821           | 176           | 158           | 53.7 JH         | 147               | 125               | 284    |
| PCB-180 C-E 180+193     | ng/Kg dry wt  | 1210          | 1380           | 2260           | 1830          | 1330          | 9630          | 1880           | 2500           | 3250          | 1580          | 1590          | 536             | 1040              | 964               | 1490   |

Appendix F -- 2009 Lower Columbia Slough Sediment between River Mile 5.9 and 8.7.

Congener Data Table

| Parameter           | Unit         | 35980<br>OF 59 #59<br>09/15/2009 09 | 35976<br>OF #60<br>09/14/2009 10 | 35974<br>OF #61<br>09/14/2009 09 | 35974<br>OF #61<br>09/14/2009 09 | 35971<br>OF #61A<br>09/09/2009 10 | 35969<br>OF62<br>09/09/2009 09 | 35968<br>Wastech<br>09/10/2009 10 | 35967<br>Prec +BR<br>09/10/2009 09 | 35967<br>Prec +BR<br>09/10/2009 09 | 35965<br>OF 63<br>09/08/2009 11 | 35966<br>OF 64<br>09/08/2009 10 | 35962<br>OF 65<br>09/08/2009 09 | 35981<br>IS Triplicate A<br>09/18/2009 | 35981<br>IS (Triplicate B)<br>09/18/2009 | 35981<br>IS (Triplicate C)<br>09/18/2009 |
|---------------------|--------------|-------------------------------------|----------------------------------|----------------------------------|----------------------------------|-----------------------------------|--------------------------------|-----------------------------------|------------------------------------|------------------------------------|---------------------------------|---------------------------------|---------------------------------|--|--|--|
| PCB-181             | ng/Kg dry wt | 2.04 U                              | 2.05 U                           | 10.4 J                           | 11.2                             | 2.06 U                            | 44.5                           | 8.73 J                            | 11.1                               | 10.7                               | 6.89 J                          | 2.5 U                           | 1.96 U                          | 2.05 U                                 | 2.04 U                                   | 4 U                                      |
| PCB-182             | ng/Kg dry wt | 1.93 U                              | 1.94 U                           | 8.65 J                           | 7.77                             | 1.95 U                            | 1.94 U                         | 1.99 U                            | 1.94 U                             | 1.91 U                             | 2.18 U                          | 2.37 U                          | 1.86 U                          | 1.94 U                                 | 1.93 U                                   | 3.79 U                                   |
| PCB-183 C-E 183+185 | ng/Kg dry wt | 484                                 | 690 J                            | 689                              | 652                              | 472                               | 3600                           | 672                               | 875                                | 1110                               | 529                             | 525                             | 183                             | 378                                    | 361                                      | 406                                      |
| PCB-184             | ng/Kg dry wt | 1.81 U                              | 1.82 U                           | 1.91 U                           | 0.352 U                          | 1.83 U                            | 1.83 U                         | 1.87 U                            | 1.82 U                             | 1.8 U                              | 2.05 U                          | 2.22 U                          | 1.75                            | 1.83 U                                 | 1.81 U                                   | 3.56 U                                   |
| PCB-185 C-E 183+185 | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-186             | ng/Kg dry wt | 1.84 U                              | 1.85 U                           | 1.93 U                           | 0.357 U                          | 1.86 U                            | 1.85 U                         | 1.89 U                            | 1.85 U                             | 1.82 U                             | 2.08 U                          | 2.26 U                          | 1.77 U                          | 1.85 U                                 | 1.84 U                                   | 3.61 U                                   |
| PCB-187             | ng/Kg dry wt | 836                                 | 958                              | 1250                             | 1170                             | 852                               | 5700                           | 1430                              | 1540                               | 1890                               | 879                             | 882                             | 295                             | 690                                    | 640                                      | 677                                      |
| PCB-188             | ng/Kg dry wt | 4.82 J                              | 2.57 U                           | 2.69 U                           | 0.496 U                          | 2.58 U                            | 2.57 U                         | 2.63 U                            | 2.57 U                             | 2.53 U                             | 2.89 U                          | 3.14 U                          | 2.46 U                          | 2.57 U                                 | 2.55 U                                   | 5.02 U                                   |
| PCB-189             | ng/Kg dry wt | 31.7                                | 35.6                             | 54.2                             | 52.4                             | 28.9                              | 211                            | 40.6                              | 55.1                               | 71.1                               | 44.2                            | 37.5                            | 14.2                            | 28.8                                   | 23.3                                     | 39.8                                     |
| PCB-19              | ng/Kg dry wt | 40.7                                | 57                               | 41.3                             | 46.9                             | 77.8                              | 2710                           | 57.4                              | 96.4                               | 106                                | 41.3                            | 49.4                            | 11                              | 17.3                                   | 18.6                                     | 27.1                                     |
| PCB-190             | ng/Kg dry wt | 99.7                                | 119                              | 201                              | 171                              | 123                               | 1030                           | 160                               | 204                                | 312                                | 144                             | 141                             | 49.9                            | 88.4                                   | 85.1                                     | 146                                      |
| PCB-191             | ng/Kg dry wt | 20                                  | 25.1                             | 45.3                             | 37.6                             | 25.8                              | 157                            | 36.7                              | 45                                 | 60.5                               | 28.1                            | 30.4                            | 10.5                            | 20.8                                   | 17.6                                     | 29.8                                     |
| PCB-192             | ng/Kg dry wt | 3.69 U                              | 3.72 U                           | 3.89 U                           | 0.718 U                          | 3.74 U                            | 3.72 U                         | 3.81 U                            | 3.71 U                             | 3.66 U                             | 4.18 U                          | 4.53 U                          | 3.56 U                          | 3.72 U                                 | 3.69 U                                   | 7.27 U                                   |
| PCB-193 C-E 180+193 | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-194             | ng/Kg dry wt | 381                                 | 371                              | 571                              | 502                              | 303                               | 2680                           | 441                               | 606                                | 785                                | 384                             | 379                             | 123                             | 285                                    | 284                                      | 452                                      |
| PCB-195             | ng/Kg dry wt | 140                                 | 143                              | 206                              | 188                              | 113                               | 893                            | 175                               | 215                                | 291                                | 148                             | 143                             | 46.8                            | 111                                    | 111                                      | 168                                      |
| PCB-196             | ng/Kg dry wt | 199                                 | 210                              | 294                              | 267                              | 186                               | 1050                           | 248                               | 325                                | 400                                | 217                             | 200                             | 66.6                            | 158                                    | 164                                      | 242                                      |
| PCB-197 C-E 197+200 | ng/Kg dry wt | 68.1                                | 69.8                             | 92.3                             | 84.1                             | 68.5                              | 468                            | 85.3                              | 115                                | 134                                | 72.8                            | 69.5                            | 23                              | 54                                     | 54.5                                     | 65.2                                     |
| PCB-198 C-E 198+199 | ng/Kg dry wt | 430                                 | 456                              | 610                              | 571                              | 417                               | 2770                           | 548                               | 684                                | 813                                | 464                             | 455                             | 148                             | 349                                    | 355                                      | 507                                      |
| PCB-199 C-E 198+199 | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-2               | ng/Kg dry wt | 21.3 JH                             | 23.3 JH                          | 30.4 JH                          | 55.1 JH                          | 30.9 JH                           | 49.5 JH                        | 20.9 JH                           | 19.1 JH                            | 21.4 JH                            | 31.8 JH                         | 31.3 JH                         | 16.7 JH                         | 20.2 JH                                | 20.5 JH                                  | 83.7 JH                                  |
| PCB-20 C-E 20+28    | ng/Kg dry wt | 999                                 | 1520                             | 1010                             | 1100                             | 1250                              | 35400                          | 753                               | 1120                               | 1420                               | 553                             | 565                             | 165                             | 524                                    | 469                                      | 810                                      |
| PCB-200 C-E 197+200 | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-201             | ng/Kg dry wt | 65.6                                | 61.5                             | 75.1                             | 78.3                             | 59.4                              | 412                            | 73.2                              | 98.7                               | 113                                | 60.6                            | 58.4                            | 20.5                            | 49.7                                   | 47.5                                     | 54.6                                     |
| PCB-202             | ng/Kg dry wt | 113                                 | 121                              | 138                              | 136                              | 90                                | 655                            | 129                               | 157                                | 183                                | 109                             | 111                             | 38.1                            | 82.1                                   | 82                                       | 110                                      |
| PCB-203             | ng/Kg dry wt | 277                                 | 284                              | 373                              | 348                              | 240                               | 1200                           | 327                               | 405                                | 483                                | 277                             | 266                             | 86.8                            | 211                                    | 205                                      | 317                                      |
| PCB-204             | ng/Kg dry wt | 1.77 U                              | 1.78 U                           | 1.86 U                           | 0.344 U                          | 1.79 U                            | 1.79 U                         | 1.83 U                            | 1.78 U                             | 1.76 U                             | 2.01 U                          | 2.18 U                          | 1.71 U                          | 1.79 U                                 | 1.77 U                                   | 3.49 U                                   |
| PCB-205             | ng/Kg dry wt | 24.4                                | 26.7                             | 37.3                             | 32.2                             | 5.1 U                             | 147                            | 28.3                              | 38.1                               | 48                                 | 25.9                            | 27.3                            | 8.98 J                          | 21.6                                   | 17.3                                     | 25.1                                     |
| PCB-206             | ng/Kg dry wt | 291                                 | 270                              | 245                              | 297                              | 183                               | 1170                           | 270                               | 319                                | 357                                | 253                             | 274                             | 79.6                            | 216                                    | 212                                      | 287                                      |
| PCB-207             | ng/Kg dry wt | 35.1                                | 37.6                             | 34.2                             | 39.2                             | 4.41                              | 206                            | 36.1                              | 41.4                               | 43.5                               | 36.8                            | 34.2                            | 11.6                            | 26.4                                   | 25.1                                     | 35.9                                     |
| PCB-208             | ng/Kg dry wt | 95.1                                | 88.8                             | 101                              | 128                              | 59.3                              | 342                            | 96.2                              | 91.9                               | 95.1                               | 80.6                            | 88.1                            | 27.3                            | 71.8                                   | 62.1                                     | 80.9                                     |
| PCB-209             | ng/Kg dry wt | 197                                 | 221                              | 200                              | 353                              | 80.9                              | 321                            | 337                               | 192                                | 207                                | 187                             | 115                             | 53.9                            | 151                                    | 148                                      | 205                                      |
| PCB-21 C-E 21+33    | ng/Kg dry wt | 369                                 | 588                              | 398                              | 435                              | 573                               | 20100                          | 390                               | 549                                | 675                                | 203                             | 222                             | 65.5                            | 217                                    | 189                                      | 326                                      |
| PCB-22              | ng/Kg dry wt | 304                                 | 502                              | 298                              | 337                              | 433                               | 12300                          | 258                               | 378                                | 465                                | 171                             | 174                             | 50.2                            | 162                                    | 144                                      | 237                                      |
| PCB-23              | ng/Kg dry wt | 2.95 U                              | 2.97 U                           | 3.11 U                           | 0.574 U                          | 2.99 U                            | 55.1                           | 3.04 U                            | 2.97 U                             | 2.93 U                             | 3.34 U                          | 3.62 U                          | 2.85 U                          | 2.97 U                                 | 2.95 U                                   | 5.81 U                                   |
| PCB-24              | ng/Kg dry wt | 5.96 J                              | 9.53 J                           | 5.52 J                           | 5.36                             | 9.32 J                            | 538                            | 6.37 J                            | 9.32 J                             | 11.6                               | 4.63 J                          | 3.78 U                          | 2.97 U                          | 3.18 J                                 | 3.08 U                                   | 6.18 J                                   |
| PCB-25              | ng/Kg dry wt | 82.1                                | 124                              | 98.5                             | 105                              | 103                               | 2990                           | 105                               | 81.9                               | 102                                | 51.2                            | 43.1                            | 16.3                            | 44.2                                   | 37.7                                     | 69.2                                     |
| PCB-26 C-E 26+29    | ng/Kg dry wt | 173                                 | 253                              | 186                              | 205                              | 213                               | 7420                           | 273                               | 201                                | 247                                | 104                             | 95.1                            | 30                              | 92.3                                   | 79                                       | 153                                      |
| PCB-27              | ng/Kg dry wt | 41.5                                | 51.5                             | 41.6                             | 45.2                             | 61.3                              | 2550                           | 35.4                              | 65.7                               | 80.1                               | 27.4                            | 36.6                            | 6.98 J                          | 19.5                                   | 14.6                                     | 37.8                                     |
| PCB-28 C-E 20+28    | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-29 C-E 26+29    | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-3               | ng/Kg dry wt | 23.3                                | 27.2                             | 40.3                             | 52.3                             | 29.3                              | 166 J                          | 22.1                              | 25.4                               | 29.5                               | 32.2                            | 30.8                            | 15.9 U                          | 21.3                                   | 21.8                                     | 59.9                                     |
| PCB-30 C-E 18+30    | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-31              | ng/Kg dry wt | 737                                 | 1120                             | 739                              | 852                              | 982                               | 33900                          | 659                               | 994                                | 1260                               | 381                             | 432                             | 121                             | 398                                    | 347                                      | 640                                      |
| PCB-32              | ng/Kg dry wt | 164                                 | 213                              | 150                              | 175                              | 262                               | 8980                           | 153                               | 260                                | 300                                | 93.4                            | 95.9                            | 25.1 JH                         | 74.8                                   | 63.8                                     | 128                                      |
| PCB-33 C-E 21+33    | ng/Kg dry wt | C-E                                 | C-E                              | C-E                              | C-E                              | C-E                               | C-E                            | C-E                               | C-E                                | C-E                                | C-E                             | C-E                             | C-E                             | C-E                                    | C-E                                      | C-E                                      |
| PCB-34              | ng/Kg dry wt | 7.36 J                              | 9.18 J                           | 7.12 J                           | 6.77                             | 3.75 U                            | 152                            | 4.54 J                            | 7.09 J                             | 11.1                               | 4.19 U                          | 4.55 U                          | 3.57 U                          | 3.79 J                                 | 3.7 U                                    | 7.28 U                                   |
| PCB-35              | ng/Kg dry wt | 25.1                                | 37.5                             | 27.6                             | 31.4                             | 30.4                              | 491                            | 19.2                              | 28.4                               | 30.8                               | 21.7                            | 22.8                            | 7.31 J                          | 16.4                                   | 15.5                                     | 21                                       |
| PCB-36              | ng/Kg dry wt | 2.31 U                              | 2.33 U                           | 2.78 J                           | 0.45 U                           | 2.34 U                            | 2.33 U                         | 2.39 U                            | 2.33 U                             | 2.29 U                             | 2.62 U                          | 2.84 U                          | 2.23 U                          | 2.33 U                                 | 2.31 U                                   | 4.55 U                                   |
| PCB-37              | ng/Kg dry wt | 360                                 | 564                              | 381                              | 383                              | 435                               | 9710                           | 246                               | 402                                | 469                                | 257                             | 246                             | 75.3                            | 207                                    | 179                                      | 279                                      |
| PCB-38              | ng/Kg dry wt | 3.1 U                               | 3.12 U                           | 3.26 U                           | 0.603 U                          | 3.14 U                            | 3.13 U                         | 3.2 U                             | 3.12 U                             | 3.07 U                             | 3.51 U                          | 3.81 U                          | 2.99 U                          | 3.13 U                                 | 3.1 U                                    | 6.1 U                                    |
| PCB-39              | ng/Kg dry wt | 8.15 J                              | 10.6                             | 10.5 J                           | 10.2                             | 7.85 J                            | 132                            | 5.57 J                            | 8.94 J                             | 13.9                               | 5.73 J                          | 5.42 J                          | 2.06 U                          | 5.06 J                                 | 3.57 J                                   | 6.77 J                                   |

Appendix F -- 2009 Lower Columbia Slough Sediment between River Mile 5.9 and 8.7.

Congener Data Table

| Lab ID #                                     | 35980         | 35976         | 35974          | 35974          | 35971         | 35969         | 35968         | 35967          | 35967          | 35965         | 35966         | 35962         | 35981           | 35981             | 35981             |
|--|---------------|---------------|----------------|----------------|---------------|---------------|---------------|----------------|----------------|---------------|---------------|---------------|-----------------|-------------------|-------------------|
| Site Name                                    | OF 59 #59     | OF #60        | OF #61         | OF #61         | OF #61A       | OF62          | Wastech       | Prec +BR       | Prec +BR       | OF 63         | OF 64         | OF 65         | IS Triplicate A | IS (Triplicate B) | IS (Triplicate C) |
| Parameter                                    | 09/15/2009 09 | 09/14/2009 10 | 09/14/2009 09  | 09/14/2009 09  | 09/09/2009 10 | 09/09/2009 09 | 09/10/2009 10 | 09/10/2009 09  | 09/10/2009 09  | 09/08/2009 11 | 09/08/2009 10 | 09/08/2009 09 | 09/18/2009      | 09/18/2009        | 09/18/2009        |
| Unit   | Sample        | Sample        | Sample - Field | Sample - Field | Sample        | Sample        | Sample        | Sample - Field | Sample - Field | Sample        | Sample        | Sample        | Sample          | Sample            | Sample            |
| PCB-4 ng/Kg dry wt                           | 32.6          | 67.4          | 37.6           | 43.5           | 64.7          | 4510          | 75.1          | 76.1           | 97.1           | 66.9          | 163           | 11.5          | 23.1            | 19                | 44.1              |
| PCB-40 C-E 40+41+71 ng/Kg dry wt             | 704           | 868           | 731            | 790            | 918           | 33000         | 573           | 912            | 1050           | 423           | 483           | 124           | 337             | 344               | 410               |
| PCB-41 C-E 40+41+71 ng/Kg dry wt             | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-42 ng/Kg dry wt                          | 363           | 448           | 410            | 461            | 452           | 14800         | 286           | 478            | 519            | 220           | 242           | 69.1          | 191             | 180               | 257               |
| PCB-43 C-E 43+73 ng/Kg dry wt                | 40            | 43.5          | 41.7           | 40             | 64            | 2380          | 30.5          | 51.3           | 58.9           | 14.5          | 16            | 4.13 U        | 22.2            | 10.1              | 31.2              |
| PCB-44 C-E 44+47+65 ng/Kg dry wt             | 1260          | 1580          | 1580           | 1730           | 1580          | 54200         | 1460          | 1760           | 2200           | 862           | 844           | 284           | 739             | 660               | 927               |
| PCB-45 C-E 45+51 ng/Kg dry wt                | 280           | 322           | 271            | 285            | 423           | 15300         | 310           | 373            | 474            | 137           | 139           | 36.4          | 126             | 115               | 187               |
| PCB-46 ng/Kg dry wt                          | 90.1          | 114           | 99.9           | 96.5           | 166           | 5470          | 113           | 145            | 178            | 49.6          | 51.5          | 14.4          | 44.3            | 39.4              | 64.5              |
| PCB-47 C-E 44+47+65 ng/Kg dry wt             | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-48 ng/Kg dry wt                          | 217           | 295           | 241            | 256            | 314           | 14900         | 173           | 300            | 372            | 120           | 161           | 38.1          | 127             | 115               | 152               |
| PCB-49 C-E 49+69 ng/Kg dry wt                | 831           | 1060          | 1180           | 1280           | 1030          | 37700         | 2600          | 1120           | 1330           | 551           | 529           | 187           | 509             | 468               | 545               |
| PCB-5 ng/Kg dry wt                           | 4.43 U        | 4.46 U        | 4.66 U         | 4.09           | 4.48 U        | 140           | 4.57 U        | 4.46 U         | 5.08 J         | 5.02 U        | 5.44 U        | 4.28 U        | 4.46 U          | 4.43 U            | 8.71 U            |
| PCB-50 C-E 50+53 ng/Kg dry wt                | 206           | 243           | 212            | 235            | 309           | 10800         | 271           | 295            | 365            | 122           | 119           | 32.4          | 98.2            | 86.4              | 143               |
| PCB-51 C-E 45+51 ng/Kg dry wt                | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-52 ng/Kg dry wt                          | 1750          | 2280          | 2460           | 2890           | 2210          | 82200         | 7020          | 2850           | 3610           | 1300          | 1080          | 472           | 1210            | 1030              | 1620              |
| PCB-53 C-E 50+53 ng/Kg dry wt                | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-54 ng/Kg dry wt                          | 3.46 J        | 3.11 U        | 3.97 J         | 3.55           | 3.13 U        | 73.6          | 3.64 J        | 5 J            | 4.19 J         | 3.5 U         | 3.8 U         | 2.98 U        | 3.12 U          | 3.09 U            | 6.08 U            |
| PCB-55 ng/Kg dry wt                          | 4.17 U        | 4.2 U         | 4.39 U         | 0.811 U        | 22.1          | 4.21 U        | 4.3 U         | 4.2 U          | 4.13 U         | 4.72 U        | 5.12 U        | 4.03 U        | 4.2 U           | 4.17 U            | 13.5 J            |
| PCB-56 ng/Kg dry wt                          | 529           | 596           | 525            | 570            | 517           | 10400         | 357           | 616            | 695            | 364           | 358           | 120           | 292             | 266               | 265               |
| PCB-57 ng/Kg dry wt                          | 4.19 J        | 9.79 J        | 6.37 J         | 7.59           | 6.71 J        | 187           | 38.6          | 3.35 J         | 5.25 J         | 5.37 J        | 3.62 J        | 1.71 U        | 4.66 J          | 3.83 J            | 6.62 J            |
| PCB-58 ng/Kg dry wt                          | 5.39 J        | 6.7 J         | 8.2 J          | 11.7 J         | 4.5 J         | 66.5 J        | 232           | 6.05 J         | 4.67 J         | 3.51 J        | 3.99 J        | 2.68 U        | 4.16 J          | 2.8 J             | 5.45 U            |
| PCB-59 C-E 59+62+75 ng/Kg dry wt             | 130           | 161           | 141            | 153            | 171           | 4840          | 113           | 147            | 176            | 79.4          | 87.3          | 22 J          | 66.4            | 62.3              | 85.3              |
| PCB-6 ng/Kg dry wt                           | 28.4          | 52.1          | 39.2           | 40.3           | 46.5          | 2320          | 42.5          | 33.9           | 46.2           | 27.6          | 26.1          | 11.2          | 19.9            | 17.4              | 47.4              |
| PCB-60 ng/Kg dry wt                          | 184           | 235           | 187            | 190            | 240           | 5390          | 142           | 260            | 289            | 144           | 170           | 45.6          | 97.3            | 85.3              | 111               |
| PCB-61 C-E 61+70+74+76 ng/Kg dry wt          | 1760          | 2220          | 2340           | 2640           | 1970          | 47100         | 1750          | 2470           | 3010           | 1510          | 1400          | 544           | 1230            | 1100              | 1220              |
| PCB-62 C-E 59+62+75 ng/Kg dry wt             | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-63 ng/Kg dry wt                          | 48.6          | 57.1          | 52.6           | 62.7           | 54.7          | 1240          | 34.7          | 47.9           | 57.2           | 33.1          | 38.1          | 12            | 29.9            | 26.2              | 26.3              |
| PCB-64 ng/Kg dry wt                          | 565           | 680           | 626            | 663            | 732           | 22500         | 467           | 713            | 826            | 374           | 406           | 118           | 302             | 279               | 351               |
| PCB-65 C-E 44+47+65 ng/Kg dry wt             | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-66 ng/Kg dry wt                          | 1190          | 1380          | 1400           | 1550           | 1070          | 20400         | 863           | 1320           | 1560           | 885           | 808           | 293           | 712             | 656               | 626               |
| PCB-67 ng/Kg dry wt                          | 38.1          | 50.6          | 49.3           | 55.9           | 43.6          | 904           | 68.9          | 35.2           | 42.4           | 27.2          | 32.1          | 9.71          | 24.9            | 23.2              | 24.7              |
| PCB-68 ng/Kg dry wt                          | 11.3          | 16.2          | 24.1           | 29.9           | 8.61 J        | 93.6          | 34.2          | 11.5           | 10.9           | 12.3          | 8.17 J        | 4.41 J        | 10.5            | 8.4 J             | 9.74 J            |
| PCB-69 C-E 49+69 ng/Kg dry wt                | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-7 ng/Kg dry wt                           | 5.06 J        | 7.81 J        | 6.77 J         | 7.79           | 9.69 J        | 464           | 8.92 J        | 6.64 J         | 9.76 J         | 7.36 J        | 5.79 J        | 3.25 U        | 4.76 J          | 4.21 J            | 9.49 J            |
| PCB-70 C-E 61+70+74+76 ng/Kg dry wt          | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-71 C-E 40+41+71 ng/Kg dry wt             | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-72 ng/Kg dry wt                          | 20.6          | 25.5          | 54.5           | 70.7           | 16            | 194           | 73.9          | 20.4           | 18.4           | 16.5          | 12 J          | 4.95 J        | 14.3            | 14.0              | 12 J              |
| PCB-73 C-E 43+73 ng/Kg dry wt                | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-74 C-E 61+70+74+76 ng/Kg dry wt          | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-75 C-E 59+62+75 ng/Kg dry wt             | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-76 C-E 61+70+74+76 ng/Kg dry wt          | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E             | C-E               | C-E               |
| PCB-77 ng/Kg dry wt                          | 158           | 189           | 186            | 171            | 157           | 2470 J        | 95.2          | 197            | 228            | 133           | 111           | 40.9          | 89              | 79.9              | 119               |
| PCB-78 ng/Kg dry wt                          | 2.64 U        | 2.66 U        | 2.78 U         | 0.513 U        | 2.67 U        | 2.66 U        | 2.72 U        | 2.65 U         | 2.62 U         | 2.99 U        | 3.24 U        | 2.55 U        | 2.66 U          | 2.64 U            | 5.19 U            |
| PCB-79 ng/Kg dry wt                          | 12.5          | 15.6          | 33.5           | 37             | 13.5          | 115           | 26.2          | 25.8           | 26.5           | 19.4          | 13.2          | 7.9 J         | 12.3            | 9.74 J            | 16.6 J            |
| PCB-8 ng/Kg dry wt                           | 102           | 176           | 118            | 122            | 167           | 10700         | 144           | 133            | 190            | 91.5          | 98.7          | 30.7 JH       | 63.1 JH         | 59.1 JH           | 126               |
| PCB-80 ng/Kg dry wt                          | 2.56 U        | 2.58 U        | 2.7 U          | 0.498 U        | 2.59 U        | 2.58 U        | 43.4          | 13.2           | 16.9           | 2.9 U         | 3.15 U        | 2.47 U        | 2.58 U          | 2.56              | 5.04 U            |
| PCB-81 ng/Kg dry wt                          | 3.84 J        | 5.48 J        | 8.79 J         | 6.48           | 6.26 J        | 116           | 4.44 J        | 5.58 J         | 7.43 J         | 5.76 J        | 7.1 J         | 2.25 U        | 2.77 J          | 2.71 J            | 4.59 U            |
| PCB-82 ng/Kg dry wt                          | 264           | 308           | 389            | 448            | 305           | 1590          | 381           | 617            | 627            | 286           | 190           | 99.8          | 191             | 161               | 197               |
| PCB-83 ng/Kg dry wt                          | 144           | 177           | 174            | 239            | 119           | 912           | 344           | 256            | 317            | 204           | 114           | 53.2          | 106             | 66.7              | 94.1              |
| PCB-84 ng/Kg dry wt                          | 631           | 806           | 1070           | 1330           | 769           | 8220          | 1270          | 1340           | 1510           | 710           | 426           | 254           | 523             | 438               | 483               |
| PCB-85 C-E 85+116+117 ng/Kg dry wt           | 331           | 414           | 527            | 622            | 368           | 2740          | 537           | 778            | 787            | 388           | 275           | 142           | 259             | 214               | 246               |
| PCB-86 C-E 86+87+97+108+119+125 ng/Kg dry wt | 1380          | 1730          | 2440           | 2750           | 1590          | 11900         | 2460          | 3390           | 3480           | 1600          | 1090          | 588           | 1200            | 1010              | 1040              |

Appendix F -- 2009 Lower Columbia Slough Sediment between River Mile 5.9 and 8.7.

Congener Data Table

| Lab ID #                        | 35980         | 35976         | 35974          | 35974          | 35971         | 35969         | 35968         | 35967          | 35967          | 35965        | 35966         | 35962         | 35981           | 35981             | 35981             |        |
|---------------------------------|---------------|---------------|----------------|----------------|---------------|---------------|---------------|----------------|----------------|--------------|---------------|---------------|-----------------|-------------------|-------------------|--------|
| Site Name                       | OF 59 #59     | OF #60        | OF #61         | OF #61         | OF #61A       | OF62          | Wastech       | Prec +BR       | Prec +BR       | OF 63        | OF 64         | OF 65         | IS Triplicate A | IS (Triplicate B) | IS (Triplicate C) |        |
| Parameter                       | 09/15/2009 09 | 09/14/2009 10 | 09/14/2009 09  | 09/14/2009 09  | 09/09/2009 10 | 09/09/2009 09 | 09/10/2009 10 | 09/10/2009 09  | 09/10/2009 09  | 09/08/2009 1 | 09/08/2009 10 | 09/08/2009 09 | 09/18/2009      | 09/18/2009        | 09/18/2009        |        |
| Unit                            | Sample        | Sample        | Sample - Field | Sample - Field | Sample        | Sample        | Sample        | Sample - Field | Sample - Field | Sample       | Sample        | Sample        | Sample          | Sample            | Sample            |        |
| PCB-87 C-E 86+87+97+108+119+125 | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E          | C-E           | C-E           | C-E             | C-E               | C-E               |        |
| PCB-88 C-E 88+91                | ng/Kg dry wt  | 361           | 423            | 635            | 774           | 393           | 5200          | 1720           | 650            | 719          | 364           | 230           | 128             | 277               | 235               | 259    |
| PCB-89                          | ng/Kg dry wt  | 38.1          | 40.7           | 40.5           | 45.9          | 46            | 403           | 46             | 68.8           | 70.6         | 26.6          | 21.9          | 9.18 J          | 22.4              | 17.2              | 11.3 J |
| PCB-9                           | ng/Kg dry wt  | 7.7 J         | 14.3           | 10 J           | 10.2          | 12.8          | 790           | 13.2           | 9.16 J         | 13.7         | 10.1 J        | 9.28 J        | 4.63 U          | 6.28 J            | 5.42 J            | 9.69 J |
| PCB-90 C-E 90+101+113           | ng/Kg dry wt  | 2130          | 2610           | 3790           | 4420          | 2350          | 20300         | 5650           | 4540           | 4950         | 2520          | 1730          | 889             | 1870              | 1550              | 1550   |
| PCB-91 C-E 88+91                | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E          | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-92                          | ng/Kg dry wt  | 446           | 501            | 741            | 903           | 462           | 3540          | 1720           | 928            | 947          | 514           | 356           | 181             | 404               | 318               | 321    |
| PCB-93 C-E 93+98+100+102        | ng/Kg dry wt  | 136           | 129            | 175            | 193           | 131           | 1190          | 338            | 168            | 189          | 91.7          | 63.1          | 30 J            | 71.2              | 63.0              | 69.6 J |
| PCB-94                          | ng/Kg dry wt  | 22.2          | 17.5           | 22             | 27.1          | 1.41 U        | 277           | 24.3           | 26.8           | 27.3         | 10.7 J        | 9.21 J        | 1.35 U          | 11.2              | 10.3              | 9.69 J |
| PCB-95                          | ng/Kg dry wt  | 1960          | 2390           | 3390           | 4290          | 2380          | 30200         | 8540           | 4230           | 4980         | 2170          | 1510          | 778             | 1750              | 1430              | 1330   |
| PCB-96                          | ng/Kg dry wt  | 24.7          | 25.2           | 28.6           | 33.8          | 27.7          | 459           | 46.5           | 36.7           | 43.3         | 15.3          | 11.9 J        | 4.38 J          | 12.6              | 12.7              | 15.7 J |
| PCB-97 C-E 86+87+97+108+119+125 | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E          | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-98 C-E 93+98+100+102        | ng/Kg dry wt  | C-E           | C-E            | C-E            | C-E           | C-E           | C-E           | C-E            | C-E            | C-E          | C-E           | C-E           | C-E             | C-E               | C-E               | C-E    |
| PCB-99                          | ng/Kg dry wt  | 931           | 1110           | 1740           | 2090          | 933           | 5500          | 2140           | 1920           | 1810         | 958           | 631           | 341             | 776               | 680               | 645    |

C-E Co-Elution



Appendix H: Lower Columbia Slough Sediment Screening Levels

| Lower Columbia Slough Screening Levels - (1/27/12) |   |                     |                                  |            |
|--|---|---------------------|----------------------------------|------------|
| Compound   | Source Control Screening Level (mg/kg-dry weight) |                     | Sediment Screening Level (mg/kg) |            |
|  | Concentration                                     | Basis               | Concentration                    | Basis      |
| <b>Metals</b>                                      |   |                     |                                  |            |
| Aluminum   | NA  |                     | 16867                            | baseline   |
| Antimony   | 4   | background (Oregon) | 4.0                              | background |
| Arsenic  | 7   | background (Oregon) | 7.0                              | background |
| Barium   | NA  |                     | 124                              | baseline   |
| Cadmium  | 1   | background (Oregon) | 1.0                              | background |
| Chromium   | 42  | background          | 44                               | baseline   |
| Copper   | 36  | background          | 38                               | baseline   |
| Cobalt   | NA  |                     | 10                               | baseline   |
| Lead   | 17  | background          | 41                               | baseline   |
| Manganese  | 1100  | toxicity            | 1100                             | toxicity   |
| Mercury (inorganic)                                | 0.07  | background          | 0.1                              | baseline   |
| Selenium   | 2   | background          | 2.0                              | background |
| Silver   | 4.5   | toxicity            | 4.5                              | toxicity   |
| Nickel   | 38  | background          | 38                               | background |
| Thallium   | 0.7   | bioaccum            | 0.7                              | bioaccum   |
| Zinc   | 123   | toxicity            | 244                              | baseline   |
| <b>Polychlorinated Biphenyls</b>                   |   |                     |                                  |            |
| Aroclor 1254                                       | 0.01  | MRL                 | 0.027                            | baseline   |
| Aroclor 1260                                       | 0.01  | MRL                 | 0.019                            | baseline   |
| Aroclor 1248                                       | 0.01  | MRL                 | 0.007                            | baseline   |

Appendix H: Lower Columbia Slough Sediment Screening Levels

| Lower Columbia Slough Screening Levels - (1/27/12) |   |          |                                  |                  |
|--|---|----------|----------------------------------|------------------|
| Compound   | Source Control Screening Level (mg/kg-dry weight) |          | Sediment Screening Level (mg/kg) |                  |
|  | Concentration                                     | Basis    | Concentration                    | Basis            |
| <b>Pesticides</b>                                  |   |          |                                  |                  |
| alpha-BHC  | 0.001   | MRL      | 1.8 x 10 <sup>-5</sup>           | baseline         |
| beta-BHC   | 0.001   | MRL      | 1.9 x 10 <sup>-5</sup>           | baseline         |
| gamma-BHC  | 0.0009  | toxicity | 2.9 x 10 <sup>-5</sup>           | MRL/baseline (a) |
| delta-BHC  | NA  |          | 1.7 x 10 <sup>-5</sup>           | MRL/baseline (a) |
| DDD  | 0.002   | bioaccum | 0.005                            | baseline         |
| DDE  | 0.002   | bioaccum | 0.007                            | baseline         |
| DDT  | 0.002   | bioaccum | 0.0006                           | baseline         |
| Endosulfan   | 0.35  | bioaccum | 0.35                             | bioaccum         |
| Endrin Aldehyde                                    | 0.003   | toxicity | 0.003                            | toxicity         |
| Aldrin   | 0.001   | MRL      | 1.8 x 10 <sup>-4</sup>           | baseline         |
| Chlordane  | 0.001   | MRL      | 0.002                            | baseline         |
| Dieldrin   | 0.001   | MRL      | 7 x 10 <sup>-4</sup>             | baseline         |
| Heptachlor   | 0.001   | MRL      | 5.4 x 10 <sup>-5</sup>           | MRL/baseline (a) |
| <b>Semivolatile Organics</b>                       |   |          |                                  |                  |
| 2-Methylnaphthalene                                | 0.02  | toxicity | 0.020                            | toxicity         |
| Acenaphthene                                       | 0.29  | toxicity | 0.290                            | toxicity         |
| Acenaphthylene                                     | 0.16  | toxicity | 0.160                            | toxicity         |
| Anthracene   | 0.057   | toxicity | 0.620                            | toxicity test    |
| Benzo(a)anthracene                                 | 0.032   | toxicity | 0.695                            | toxicity test    |
| Benzo(a)pyrene                                     | 0.032   | toxicity | 0.718                            | toxicity test    |
| Benzo(b)fluoranthene                               | NA  |          | 1.380                            | toxicity test    |
| Benzo(g,h,i)perylene                               | 0.3   | toxicity | 0.457                            | toxicity test    |
| Benzo(k)fluoranthene                               | 0.03  | toxicity | 0.465                            | toxicity test    |
| Bis(2-ethylhexyl)phthalate                         | 0.75  | toxicity | 4.387                            | toxicity test    |
| Chrysene   | 0.057   | toxicity | 1.015                            | toxicity test    |
| Dibenzo(a,h)anthracene                             | 0.06  | toxicity | 0.140                            | toxicity test    |
| Dibenzofuran                                       | 5.1   | toxicity | 5.100                            | toxicity         |
| Fluoranthene                                       | 0.111   | toxicity | 2.303                            | toxicity test    |
| Fluorene   | 0.077   | toxicity | 0.077                            | toxicity         |
| Indeno(1,2,3-cd)pyrene                             | 0.017   | toxicity | 0.598                            | toxicity test    |
| Naphthalene  | 0.176   | toxicity | 0.176                            | toxicity         |
| Phenanthrene                                       | 0.042   | toxicity | 0.870                            | toxicity test    |
| Phenol   | 0.048   | toxicity | 0.066                            | MRL/baseline (a) |
| 4-methylphenol                                     | NA  |          | 0.066                            | MRL/baseline (a) |
| Pyrene   | 1.9   | CTL      | 2.020                            | toxicity test    |
| Pentachlorophenol                                  | 0.1   | bioaccum | 0.100                            | bioaccum         |

N/A - Not Available

(a) MRL/baseline indicates that the compound was not detected in the 2009 incremental samples at the indicated MRL